Supplementary Material 1.

Distribution of fine-scale species richness in 5 x 5cm micro-quadrats (Figure S1).



5 6

Figure S1 Distribution of fine-scale species richness in 5 x 5cm micro-quadrats (all species were considered and all data merged). Data represent the 12 transects (each 5m long) that was monitored over 5 years. The median species richness was 4 species (Q1=3 Q3=5 species).
Species richness categories refers to the number of species found in a micro-quadrat.

10

11 Supplementary Material 2.

12 Diversity of species combinations and spatial scaling

13 Plant species form various species combinations in nature. For assessing within-stand spatial 14 patterns of species combinations we used a diversity measure from a family of information theory models developed by Juhász-Nagy (Juhász-Nagy and Podani 1983, Juhász-Nagy 15 1993). We chose this model because it appeared to be especially sensitive detecting fine-scale 16 17 changes in coexisting relationships within communities and could be used successfuly to 18 describe vegetation changes during succession or degradation (Juhász-Nagy and Podani, 19 1983; Margóczi, 1993; Gosz, 2000; Bartha et al., 1995, 1998, 2004, 2011, 2020; Virágh 2008; 20 Bakacsy, 2019). Using this model we expect to detect some early signals of vegetation 21 changes due to changing the management regimes.

22 Spatial scaling is an inherent part of the methodology (Juhász-Nagy, 1993; Podani et al., 1993; Bartha et al., 1998) and the biological interpretation of models linked directly to the 23 assembly dynamics (Bartha et al., 1995, 1998; Bartha et al., 2020). Traditional diversity 24 25 indices often assessed at a single scale or at few arbitrarily chosen scales and they are often estimated from small samples. In contrast, Juhász-Nagy's models are assessed at a series of 26 27 gradually increasing sampling units sizes (spatial scaling process) (Podani, 1987; Podani et 28 al., 1993). The transect sampling design we used in this study was specifically developed and optimized for estimating these models (Bartha et al., 2004). Compositional Diversity (CD) is 29 30 the diversity of realized (observed) species combinations at a given scale; which is calculated 31 as the Shannon entropy of the frequency distribution of species combinations observed within 32 the sampling units:

$$CD = -\sum_{k=1}^{\omega} p_k \log_2 p_k,$$

33 34

35 where $\omega = 2^{S}$ is the number of possible species combinations, S is the number of species and 36 p_k is the relative frequency of the **k**th species combinations in the sample. To estimate CD 37 large sample sizes are required (100 units or more) and it is calculated at a series of increasing 38 sampling unit sizes (gradually merging 2, 3, 4, etc. adjacent micro-quadrats along transect 39 (Figure S2., S3). Spatial scaling used to be performed by computerized resampling of the 40 base-line transects (Podani, 1987; Podani et al., 1993) with gradually increasing sampling unit 41 sizes.

42 Illustration of how CD (the Shannon diversity of species combinations) differs from the

43 traditional Shannon diversity calculated from species abundances is shown in Figure S4.44



CD = -0.1*log₂(0.1) -0.15*log₂(0.15) -0.15*log₂(0.15) -0.35*log₂(0.35) -0.25*log₂(0.25)=2.183 bits

Figure S2. Illustration of the computerized sampling from the base-line transect data and the
 calculation of Compositional Diversity using artificial data Example for calculating
 Compositional Diversity with grain size=1

50 1, The baseline transect (20 units long with 3 species) resampled with computer (with grain 51 size =1) and a binary coenological table is created.

52 2, Species combinations calculated from the binary coenological table.

3, Number of realized species combinations (NRC) are the number of combinations with nonzero frequency (from 3 species the potential maximum number of combinations would be 8,
however, only 5 had non-zero frequency in our example (NRC=5).

56 4, Compositional Diversity (CD), i.e. the diversity of species combinations, which is 57 calculated based on the relative frequency of species combinations.



(10 x 5 cm) (two adjacent small quadrats merged):

 $CD = -0.2*log_2(0.2) - 0.2*log_2(0.2) - 0.1*log_2(0.1) - 0.3*log_2(0.3) - 0.2*log_2(0.2) = 2.246 \text{ bits}$

60

61 **Figure S3.** Example for calculating Compositional Diversity with grain size=2

1, The baseline transect (20 units long with 3 species) resampled with computer (with grain
 size =2) and a binary coenological table is created.

64 2, Species combinations calculated from the binary coenological table.

65 3, Number of realized species combinations (NRC) are the number of combinations with non-

c66 zero frequency (from 3 species the potential maximum number of combinations would be 8,

67 however, only 5 had non-zero frequency in our example (NRC=5).

- 68 4, Compositional Diversity (CD), i.e. the diversity of species combinations is calculated based
- 69 on the relative frequency of species combinations.

70

59

Spatial scaling with grain size= (



71 Figure S4. Illustration of how CD (the Shannon diversity of species combinations) differs 73 from the traditional Shannon diversity calculated from species abundances. In the example we 74 show three different communities with the same abundance of species but with different 75 patterns of species co-occurrences (different combinations of species). The traditional 76 abundance diversity estimates do not differ between these communities. In contrast, 77 Compositional Diversity (CD) is able to detect the fine changes in coexistence relationships.

79 Illustration for the importance of estimating Compositional diversity (CD) at many different 80 grain sizes. At very small grain sizes many of the sampling units are empty or contain only one species (consequently we have relatively few species combinations). At very large grain 81 82 size it can appear that all sampling units have all species (i.e. we have only one species 83 combinations). The largest number of realized species combinations tends to appear in 84 intermediate grain sizes. Consequently, CD shows a unimodal maximum curve if it is plotted 85 against the grain size. (Figure S5, S6)

The maximum Compositional Diversity is usually different in different communities and it 86

87 can change also due to stress, disturbance or succession (Juhász-Nagy and Podani, 1983;

88 Margóczi, 1993; Bartha et al., 2004, 2011; Virágh et al., 2008; Bartha et al., 2020). Careful

spatial scaling is necessary to find precisely the resolution where the function maximum 89

90 appears. In this paper 23 different grain sizes were tested. (Figure S5, S6).

91



Figure S5. Spatial scaling of Compositional diversity with real data (one transects in 2015 at 94 the grazed site). The arrow shows the spatial scale where the function maximum was detected. 95



Figure S6. Illustration of improper scaling. Biased results for curve and for maximum scale 98 because few grain sizes were tested (real data, the same transect in 2015 at the grazed site). 99 Arrow shows the spatial scale where the function maximum was detected (10cm was the 100 proper scale but 5cm was detected due to improper spatial scaling).

101

102 Illustration of choosing the right number of species in the analyses.

103

104 The number of potential species combinations of s species is 2^{s} , i.e. the number of potential species combinations increases exponentially when we increase the number of species in the 105 106 analyses (Figure S7). Selecting too many species in the analyses might results in biased 107 estimates because it would require very large sample sizes that usually cannot be collected in 108 practice. Therefore, we usually focus on a reasonable subset of species (usually the most 109 important dominant species forming the vegetation matrix).

110

111 The number of species combinations found in an analysis is constrained by the maximum 112 number of sampling units. The sample size was 100 in our case that allow 6-7 species (if s=6 113 2^{s} =64 if s=7 2^{s} =128) to be considered (Figure S7). However, during analyses we do not prefer 114 to fix the number of species considered because the number of species playing an important 115 role tends to change during vegetation dynamics. Instead, we fix the maximum number of species that can be analysed. Setting an abundance threshold (larger or equal to 20 presences 116 117 per transect) was an optimal choice in our study resulting in 6-8 species and in 6-96 species 118 combinations. We assume here that species appearing in less than 20% of sampling units 119 might not be so important functionally.





124 However, compositional diversity curves appear in truncated forms if it is calculated from too 125 many species (Figure S8). Maximum values and the spatial scale of maxima cannot be detected precisely from these truncated curves. For 15, 35 or 45 species the max. CD would 126 127 be larger than 7 bits. However, due to the limited sample size (m=100) we reached a threshold where all microquadrats had different species combinations and all combinations were equally 128 frequent, this is 6.643 bits in our case. Larger CD values cannot be detected. 129

130

123



131 132 Figure S8. C compositional diversity curves appear in truncated forms if it is calculated from 133 too many species (maximum values and the spatial scale of maxima cannot be detected 134 precisely from these truncated curves. For 15, 35 or 45 species the max. CD would be larger 135 than 7 bits.)

In this study we used an abundance threshold (minimum 20 presence in a 100 units (5m long)
transect. This threshold resulted in 6-8 species per transects and represented the most
important dominant (matrix forming) species of the grassland.

140

141 Characteristic scales detected in this study

142

After careful scaling and careful selection of proper abundance threshold, the detected maximum scales did not differ between the grazed (transects G1-G6) and mowed (transects M1-M6) vegetation patches (Figure S9). All maximum data appeared at 5 x 5 cm (or rarely at 5 x 10 cm) (Figure S9). In contrast to the heterogeneity in species composition (spatial heterogeneity) and the compositional variability due to fluctuating weather conditions (temporal heterogeneity), the estimated scales of maximum Compositional Diversity appeared at very fine scales and remained stable in our study.







155 References

- 156
- 157 Bakacsy, L. 2019. Invasion impact is conditioned by initial vegetation states. Community Ecology 20: 158 11-19.
- 159 Bartha, S., Collins, S. L., Glenn, S. M., Kertész, M., 1995. Fine-scale spatial organization of tallgrass 160 prairie vegetation along а topographic gradient. Folia Geobot., 30, 169–184. 161 https://doi.org/10.1007/BF02812096
- 162 Bartha, S., Campatella, G., Canullo, R., Bódis, J., Mucina, L., 2004. On the importance of fine-scale
- 163 spatial complexity in vegetation restoration. International Journal of Ecology and Environmental 164
- Sciences, 30, 101–116.

- Bartha, S., Czárán, T., Podani, J., 1998. Exploring plant community dynamics in abstract
 coenostate spaces. Abstr. Bot., 22, 49–66.
- Bartha, S., Campetella, G., Ruprecht E., Kun A., Házi J., Horváth A., Virágh K., Molnár Zs. 2008.
 Will inter-annual variability in sand grassland communities increase with climate change?
 Community Ecology 9 Suppl.: 13-21.
- Bartha, S., Campetella, G., Kertész, M., Hahn, I., Kröel-Dulay, Gy., Rédei, T., Kun, A., Virágh, K.,
 Fekete, G., Kovács-Láng, E., 2011. Beta diversity and community differentiation in dry perennial
 sand grasslands. Annali di Botanica, 1, 9–18. DOI: <u>https://doi.org/10.4462/annbotrm-9118</u>
- Bartha S., Campetella, G., Chelli, S., Canullo, R. 2020. Patterns of herb layer diversity
 in coppiced Italian beech forests. Diversity 12(3) 101. https://doi.org/10.3390/d12030101
- Collins, S.L., Xia, Y., 2015. Long-term dynamics and hotspots of change in a desert grassland plant
 community. Am. Nat., 185, E30–E43. doi: 10.1086/679315.
- Gosz, J., Peters, D., Kertész, M., Kovács-Láng, E., Kröel-Dulay, Gy., Bartha, S. 2000. Organization of
 grasslands along ecological gradients: US-Hungarian LTER Grassland cooperation. pp. 67-76. In:
 Lajtha,, K. & Vanderbilt, K. (eds.), Cooperation in Long term Ecological Research in Central and
 Eastern Europe: Proceedings of the ILTER Regional Workshop, 22-25 June, 1999, Budapest,
 Hungary, Oregon state University, Corvallis, OR
- 182 Juhász-Nagy, P., 1993. Notes on compositional diversity. Hydrobiologia, 249, 173–182.
 183 doi.org/10.1007/BF00008852
- Juhász-Nagy, P. and Podani, J. 1983. Information theory methods for the study of spatial processes
 and succession. Vegetatio 51: 129-140.
- 186 Margóczi, K. 1993. Comparative analysis of successional stages of sandy vegetation a case study.
 187 Tiscia 27: 3-8.
- 188 Podani, J. 1987. Computerized sampling in vegetation studies. Coenoses 2: 9-18.
- Podani, J., Czárán, T., Bartha, S., 1993. Pattern, area and diversity: the importance of spatial scale in
 species assemblages. Abstracta Botanica, 17, 289–302.
- Tóthmérész, B., Erdei, Zs., 1992. The effect of dominance in information theory characteristics of
 plant communities. Abstracta Botanica, 16, 43–47.
- 193 Virágh, K., Horváth, A., Bartha, S., Somodi, I., 2008. A multiscale methodological approach for
 194 monitoring the effectiveness of grassland management. Community Ecol., 9, 237–246.
 195 doi.org/10.1556/ComEc.9.2008.2.13