

1 **„Everyone does it a bit differently!": Evidence for a positive relationship between**
2 **micro-scale land-use diversity and plant diversity in hay meadows**
3 **Short title: Relationship between plant and land-use diversity**
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17 **Abstract:**

18 High nature-value grasslands including mountain hay meadows are among the most species-rich
19 habitats in Europe. Mountain hay meadows were developed and maintained by traditional, small-scale
20 management systems having high micro-scale land-use diversity (MSLUD), i.e. the parcel-scale diversity
21 of management elements which usually depend on individual decisions and family traditions of local
22 farmers. Detailed studies documenting the effects of micro-scale land-use diversity on vegetation are
23 absent. The main objectives of our study were to analyse the effect of micro-scale land-use diversity and
24 evenness on local plant diversity and cover of the main plant functional types. Field work was carried
25 out in the Gyimes region (Eastern Carpathians, Romania).

26 We conducted semi-structured interviews with the owners and managers of the studied meadow parcels
27 in order to reveal the number of applied management elements (N_m) and applied frequencies of these
28 management elements (e.g. manuring, mowing, seed sowing and weed control) per parcel and to
29 determine the three differently used hay meadow types from interviews. For quantifying MSLUD, the
30 Shannon diversity formula was used, in the case of micro-scale land-use evenness (MSLUE), the original
31 Pielou's evenness formula was applied. To document parcel-scale vegetation features, 4x4-meter
32 quadrats were surveyed in every parcel.

33 We found significant differences in the N_m , MSLUD and MSLUE among the three management types. In
34 models where MSLUD, MSLUE and N_m were built in, we got better model fits and more parsimonious
35 models than in cases where just management type was built into the models. Management elements
36 (manuring, seed sowing) also had a significant effect on vegetation.

37 Our results highlight that micro-scale land-use diversity plays a significant role in the maintenance of
38 plant diversity in traditional, small-scale farming systems. The main drivers behind the high micro-scale
39 land-use diversity may be farmers' personal decisions and family traditions. We argue that for an
40 adequate ecological understanding and conservation of these traditional, small-scale land-use systems,
41 the development of adequate ways of evaluation as well as detailed studies of the effects of several
42 different management elements and land-use diversity on vegetation are needed.
43

44 **Keywords:** mountain hay meadows, traditional management system, East-Central Europe,
45 conservation, Shannon-diversity
46

47 **1. Introduction**

48 High nature-value semi-natural grasslands are considered among the most species-rich
49 habitats in Europe, and are characteristic elements of many cultural landscapes (Fischer and
50 Stöcklin, 1997; Fischer and Wipf, 2002; Mykkestad and Saetersdal, 2004). The main reasons
51 for the diversity of semi-natural grasslands are local, regional (Mykkestad and Saetersdal,
52 2004), and historical factors (Marini et al., 2009), as well as landscape configuration (Janišová
53 et al., 2014), and the traditional, long-term, small-scale, non-intensive land use (Babai and
54 Molnár, 2014; Dorresteyn et al., 2015; Poschlod et al., 2005; Pykälä, 2000).

55 Traditional small-scale farming is characterised by low-input, labour-intensive practices on
56 relatively small parcels. These systems have developed and maintained cultural landscapes
57 with high natural, cultural and aesthetic values all over Europe (Dahlström et al., 2013;

58 Plieninger et al., 2006). Grassland management is an important part of these systems,
59 especially in cultural landscapes where grasslands are semi-natural (of woodland origin), and
60 of high nature value (Babai et al., 2014; Vadász et al., 2016).

61 These traditional land-use systems almost disappeared from Western Europe during the
62 second half of the 20th century (Marini et al., 2009; Meilleur, 1986; Plieninger et al., 2006).
63 Their drastic decrease in Central and Eastern Europe was first caused by communist agricultural
64 policies (Friedmann and McMichael, 1989), followed by political, economic, and social crises
65 after 1990, and finally the diverse effects of the new regulatory systems after the accession to
66 the European Union (e.g. Dorresteijn et al., 2015; Tudor, 2015). Diversity of land use
67 decreased, while its intensity and spatial extent increased, or in many marginal landscapes land
68 use was abandoned (Dengler et al., 2014; MacDonald et al., 2000; Niedrist et al., 2009;
69 Ruprecht et al., 2010; Strijker, 2005). These processes had a negative effect on grassland
70 diversity, causing homogenization of grassland vegetation (Csergő et al., 2013; Mykkestad and
71 Saetersdal, 2003; Spiegelberger et al., 2010).

72 Some of the small-scale land-use systems have avoided the abovementioned drastic changes
73 in marginal, mainly mountainous landscapes of Europe (Babai and Molnár, 2014; von
74 Glasenapp and Thornton, 2011; Tudor, 2015). The main reasons for their survival are economic
75 and natural constraints (cf. Babai et al., 2015). Nature conservation measures also stimulated
76 their survival, or in some cases, their partial revival (Dahlström et al., 2013). Surviving systems
77 give us a chance to study the functioning of traditional small-scale land-use systems which are
78 highly important for the conservation of these species-rich landscapes (Babai et al., 2015;
79 Dahlström et al., 2013; Škodová et al., 2015; Söderström et al., 2001; Sutcliffe and Larkham,
80 2011). By their uniqueness and particular status, such landscapes are threatened in Europe
81 (e.g., Alps – von Glasenapp and Thornton, 2011; North-Eastern Carpathians – Škodová et al.,
82 2015).

83 Several publications highlight the positive impacts of certain management practices
84 (especially the frequency of mowing, Tälle et al., 2018) on local plant diversity, and the possibly
85 important role of management diversity (Marini et al., 2009; Meilleur, 1986; Mykkestad and
86 Sætersdal, 2004; Niedrist et al., 2009; Poschlod et al., 2005; Škodová et al., 2015; Söderström
87 et al., 2001). However, we haven't found detailed studies measuring the effects of micro-scale
88 land-use diversity (MSLUD), evenness (MSLUE) and number of management elements (N_m) on
89 vegetation. Fischer et al. (1996) and Poschlod et al. (2005) emphasize the importance of the
90 small-scale decisions of farmers on the preservation of traditional landscape mosaics and local
91 α and β diversity without providing field evidence. The special effect of land-use diversity on
92 biological diversity has been only studied at a macro- or landscape scale and from a modelling
93 perspective (e.g., Olsson et al., 2000; Yoshida and Tanaka, 2005; Fischer et al., 2008).

94 We studied MSLUD in a traditional cultural landscape in the Eastern Carpathians in Romania
95 (Gyimes) with small-scale spatial mosaicity where species-rich hay meadows are managed by
96 low-intensity traditional management by the local Hungarian Csángó community (Babai et al.,
97 2014). Previous studies (Babai and Molnár, 2014; Babai et al., 2014) show that grassland
98 management in Gyimes is similar to historical or recently abandoned systems of other
99 mountainous landscapes in Europe (e.g. French Alps – Meilleur, 1986; Swiss Alps – Netting,
100 1981; Austrian Alps – von Glasenapp and Thornton, 2011; German Alps – Poschlod et al.,
101 1998). Studying this surviving, still functioning system may help us to better understand one
102 of the most important pillars of the concept of the European cultural landscape (Plieninger, T.
103 and Bieling (eds.), 2012), namely, the extensive traditional land-use system.

104 We have coined the term micro-scale land-use diversity (MSLUD), defined as the parcel-
105 scale diversity of management calculated by Shannon diversity from the different ratios of
106 management elements, and have also coined the term micro-scale land-use evenness (MSLUE),
107 defined as the parcel-scale evenness of management calculated by Pielou's evenness from the
108 ratio of MSLUD to $\log(N_m)$. Types of these elements and the frequency of their use strongly
109 depend on individual decisions and / or family traditions of local farmers (Babai et al., 2014)
110 and are expected to cause plant diversity differences among parcels.

112 The three main objectives of our study are the following:

- 113 • What forms number of management elements (N_m), micro-scale land-use diversity
114 (MLSUD) and micro-scale land-use evenness (MSLUE), and how are they built up?
- 115 • Are there any differences in number of management elements (N_m), micro-scale land-
116 use diversity (MSLUD) and micro-scale land-use evenness (MSLUE) between the main
117 land-use management types?
- 118 • Do land-use management type, number of management elements (N_m), micro-scale
119 land-use diversity (MSLUD) and micro-scale land-use evenness (MSLUE) have a
120 significant impact on local plant diversity and the cover of the main plant functional
121 types?

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123 In this paper we introduce the concept of micro-scale land-use diversity (MSLUD) as a
124 determinant of plant diversity and composition of grasslands.

126 **2 Material and Methods**

127 **2.1 Study area**

128 The study area lies in Valea Rece (Hidegségpataka) in Lunca de Jos (Gyimesközéplak) in the
129 Eastern Carpathians, Romania (coordinates: N: 46.628582, E: 25.958554). Elevation is 800-
130 1550 m above sea level. The montane-boreal climate is modified by continentality, the mean
131 annual temperature ranges from 4 to 6 °C, and the amount of annual precipitation from 700
132 to 1200 mm (Ilyés, 2007; Pálfalvi, 1995). The first settlers arrived in Gyimes in the middle of
133 the 18th century (Babai et al., 2014; Ilyés, 2007). The area of Lunca de Jos is covered by
134 forests (30,2%), hay meadows (30,4%), pastures (36,4%), and arable lands (3,0%) (Sólyom
135 et al., 2011). The human population was 5307 in 2010 ([http#1](#)). The majority of the local
136 population are small-scale farmers, dealing primarily with cattle farming. The average farmland
137 area is 3.8 ha (Knowles, 2010; Sólyom et al., 2011), 0.97 ha is used as hay meadow on average
138 in 3-5 parcels.

139 The area falls within the coniferous forest zone (acidophilous *Picea* forests – R4205) (Doniță
140 et al., 2005). Vegetation of the hay meadows primarily belongs to *Festuca rubra* hay meadows
141 (R3803) and acidofrequent grasslands (R3808), rarely to species-rich *Nardus* grasslands
142 (R3609) (Doniță et al., 2005). Dominant or frequently occurring species are *Arrhenatherum*
143 *elatius*, *Trisetum flavescens*, *Dactylis glomerata*, *Poa pratensis*, *Salvia pratensis*, *Colchicum*
144 *autumnale*, *Ranunculus acris*, *Taraxacum officinale*, *Trifolium pratense*; regionally rare and / or
145 characteristic species are *Carlina acaulis*, *Dianthus compactus*, *Gentiana utriculosa*, *Gladiolus*
146 *imbricatus*, *Trifolium pannonicum*, *Traunsteinera globosa*, *Trollius europaeus*.

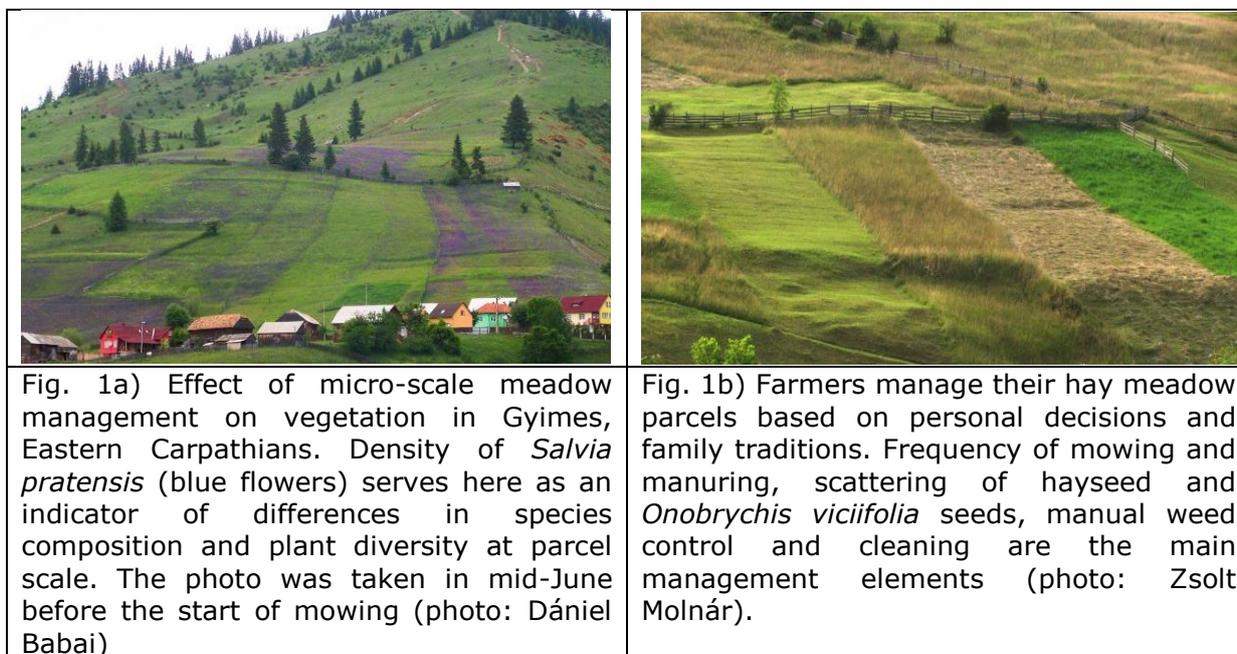
147 Local farmers divide their hay meadows into three types based on their use (Babai and
148 Molnár, 2014): 1) inner (close to the settlement) hay meadows near the farmers' homes on
149 valley floors on very gentle slopes, mown 2(3) times a year (InFI); 2) inner hay meadows on
150 steeper slopes with less intensive use (InSI); and 3) outer hay meadows on slopes farther from
151 settlements, usually at higher altitudes, usually not manured, and mown once a year (Out)
152 (Table 1).

153 The studied hay meadows have brown forest soil with SiO_2 and metallic oxides, but most of
154 them are nutrient rich as a consequence of land management (Table 1). Inner meadows (InFI
155 and InSI) are manured every 1 to 3 years, with an average amount of 8833 kg/ha, with a
156 relatively high standard deviation (SD = 3951 kg/ha), while outer hay meadows (Out) are
157 manured rarely or not at all. The amount of nitrogen used yearly (based on laboratory
158 evaluation of nutrient concentration of local averages of manure samples collected from parcels
159 of land owners) that reaches the meadows ranges from 49.17 to 147.51 kg (mean: 73.76 kg)
160 ha/year, and phosphorus from 7.95 to 23.85 kg (mean: 11.93 kg) ha/year, depending on the
161 frequency of manuring (these values are far below the European average; cf. Ondersteijn et
162 al., 2002) (Kun ined.).

165 Table 1. Nitrogen, phosphorus and potassium concentrations (based on 8 soil samples per parcel, Kun
 166 unpubl.) and slope and altitude values of the three hay meadow management types.

| Characteristics | InFI MEAN±SD | InSI MEAN±SD | Out MEAN±SD | p<0.05 |
|----------------------|-----------------|-----------------|----------------|--------|
| N (mg/kg) | 109.48±32.71a | 76.59±26.37b | 77.02±33.18b | <0.001 |
| P (mg/kg) | 66.42±64.72a | 11.66±3.55b | 9.39±2.43b | <0.001 |
| K (mg/kg) | 156.53±73.92a | 169.89±99.21a | 164.73±55.35a | 0.402 |
| Slope angle (°) | 1.56±1.78a | 24.81±9.23b | 24.14± 3.56 b | <0.001 |
| Altitude (in meters) | 856.88±26.64a | 887.63±38.75b | 959.14±72.11c | <0.001 |

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2.2 Sampling methods

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We conducted semi-structured interviews with the owners and managers (N=16) of the studied parcels (N=23) in June 2013, with a focus on those land-use elements that are important for vegetation (based on Babai and Molnár, 2014; Babai et al., 2015). 81% of the interviewees were farmers as a main occupation; their age varied between 20 and 85 years. The main questions of the interviews referred to the management type of the parcel according to the owner, frequency of mowing and manuring, presence or absence of hayseed and *Onobrychis viciifolia* sowing, manual weeding and clearing in the last 5 years.

Four sampling areas were selected in Valea Rece, each of them containing all three hay meadow management types, with similar exposure within one sampling area. The four sampling areas were chosen as close to each other as possible, and were similar in species composition.

Parcels were localized by the farmers during the interviews (precise location was determined using aerial photos if necessary). Boundaries of parcels were visible in the field, fenced or marked by stakes, anthills, etc. (for more details see Babai et al., 2014). Parcels were managed homogeneously (if homogeneity was not unequivocally clear, the parcel was excluded from the analysis). In the case of outer hay meadows (Out), we sampled their zone which was less elevated and which was nearer to the other two types (Table 1) to avoid the impact of elevation on species composition. Therefore, the potential vegetation is the same in all three hay meadow management types. Sixty-nine 4×4 m² quadrats were surveyed in June 2013. There were three types of parcels (InFI, InSI, Out, see above). Altogether 8 inner meadow parcels on valley floors, 8 parcels of inner meadows on slopes, and 7 parcels of outer meadows were surveyed. Three quadrats were placed randomly in each randomly chosen parcel. Percentage of geometrical cover values of all vascular plant species was estimated in all quadrats.

2.3 Formulating the concepts of micro-scale land use diversity and related data analysis

As indicated above, parcels were classified into the three hay meadow management types by the owner farmers. Based on our interviews with the farmers, the most important management elements on the studied parcels were: 1) frequency of mowing, 2) frequency of manuring, 3) hayseed sowing, 4) sowing with *Onobrychis* seeds, 5) manual weed control, and 6) annual clearing. Based on the interviews, all management elements ($N=6$) applied in the last 5 years were listed for all studied parcels, and with the help of the 6 different management elements we were able to delineate the three meadow types (InFI, InSI, Out) determined by local farmers as well. Management elements were quantified on a ratio scale according to the application frequency of a given management element applied by farmers on a parcel in the last 5 years. Number of management elements (N_m) was quantified using the number of applied management elements for every given parcel (lowest $N_m=2$; highest $N_m=6$). MSLUD was calculated by the Shannon diversity formula, $H = - \sum p_i * \log p_i$ (Peet 1975), p_i being the proportion of i th applied management element on every given parcel. We also introduced the related evenness term: MSLUE expressed as counted by the H/H_{max} formula, where $H_{max} = \log(N_m)$ (Heip 1974, Peet 1975). Micro-scale land-use diversity (MSLUD) has been expressed by Shannon diversity at the parcel scale, and micro-scale land-use evenness (MSLUE) is expressed at the parcel scale. MSLUD was lower when just few management elements (e.g. $N=2$ or $N=3$) were applied on a given parcel with a relatively different ratio (e.g. there was just 1 dominant management element with higher frequency and few with lower frequency), and it was higher when several management elements (e.g. $N_m=5$ or $N_m=6$) were applied with similar frequency on a given parcel. There is an algebraic relationship among MSLUE, MSLUD and N_m . $\log(N_m)$ is the theoretical maximum of MSLUD, while MSLUE is the ratio between MSLUD and $\log(N_m)$. Thus, the more MSLUD approaches its theoretical maximum, the higher MSLUE is.

2.4 Statistical analysis

Species were classified into three main functional types: 'graminoids,' 'forbs,' and 'legumes.' Normality of every variable and their relationships of importance were checked by Shapiro-Wilk normality test. In the case of normally distributed variables we used ANOVA and Tukey HSD tests to test the difference between management types, while in the case of non-normal distribution, Kruskal-Wallis test and Dunn's post hoc test were applied with Bonferroni correction method to counteract the problem of multiple comparison. Linear mixed effect models were used to model the relationship between the three main predictors (N_m , MSLUD, MSLUE) and plant diversity (species number, Shannon diversity) and plant functional types (i.e. graminoids and forbs percentage cover). In our models, management type, N_m , MSLUD, and MSLUE were fixed factors and sampled site was a random factor. We also analysed separately the first three most important management elements determined by farmers (mowing, manuring seed sowing – management elements with strongest hypothetical explanatory power) in a model comparison. Every model comparison started with a model where management type was the only predictor and all following models were compared to this in parsimony and fit. Explanatory power and goodness of fit of the models were calculated with the help of unadjusted R^2 values and Akaike information criterion (AIC). Analyses were made in R 3.5.1 (R Core Team, 2018) software environment.

3. Results

3.1. Micro-scale land-use diversity of the three main hay meadow management types

We found differences in the frequency of management elements between the three hay meadow management types (Table A.1). The outer hay meadows (Out) were not manured (or only occasionally), and were mown only once a year, i.e., they had the lowest land-use intensity. Inner meadows on valley floors (InFI) and on slopes near the village (InSI) had more intensive management; the former were the most often manured and mown meadows. Some management elements were less confined to management types, such as sowing of hayseeds (collected seeds fallen from hay in the barn), sowing *Onobrychis* seeds, and manual control of

weeds (e.g., *Helleborus purpurascens*, *Veratrum album*, *Colchicum autumnale*, young bushes and trees) using a hand scythe. Annual clearing of litter, twigs, ant and mole hills was a constant element of all three types (Table A.1).

Table 2. Number of management elements and micro-scale land-use diversity and evenness values of the three hay meadow management types in Gyimes, Eastern Carpathians.

| Index / variable | InFI MEAN±SD | InSI MEAN±SD | Out MEAN±SD | p<0.05 |
|------------------|-----------------|-----------------|----------------|--------|
| N _m | 4.00±1.11ab | 4.13±0.60a | 3.43±0.90b | 0.034 |
| MSLUD | 1.75±0.35a | 1.93±0.29ab | 2.01±0.22b | 0.036 |
| MSLUE | 0.90±0.03a | 0.93±0.05b | 0.99±0.02c | <0.001 |

There were significant differences in the number of management elements applied and MSLUD between outer hay meadows (Out) and the other two management types (InFI, InSI), while MSLUE was significantly different among all three management types (Table 2, Table A.1).

3.2. Impact of hay meadow management types and micro-scale land-use diversity on local plant diversity and cover of functional types

Plant diversity and cover of graminoid and legume species groups were significantly different among the three hay meadow management types (Table 3.). The quadrat level diversity of inner meadows in valley floors (InFI) was significantly smaller than the diversity of the other two types (InSI and Out). The plant diversity of inner meadows on slopes (InSI) was similar to outer hay meadows (Out) (Table 3). Total cover of legumes and graminoids was significantly different among the inner meadows on valley floors (InFI) and in outer meadows (Out) while inner meadows on slopes (InSI) showed intermediate values. Standard deviations of graminoids, legumes and forbs cover were moderately high or high in all cases, indicating considerable variations within each management type.

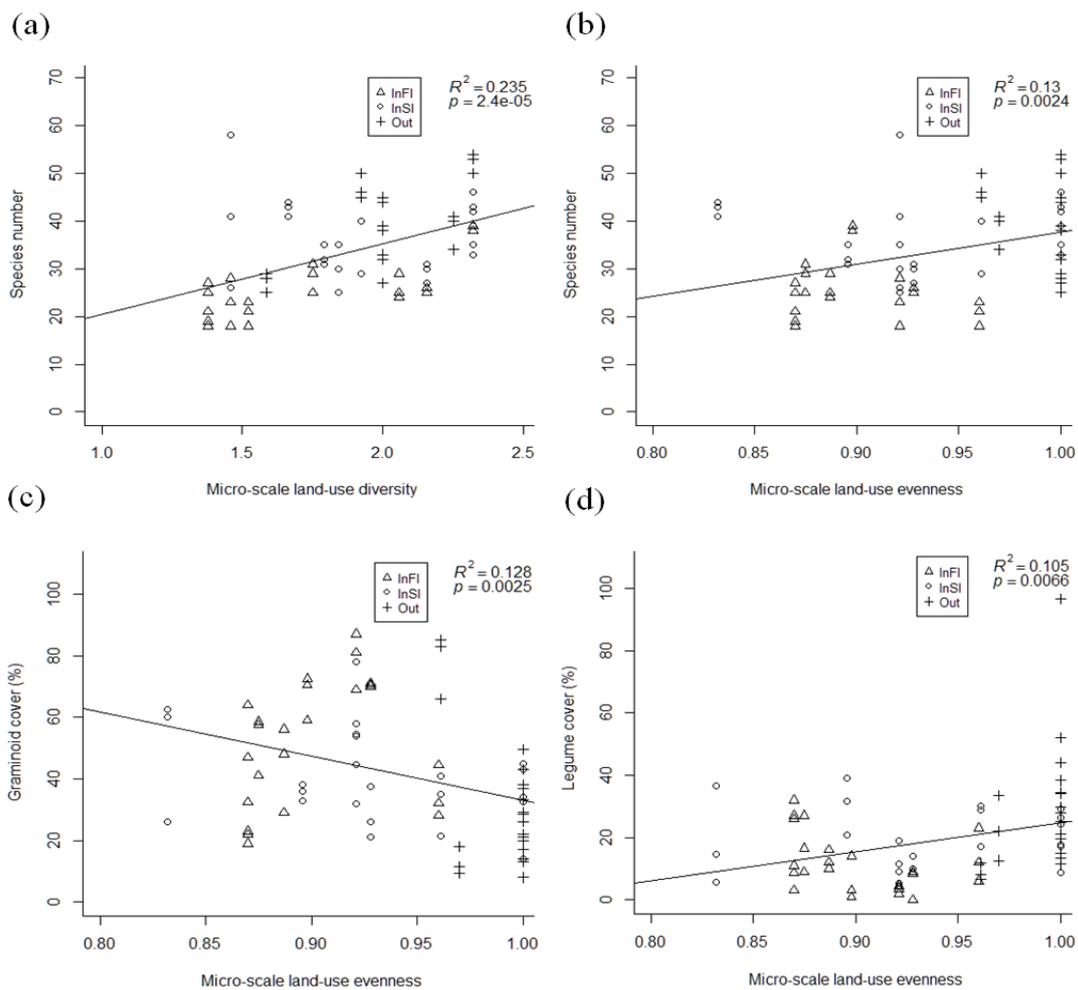
MSLUD and MSLUE explained parcel-scale plant diversity and cover of main plant functional types better than the number of management elements (Table 4). Models which had MSLUD, MSLUE and N_m built in had better parsimony and stronger explanatory power in cases of species number, forbs, graminoids, than models where management type was the only predictor. Individual management practices (manuring, mowing and hayseed sowing) also had a considerable effect on plant diversity and plant functional types cover (e.g. manuring on Shannon diversity, seed sowing on Shannon diversity and on cover of graminoids and forbs – for more details see: Table A.2). MSLUE had a stronger effect on graminoid and legume cover and species number than on Shannon diversity (Table 4). MSLUD had a significant and stronger positive relationship with species number than MSLUE (Fig. 2). MSLUE had a negative effect on graminoid cover and a positive effect on legume cover (Fig. 2).

Table 3. Shannon diversity, species number and cover values of main functional groups in the three hay meadow management types in Gyimes, in the Eastern Carpathians.

| Index / variable | InFI MEAN±SD | InSI MEAN±SD | Out MEAN±SD | p<0.05 |
|-------------------|-----------------|-----------------|----------------|--------|
| Shannon diversity | 2.25±0.29 a | 2.84±0.11 b | 2.63±0.27 b | <0.001 |
| Number of species | 25.58±6.06 a | 36.04±7.66 b | 40.10±8.51 b | <0.001 |
| Forbs cover | 45.08±17.43a | 52.02±13.92a | 55.59±16.92a | 0.100 |
| Graminoids cover | 52.20±19.75a | 40.04±14.71ab | 32.01±21.91b | 0.003 |
| Fabaceae cover | 11.85± 9.00a | 18.34±10.03ab | 27.73±19.45b | <0.001 |

Table 4. Explanatory variables were meadow type (**T**), effect of number of management elements (**N**), micro-scale land-use diversity (**D**), and evenness (**E**). Effects of explanatory variables on plant diversity variables and functional types were measured and compared by R² and Akaike information criterion (AIC) values.

| | Species number | | Shannon diversity | | Graminoids | | Forbs | | Fabaceae | |
|----------------|----------------|----------------|-------------------|----------------|------------|----------------|--------|----------------|----------|----------------|
| | AIC | R ² | AIC | R ² | AIC | R ² | AIC | R ² | AIC | R ² |
| T | 447.79 | 0.38 | 42.29 | 0.29 | 559.39 | 0.15 | 557.91 | 0.06 | 548.62 | 0.18 |
| D | 456.11 | 0.22 | 48.15 | 0.03 | 567.18 | 0.01 | 562.96 | 0.00 | 559.12 | 0.01 |
| E | 455.85 | 0.12 | 42.87 | 0.08 | 560.19 | 0.12 | 558.94 | 0.01 | 551.44 | 0.10 |
| N | 464.30 | 0.03 | 50.06 | 0.04 | 568.41 | 0.04 | 563.21 | 0.06 | 560.52 | 0.03 |
| T+D | 440.90 | 0.45 | 45.93 | 0.28 | 554.32 | 0.15 | 552.76 | 0.08 | 544.82 | 0.18 |
| T+E | 440.42 | 0.38 | 41.66 | 0.29 | 549.80 | 0.16 | 549.13 | 0.06 | 540.71 | 0.18 |
| T+N | 446.73 | 0.38 | 46.49 | 0.32 | 556.16 | 0.16 | 554.23 | 0.10 | 546.94 | 0.18 |
| T+D+E | 431.86 | 0.48 | 45.10 | 0.28 | 544.26 | 0.16 | 543.93 | 0.08 | 536.62 | 0.18 |
| T+D+N | 439.95 | 0.45 | 50.17 | 0.31 | 551.07 | 0.16 | 549.09 | 0.12 | 543.10 | 0.18 |
| T+E+N | 439.37 | 0.37 | 45.75 | 0.32 | 546.42 | 0.17 | 545.54 | 0.10 | 538.94 | 0.18 |
| T+D+E+N | 431.10 | 0.47 | 49.22 | 0.31 | 540.81 | 0.18 | 540.27 | 0.11 | 534.79 | 0.18 |



290 Fig. 2. Linear relationships with best fits. Effect of micro-scale land-use diversity (a) and evenness (b)
 291 on species number and effect of evenness on cover of graminoids (c) and legumes (d) in Gyimes,
 292 Eastern Carpathians.
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295 4. Discussion

296 4.1 Micro-scale land-use diversity and its impact on local vegetation

297 Number of applied management elements (N_m), micro-scale land-use diversity (MSLUD) and
 298 evenness (MSLUE) were different among the three main hay meadow management types in
 299 the study area (Table 2, Table A.1). Additionally, MSLUD and MSLUE had a significant impact
 300 on local (quadrat scale) plant diversity and the cover of graminoids and forbs (Fig. 2, Table 4).

301 From model comparisons it was clear that MSLUD and MSLUE had a considerable effect on the
302 parsimony of models compared with simpler models where management type was the only
303 predictor (Table 4). These results suggest that MSLUD and MSLUE as well as the composition
304 of parcel-scale management may play a significant role in the development and maintenance
305 of plant diversity in traditional, non-intensive, small-scale farming systems. Several
306 management elements of the studied hay meadows were confined to certain types, the major
307 difference being the frequency of mowing, manuring (being less intensive on outer meadows)
308 and hayseed sowing (Table A.1). Clearing management element was present on every parcel
309 while other elements appeared rather randomly (e.g., hayseed sowing). Manuring rates and
310 hayseed sowing had a significant effect on vegetation independently, especially on Shannon
311 diversity, on graminoids and forbs cover (Table A.2). There are other important management
312 elements which contribute to land-use composition and enhance land-use diversity (meadow
313 cleaning, weed control and *Onobrychis* seed sowing) and thereby plant diversity (Fig. 2, Table
314 A.1). Babai and Molnár (2014) and our interviews suggest that the main drivers behind MSLUD
315 in Gyimes are farmers' personal decisions, family traditions, labour and work organisation of
316 the farm, distance, exposure and accessibility of the parcels. An increase in human population
317 in the landscape during the last century resulted in fragmentation of the parcels (mean size <
318 1 ha; Babai et al., 2014). This has led to the development of a small-scale traditional 'precision'
319 management system with careful manuring, hayseed sowing, manual weed control, etc. (Babai
320 et al., 2015).

321 MSLUD had a visible impact on the vegetation in Gyimes (see Fig. 1). Plant diversity, species
322 number and legume cover were lowest in the most intensively managed and most productive
323 meadows in valley floors (InFls), while graminoid cover was the highest with relatively high
324 standard deviations (Table 3). Farmers in Gyimes are aware of the importance of the proportion
325 of graminoids, forbs and legumes as these considerably affect hay quality, grassy hay being
326 preferred by horses, while forb-rich hay by cattle (Babai and Molnár, 2014). Inner meadows
327 are deliberately managed differently because they are highly valued for the high quality second
328 growth cut in late summer (Babai et al., 2015).

329 Diversity and evenness of management seemed to be a more important factor affecting plant
330 diversity and composition than the number of management elements per parcel in itself. MSLUD
331 had a stronger effect on species number than evenness (MSLUE), while MSLUE had a stronger
332 effect on the cover of graminoids and legumes than land-use diversity (MSLUD) (Fig. 2). MSLUD
333 and MSLUE as indices of the parcel-scale composition of management were better predictors
334 in our study than the N_m , where only the parcel-scale number of management elements was
335 taken into consideration. Ecological mechanisms behind these patterns are not yet completely
336 clear.

337 Hayseed sowing was most common on inner meadows on valley floors and on slopes and
338 had a considerable effect on Shannon diversity, forbs and graminoids cover (Table A.2). This
339 practice may significantly contribute to the propagule dispersion in this landscape and may
340 have a significant positive impact on species number (Babai et al., 2015). Hayseed sowing
341 (from local seed sources) is not a widespread management element in European hay meadows
342 today (Babai and Molnár, 2014; Ivaşcu et al., 2016) but might have been a common practice
343 in the past, until the 19th century (Poschlod and Wallis de Vries, 2002; Poschlod and Biewer,
344 2005; and unpubl. data of the authors). *Onobrychis viciifolia* seed sowing was also a common
345 practice in our Eastern Carpathian study area. *Onobrychis* improves forage quality, helps
346 equalize the forage value of the parcels (Babai et al., 2015), and as it is not applied to all
347 parcels, it adds to land-use diversity. Exact timing of mowing can also be a key factor affecting
348 local plant diversity. Several days' or 1-2 weeks' difference in mowing time among years
349 certainly affects the composition of seeds fallen back in that year to that parcel. However,
350 correct documentation and quantification of this management practice was not possible (but
351 see an exceptional case study from England, [http#2](#)). Calculating long-term yearly differences
352 in average (!) mowing times would be a first step to document this diversity.

353 Although plant diversity significantly differed between hay meadow management types, this
354 diversity was relatively high in all three types (Table 3). Besides the relatively low intensity of

355 traditional farming (Babai et al., 2015; cf. Maurer et al., 2006; Niedrist et al., 2009), some
356 management elements (e.g. hayseed sowing), land-use diversity and evenness might have
357 contributed to this unexpected homogenous pattern. We emphasize that even the most
358 intensively used parcels in Gyimes had high species diversity compared to most European hay
359 meadows (see e.g. Niedrist et al., 2009; Plantureux et al., 2005).

361 **4.2 Nature conservation, agricultural regulations and micro-scale land-use diversity**

362 According to our interviews with local farmers and previous studies (Babai and Molnár, 2014;
363 Babai et al., 2014; 2015), the main objective of the well-developed traditional small-scale hay
364 meadow management system in Gyimes is to increase the reliability of a natural resource
365 provision, i.e. to ensure the necessary hay fodder for the winter and decrease inter-annual
366 fluctuations in its quantity and quality. The relatively high species diversity of these meadows
367 is actually only a 'by-product' of their activities.

368 Farmers in Gyimes use their hay meadows non-intensively due to natural, lifestyle and
369 regulatory constraints. Diversity of use is increased by personal decisions emerging from family
370 traditions. The optimal ratio of management elements and enhancement of MSLUD and MSLUE
371 can help farmers sustain the level of quality of hay and can help increase species diversity.
372 Land abandonment, a major cause of meadow degradation Europe-wide (Galváneek and Lepš,
373 2008; MacDonald et al., 2000; Plieninger et al., 2013; Poschlod et al., 2005; Ruprecht et al.,
374 2010) results in decreasing plant diversity in this region also (Csergő et al., 2013). However,
375 due to the economically marginal situation of the local community and the availability of the
376 European Union agricultural subsidies promoting continued land use, land abandonment is less
377 prominent in this landscape than in the adjacent regions (Demeter and Kelemen, 2012; Sólyom
378 et al., 2011) (but it exists in this landscape as well). The main reason for this is that subsidies
379 provide one of the main sources of cash for local livelihoods in this region (Babai et al., 2015;
380 Sólyom et al., 2011).

381 Agricultural regulations and subsidies, however, have negative effects as well. Mowing on
382 inner meadows in the valley floors and slopes has become more uniform in recent years and
383 has shifted to a later date due to regulations. These changes are economically disadvantageous
384 to family farms, since they can only harvest the hay late, in a sub-optimal state (Babai et al.,
385 2015). Hence, regulations can decrease MSLUD and thus can cause a decrease in plant diversity
386 in the future. For this reason, it would be very important to monitor how MSLUD would change
387 as a result of planned regulation and through this, how it would affect plant diversity.

388 In the Gyimes region this effect (i.e., the more uniform time of mowing) has become even
389 more widespread with the recent introduction of small mowing machines. On the other hand,
390 mowing machines slowed down the pace of abandonment since they make harvesting more
391 efficient, thus, farmers continue their management (about 90% of mountain hay meadows are
392 still managed; Demeter and Kelemen, 2012).

393 Similarly to other European examples (Romania – Dahlström et al., 2013; Switzerland –
394 Fischer and Wipf, 2002; von Glasenapp and Thornton, 2011; France – Meilleur, 1986), local
395 people of Gyimes have adapted their complex land-use system to the potentials and constraints
396 of their natural environment, building on their deep traditional ecological knowledge (Babai et
397 al., 2014). European Union and government regulations should take these local traditions into
398 consideration when developing regulatory systems (Babai et al., 2015) to maintain the special,
399 high MSLUD in such traditional cultural landscapes (Molnár and Berkes, 2018). Furthermore,
400 we argue that for an adequate ecological understanding and conservation of these diverse
401 small-scale land-use systems, detailed studies of the combined effects of all the different
402 management elements on vegetation (including their variability and diversity) are needed (cf.
403 Vadász et al., 2016). Developing better ways of quantifying MSLUD (e.g., using diversity indices
404 and determining their sensitivity to special situations) is a major task for future research.

405 Our closing quote from a local farmer indicates that traditional farmers in Gyimes are aware
406 of the high micro-scale land-use diversity of their management system, and that their deep
407 traditional understanding of vegetation dynamics is still alive: „If there would be no fence, the

408 *parcel boundaries would still be visible, since everyone does it (the management of the parcels)*
409 *a bit differently!" Let's help them continue!*

411 **7. Acknowledgements**

412 We are grateful for our helpful local interviewees: Csaba Ambrus, Virág Blága, Dezső Boczony, Ibolya
413 Boczony, Ervin Bodor, Dénes Csilip, Károly Ferencz, Lenke Mária Ferencz, László Gábor, András Lajos,
414 Anna Molnár, Károly Molnár, Mária Molnár, Bettina Prezsmer, Károly Prezsmer, Mihály Prezsmer, István
415 Tankó, Antal Tímár, Dezső Tímár, Gyula Tímár, Gyula János Tímár, Piroska Tímár, Zoltán Molnár, and
416 János Tímár.

417 We also thank Ábel Molnár, Attila Barczy, Dénes Saláta, Csaba Centeri, Klára Virágh, András Kelemen,
418 Orsolya Valkó, Béla Harman, Péter Ragályi, Michal Hejcman, Jan Lepš, Thomas Fricke, and Samantha
419 Charman for their help in field work and analyses.

420
421 This work was supported by the Szent István University Kutató Kari Pályázat (MKK-791-10-34/2014)
422 and by the Szent István University, KTDI (Robert Kun); by GINOP-2.3.2-15-2016-00019 project (Zsolt
423 Molnár, Sándor Bartha) furthermore by the MTA Postdoctoral Scholarship (PPD008/2017) (Daniel Babai).

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Appendix

Table 1. Management elements of the three hay meadow types at parcel scale (N=23). Black cells: presence; white cells: absence of a given management element in the last 5 years. In the case of manuring, four levels, in the case of mowing, two levels were used to distinguish different intensity regimes.

| Management elements | InEI | InEI | InEI | InEI | InEI | InEI | InEI | InEI | InSI | Out | Out | Out | Out | Out | Out |
|---|-------------------------------------|------|------|------|------|------|------|------|------|------|------|------|------|------|------|-----|-----|-----|-----|-----|-----|
| | 1a) Intensive manuring (every year) | | | | | | | | | | | | | | | | | | | | |
| 1b) Moderate manuring (every second year) | | | | | | | | | | | | | | | | | | | | | |
| 1c) Slight manuring (every third year / sparsely) | | | | | | | | | | | | | | | | | | | | | |
| 1d) Abandoned manuring, or sparsely on part of the parcel | | | | | | | | | | | | | | | | | | | | | |
| 2a) Frequent mowing (3 or more times per year) | | | | | | | | | | | | | | | | | | | | | |
| 2b) Non-frequent mowing (1, sometimes 2 per year) | | | | | | | | | | | | | | | | | | | | | |
| 3) Hayseed sowing | | | | | | | | | | | | | | | | | | | | | |
| 4) Sowing Onobrychis seed | | | | | | | | | | | | | | | | | | | | | |
| 5) Manual weed control | | | | | | | | | | | | | | | | | | | | | |
| 6) Clearing of litter, twigs etc. | | | | | | | | | | | | | | | | | | | | | |

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Table 2. Explanatory variables were meadow type (**T**), manuring intensity (**man**), mowing intensity (frequency) (**mow**), and hayseed sowing (**seed**). Effects of explanatory variables on plant diversity variables and functional types were measured and compared by R² and Akaike information criterion (AIC) values.

| | Species number | | Shannon diversity | | Graminoids | | Forbs | | Fabaceae | |
|-----------------------|----------------|----------------|-------------------|----------------|------------|----------------|--------|----------------|----------|----------------|
| | AIC | R ² | AIC | R ² | AIC | R ² | AIC | R ² | AIC | R ² |
| T | 447.79 | 0.38 | 42.29 | 0.29 | 559.39 | 0.15 | 557.91 | 0.06 | 548.62 | 0.18 |
| man | 459.60 | 0.20 | 42.88 | 0.23 | 565.68 | 0.14 | 562.56 | 0.09 | 558.76 | 0.08 |
| mow | 458.05 | 0.20 | 44.45 | 0.16 | 565.93 | 0.08 | 562.20 | 0.05 | 556.46 | 0.09 |
| seed | 463.33 | 0.02 | 47.00 | 0.09 | 555.29 | 0.37 | 556.63 | 0.20 | 553.56 | 0.15 |
| T+man | 446.50 | 0.37 | 42.60 | 0.38 | 555.63 | 0.16 | 554.16 | 0.09 | 546.20 | 0.18 |
| T+mow | 444.14 | 0.39 | 45.83 | 0.29 | 554.26 | 0.14 | 553.32 | 0.06 | 544.65 | 0.18 |
| T+seed | 445.66 | 0.37 | 43.36 | 0.35 | 544.30 | 0.43 | 548.40 | 0.22 | 541.43 | 0.25 |
| T+man+mow | 442.77 | 0.38 | 46.17 | 0.37 | 550.27 | 0.16 | 549.56 | 0.09 | 542.26 | 0.18 |
| T+man+seed | 444.32 | 0.36 | 43.27 | 0.43 | 540.89 | 0.43 | 544.78 | 0.24 | 538.99 | 0.26 |
| T+mow+seed | 441.94 | 0.38 | 46.77 | 0.35 | 539.71 | 0.42 | 543.62 | 0.22 | 537.79 | 0.25 |
| T+man+mow+seed | 440.51 | 0.37 | 47.13 | 0.43 | 536.28 | 0.42 | 540.23 | 0.24 | 535.28 | 0.25 |

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