## **ORIGINAL PAPER**



# Effects of tree size and park maintenance on soil seed bank of *Gleditsia* triacanthos, an exotic tree in urban green areas

Péter Csontos<sup>1</sup> · Tibor Kalapos<sup>2</sup> · Tito Faradhimu<sup>3</sup> · Annamária Laborczi<sup>1</sup> · Tamás Hardi<sup>4</sup> · Júlia Tamás<sup>5</sup>

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#### Abstract

The exotic honey locust (*Gleditsia triacanthos*) is often planted as ornamental tree in urban parks. In Hungary, it occasionally escapes cultivation, in other countries it has already become invasive, and thus, further spread cannot be ruled out. The production of copious long-lived seeds may contribute to its invasiveness. We investigated the soil seed bank of honey locust growing in urban parks of Budapest (Hungary). Soil samples of 1200 cm³ were taken under the canopy of 48 solitary female trees, seeds were extracted by washing, and their viability was assessed by germination test following scarification. For each tree, trunk circumference and level of park maintenance were recorded. Seed bank density varied between 0 and 500 seeds/m² and in 17 out of 48 cases, it exceeded 100 seeds/m². Larger trees tended to form better-developed soil seed bank than smaller ones, yet the level of park maintenance has a much stronger effect: trees in neglected parks produced five-times higher density on average than those in perfectly managed parks (126.4 and 24.5 seeds/m², respectively). For a better understanding of the species' invasion risk, detailed seed ecological studies are needed and to prevent its further spreading, a more careful treatment of its litter is recommended.

Keywords Alien tree · Germination test · Honey locust · Mechanical scarification · Soil seed bank · Urban parks

## Introduction

Urban green areas in modern cities provide a number of ecosystem services for inhabitants such as noise reduction, water and air purification and microclimate stabilization (Bolund and Hunhammar 1999; Chiesura 2004; Bowler et al. 2010). Additionally, these areas are essential in providing

- Institute for Soil Science and Agricultural Chemistry, Centre for Agricultural Research, Herman Ottó út 15, Budapest 1022, Hungary
- Department of Plant Systematic, Ecology and Theoretical Biology, Eötvös Loránd University, Pázmány P. stny. 1/C, Budapest 1117, Hungary
- Centre for Environmental Studies, Faculty of Natural Sciences, Eötvös Loránd University, Pázmány P. stny. 1/C, Budapest 1117, Hungary
- Institute for Regional Studies, Centre for Economic and Regional Studies, P.O. Box 420, Győr 9002, Hungary
- Department of Botany, Hungarian Natural History Museum, Könyves Kálmán krt. 40, Budapest 1087, Hungary

recreational and aesthetic services (Brander and Koetse 2011; Buchel and Frantzeskaki 2015).

Exotic ornamental species are widely planted in urban green areas and contribute to many of the park's beneficial effects in the same way as native species do (e.g., heat wave mitigation, water retention, wind protection). In certain aspects, like in increasing the beauty of parks, these alien plants may even exceed native species (Muratet et al. 2015). However, there are many non-native plant species which originally has been introduced intentionally for ornamental purposes, later escaped from cultivation, established their self-sustaining stands and became invasive, thus threatening native biodiversity in resident (semi)natural ecosystems in the countryside (Essl 2007; Čeplová et al. 2017). Undoubtedly, invasive plants are among the prominent conservation threats and have affected biodiversity seriously throughout the world already (Richardson and Pyšek 2006; Chmura 2013; Early et al. 2016).

Honey locust (*Gleditsia triacanthos*) is a deciduous dioecious tree native to central North America and has been planted throughout the world—including Europe—since 1700 (Benčať 1982). In rural areas, this species is often used as hedge plant because of its sharp thorns and resistance



to pruning. Due to its tall growth, delicate foliage turning golden in autumn, and conspicuous large seedpods staying long on the tree, honey locust is widely planted in urban green areas and parks for ornamental purposes.

From some European countries, including Hungary, honey locust was reported as sporadically occurring alien tree (Balogh et al. 2004; Pyšek et al. 2012; Glišić et al. 2014; Glogov et al. 2019). However, its recent sporadic character does not rule out that it may be seriously invasive in the future. Numerous alien species survive in its new territory in low abundance for years after its introduction (called "lag phase") before suddenly becoming invasive (Hobbs and Humphries 1995). This has already happened in some regions, especially in Argentina where honey locust becomes one of the most harmful invasive plants with great potential to continue further expanding its range (Ferreras et al. 2015; Fernandez et al. 2017). Local adaptation of plant traits, particularly those associated with seeds and seedlings, could play a key role in the process (Kudoh et al. 2007; Tognetti et al. 2019).

Honey locust produces abundant seed crop every year, and the seeds are hard-coated having a thick seed coat impermeable to water that makes the trees capable of establishing high-density soil seed bank under their canopies. The seeds can also be dispersed to large distances by birds and mammals which eat the pod for its high carbohydrate content (Blair 1990). The viability of seeds lasts long; therefore, considerable amount of seeds could accumulate in the soil especially under old trees.

The soil seed bank of alien species can be the source of unintentional spread and colonization in neighboring areas when topsoil replacement is applied during park renovation works. Regular litter removal, that includes seed pods as well, might prevent or reduce soil seed bank formation in parks. However, transport of litter to sites where it is treated (composted, incinerated or deposited to landfill)—if not carried out properly—can initiate new colonization. Therefore, human effects such as shaping the species composition and age structure of urban green areas as well as the intensity of park maintenance are vital in preventing the unwanted impact of alien species, including honey locust.

In this study, we aimed (1) to investigate soil seed bank density under the canopy of honey locust trees growing in urban green areas of Budapest; (2) to determine the effect of tree size on the soil seed bank; and (3) to clarify the influence of park maintenance intensity on the development of soil seed bank under the canopy of honey locust trees. It was hypothesized that larger trees tend to have better-developed soil seed bank under the canopy compared to smaller trees, and soil seed bank will be less developed in well-managed parks than in neglected ones due to the regular removal of fallen seedpods during park maintenance.



To investigate the soil seed bank under ornamental honey locust trees, a total of 48 solitary female individuals growing in parks or other urban green areas (collectively named parks hereafter) were selected across Budapest (Fig. 1). Double-trunk trees and those having damaged crown were excluded. Exact GPS coordinates of the selected trees are listed in "Appendix 1".

The honey locust trees sampled in parks were divided into three size categories as follows: (1) small trees (trunk circumference is lower than 0.60 m); (2) medium-sized trees (trunk circumference is between 0.60 and 1.20 m); and (3) big trees (trunk circumference is larger than 1.20 m). Tree trunk circumference was measured at breast height with 0.01 m accuracy by using a 1.5 m tape measure.

According to park maintenance intensity, three different levels were distinguished: (A) perfectly maintained parkthe condition of urban parks which are very well treated, for instance, fallen leaf litter, branches and seed pods regularly removed, lawn (if present) cut short, etc.; (B) moderately maintained parks—the urban parks receive the above treatments irregularly thus some litter, fruits and hay may accumulate for shorter periods; (C) neglected parks—abandoned parks or other urban green areas without maintenance resulting in the accumulation of litter and seeds from woody and herbaceous plants. Each park was classified into one of these three categories based on their appearance at the time of the visit. We are aware that the above classification of parks includes some uncertainties because maintenance intensity of certain parks might have changed through time. The combination of tree size and maintenance intensity resulted in nine categories altogether. The number of sampled trees in each category is listed in Table 1.

For soil seed bank studies, 5 cm deep soil cores were cut by using a metal corer of 80 cm<sup>2</sup> surface area, in three replicates, under the canopy of each tree specimen. The three sampling points were located approximately 1.5 m from the trunk base in East, South and West directions, respectively. The three soil cores were merged and kept in a labeled plastic bag thus forming a soil sample of 1200 cm<sup>3</sup> which satisfies the minimum sampling volume requirement according to Roberts (1981). Soil sampling was completed from July 2018 to October 2018. Following sampling, holes on the surface were filled with soil from nearby to restore plain ground. Soil samples were transported to laboratory in the Eötvös Loránd University, Budapest, and kept in dark and cool place until processing.

Between October 2018 and November 2018, soil samples were washed and sieved using a mesh size of 3.15 mm, and then all material retained on this mesh was inspected for seeds. Seeds of honey locusts were extracted by hand



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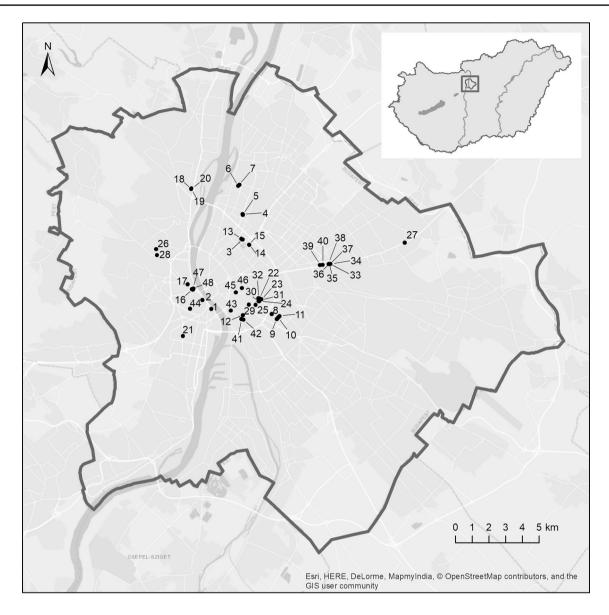


Fig. 1 Position of honey locust trees sampled for their soil seed bank in Budapest, Hungary

**Table 1** Number of honey locust trees sampled in parks of Budapest, Hungary, according to the tree size and park maintenance categories distinguished

Park maintenance	Tree size						
type	Small cf: < 0.60 m	Medium cf: 0.60– 1.20 m	Big cf: > 1.20 m				
Perfect	5	5	7				
Moderate	5	5	5				
Neglected	5	6	5				

cf, trunk circumference at breast height

from this material and were put to viability test. Seed viability was assessed by germination test under room temperature (around 21 °C). Prior to germination test, seeds were mechanically scarified to break physical dormancy by using a plier to cut a small opening on the seed coat without damaging the embryo inside (Strydom et al. 2017) and were sterilized in 5% sodium hypochlorite solution for 10 min. Following sterilization seeds were washed thoroughly with running tap water. For the germination test, seeds were placed into Petri dishes (9 cm diameter) on top of filter paper and moistened with tap water as necessary. Petri dishes containing the seeds were placed into a germination room receiving natural light. After 5 days, the result of germination tests was evaluated. Seeds were considered germinated if the radicle emerged from the seed and had reached



at least 2 mm length. Seeds that imbibed but did not produce a radicle were determined as dead seeds. Seeds remained hard were again mechanically scarified and inspected for an additional 5-day period. In order to obtain values comparable to other studies, seed densities were expressed for 1 m<sup>2</sup> surface area.

For statistical comparisons of soil seed bank densities of different tree size and park maintenance categories, non-parametric tests were used because data did not meet the normality and homoscedasticity assumptions of parametric tests'. Group medians were compared by the means of Kruskal–Wallis test followed by Dunn's post hoc test for multiple comparisons. The relationship between seed bank density and tree trunk circumference was studied by using least-square linear regression. For each statistical test, the significance level was P < 0.05. Statistical analyses were conducted by using the GraphPad InStat (1998) version 3.06 software.

## Results

From the soil samples taken from urban parks under canopies of 48 honey locust tree specimens, altogether 128 seeds were extracted. Following mechanical scarification, 123 seeds germinated, indicating that 96.1% of the seeds were viable and thus can be considered as constituents of the soil seed bank ("Appendix 1").

In perfectly maintained parks, increasing tree size was associated with slightly higher soil seed bank densities: 23.8 seeds/m<sup>2</sup> were encountered for big trees, while only 8.3 seeds/m<sup>2</sup> for medium-sized ones (mean values). Soil seed bank was completely absent under small trees. However, the increasing trend of seed bank toward larger trees was not supported by Kruskal–Wallis test due to relatively high standard deviations (Fig. 2a).

For neglected parks, much higher seed densities occurred in each tree size category. Gradually improved soil seed banks were found from small trees (125.0 seeds/ $\rm m^2$ ) through medium-sized trees (256.9 seeds/ $\rm m^2$ ) to big ones (366.6 seeds/ $\rm m^2$ ), with significant differences between small tree and big tree categories (Median(Me)<sub>small</sub>: 124.98, Me<sub>big</sub>: 374.94, Mean rank difference(MRD): 9.30, P < 0.01) (Fig. 2b).

In moderately maintained parks, there was no significant difference between tree size categories according to seed density in the soil. Even a tendency could not be figured out. Nevertheless, in this comparison, medium-sized trees had the highest number of seeds in the soil (133.3 seeds/m<sup>2</sup>) ("Appendix 2").

Park maintenance intensity resulted in clear and consistent trends for soil seed bank densities in each tree size category. For small trees, the largest soil seed bank was found

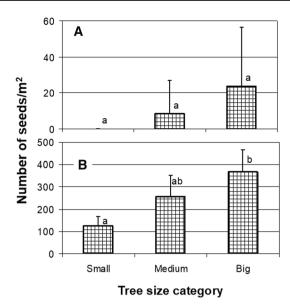


Fig. 2 The mean density of soil seed bank under different size categories of honey locust trees for perfectly maintained (a) and neglected (b) parks in Budapest, Hungary. Whiskers show +1 SD, different letters above columns indicate statistical group membership

in neglected parks (125.0 seeds/m²) followed by moderately maintained parks (16.6 seeds/m²), and there was no soil seed bank at all in perfectly maintained urban parks (Fig. 3a). The increasing value of seed bank density parallel to the increasing rate of park abandonment was statistically supported, as there were significant differences between groups of perfect versus neglected (Me<sub>perfect</sub>: 0.0, Me<sub>neglect</sub>: 124.98, MRD: 8.50, P < 0.01) and moderate versus neglected park maintenance types (Me<sub>moderate</sub>: 0.0, Me<sub>neglect</sub>: 124.98, MRD: 6.50, P < 0.05).

Considering the medium-sized tree category, also the specimens growing in neglected parks had the highest density of soil seed bank (256.9 seeds/m²) under their canopy. For moderately maintained parks, seed bank density was 133.3 seeds/m², followed by perfectly maintained parks where only 8.3 seeds/m² was found (Fig. 3b). Average seed density of the moderate treatment category showed no statistically significant difference with the neighboring groups due to the high standard deviation (SD = 119.3). However, seed bank density of the neglected group was significantly higher than seed bank density of the perfectly maintained parks (Me<sub>neglect</sub>: 249.96, Me<sub>perfect</sub>: 0.0, MRD: 8.72, P<0.01).

For big trees—similarly to the previous two tree size categories—soil seed bank density under honey locust tree canopies showed an increasing trend as park maintenance intensity decreased (Fig. 3c). A seed bank density of 23.8 seeds/m<sup>2</sup> was encountered in perfectly maintained parks while a somewhat higher value, 33.3 seeds/m<sup>2</sup>, was found in moderately maintained parks, yet the difference did not prove significant in statistical test. The highest seed



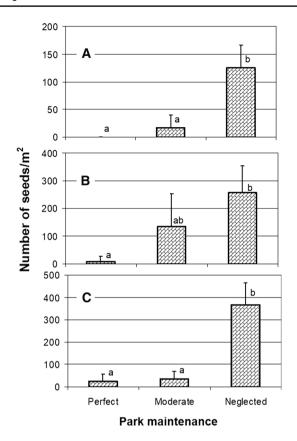
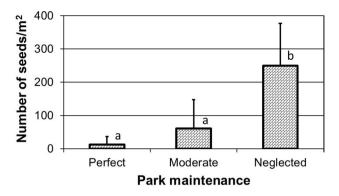


Fig. 3 Soil seed bank densities under the canopy of small-sized (a), medium-sized (b) and big-sized (c) honey locust trees found in urban parks of Budapest (Hungary) with different levels of park maintenance intensity

bank density, 366.6 seeds/m<sup>2</sup>, was observed under big trees in neglected parks, and this was the highest value among all groups examined ("Appendix 2"). Moreover, it was significantly higher than the densities under big trees in moderately or perfectly maintained parks (Me<sub>neglect</sub>: 374.94, Me<sub>moderate</sub>: 41.66, MRD: 7.90, P < 0.05 and Me<sub>neglect</sub>: 374.94, Me<sub>perfect</sub>: 0.0, MRD: 8.93, P < 0.01, respectively).

When the influence of park maintenance intensity was considered for all trees irrespective of their size, the average seed bank densities for perfect, moderate and neglected park maintenance categories were 12.3, 61.1 and 250.0 seeds/m², respectively (Fig. 4). The statistical test showed a remarkably significant increase in seed bank density of neglected parks compared to the other two maintenance categories (Me<sub>neglect</sub>: 229.13, Me<sub>moderate</sub>: 41.66, MRD: 17.73, P < 0.001 and Me<sub>neglect</sub>: 229.13, Me<sub>perfect</sub>: 0.0, MRD: 25.56, P < 0.001).

For neglected parks, significant ( $R^2 = 0.6142$ , P < 0.0005) positive linear relationship was found between trunk circumference (independent variable) and soil seed bank density under canopy (dependent variable) of honey locust trees: seed bank density increased with increasing tree size (Fig. 5).



**Fig. 4** Effect of park maintenance intensity on the soil seed bank density under the canopies of honey locust specimens irrespective to the tree size categories in urban parks of Budapest, Hungary

## **Discussion**

Following mechanical scarification, we found 96.1% germination ratio for honey locust seeds. It is practically the same rate as it was found for this species after chemical scarification (Ferus et al. 2013; Kheloufi 2017) and is among the highest values published for a wide range of taxonomically related hardseeded species (Jones et al. 2016; Siles et al. 2016; Goets et al. 2018). It means that honey locust has a high seed viability that should be considered when assessing its invasiveness.

In the present study, seed bank densities varied between 0 and 500 seeds/m<sup>2</sup> under canopies of honey locust trees. In 17 out of 48 cases, the density was above 100 seeds/m<sup>2</sup>. In a broader comparison, seed bank densities reported in this research are not outstanding since higher densities for certain plant species (including invasive ones) were often found in other studies (Wang et al. 2017; Goets et al. 2018).

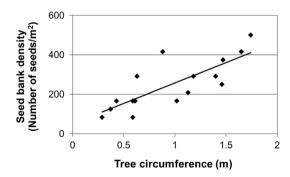


Fig. 5 Linear regression of soil seed bank density under tree canopies against tree trunk circumference of honey locust specimens (n=16) in neglected parks in Budapest, Hungary

But regarding tree species, where seeds are often absent or in a low quantity in the soil seed bank (Honu and Dang 2002;



Nano et al. 2013; Douh et al. 2018), the records for honey locust are still considerable.

Increasing density of soil seed bank under the canopy of woody species parallel with the age of the specimens is a plausible assumption, although rarely documented by published studies (but see Allen et al. 1995; Cseresnyés and Csontos 2012; Aguirre-Acosta et al. 2014). The accumulation of seed bank should be more pronounced for trees known about considerable survival of their seeds, retaining germinability for long periods in the soil, like the case of honey locust (Ferreras et al. 2015). In the present study, the influence of tree size on soil seed bank density was inconsistent. Under park maintenance of either moderate or perfect intensity, seed density did not correlate with tree circumference. Lack of correlation can be caused either by habitat effects on seed production of honey locust specimens, as it has already been described for other species (Rocha and Aguilar 2001; Aguirre-Acosta et al. 2014), or by the park maintenance itself. The latter seems to be the more obvious reason because in neglected parks, where more or less natural accumulation of soil seed bank could take place, significantly higher seed density was observed under big trees than under small ones, and also in the direct correlation between tree circumference and seed density we obtained positive and significant relationship in a linear regression test. Our assumption on tree size-dependent seed bank is thus only partially true because the size effect is considerably influenced by park maintenance activity.

Accordingly, in the comparisons based on park maintenance categories, significant differences in soil seed bank density occurred in each tree size categories, having always the highest densities in neglected parks. This trend was even more pronounced when all trees, irrespective of trunk circumference, were compared according to the park maintenance categories, and the differences were supported with high level of significance. This result is consistent with our hypothesis that park management intensity has a strong effect on the soil seed bank density of honey locust trees.

In evaluating the invasion risk of honey locust, the high capacity for forming soil seed bank demonstrated here must be considered. Species with well-developed soil seed banks are common among invasive plants (Gioria et al. 2012, 2019). Propagule pressure was also emphasized as a key factor of invasion for trees (Lamarque et al. 2011). Persistent soil seed bank—which is considered as a mode of dispersal over time (Grime 1989)—can easily provide such a continuous propagule pressure. For this, the hard-coated, long viable seeds of honey locust provide a typical example.

Additionally, Herron et al. (2007) state that a species is more likely become invasive in new localities, if (1) it has already become invasive in one region outside of its native range; and (2) if it has a large native latitudinal distribution range. The case of honey locust fulfills both of these requirements as being seriously invasive in Argentina (Fernandez et al. 2017), and having a natural range between latitudes N 26° 41' and N 43° 31' in North America (Estrada-Castillón et al. 2002).

Based on the facts mentioned above, honey locust must be considered as a potentially invasive species. Just because it has not yet become seriously invasive in Hungary, it does not mean that it will not become invasive in the future. Lengthy lag-phase has been described for several cases in biological invasions (Crooks and Soulé 1999; Aikio et al. 2010), and we will probably make the right decision regarding honey locust if we assume that we are now witnessing its extended lag-phase. Therefore, we need to consider appropriate methods to prevent honey locust invasion.

Within urban parks regular removal of litter (which includes leaves, twigs and pods) can provide a satisfactory prevention of invasion as it is demonstrated by our results. However, special attention should be made for the proper disposal of the litter containing reproductive propagules of honey locust. In case of improper treatment, e.g., simple displacement of litter from the park or private garden to a peripheral site, even if it is an official landfill site of the settlement, the translocated litter can become a new center of invasion. Plaza et al. (2018) completed a literature review of studies on the vegetation of waste sites and concluded that more than 95% of the species found in rubbish dumps or landfills have invasive character and eight of them are among the 100 worst invasive species in Europe. It is also emphasized that compost from these sites are now being used for soil improvement in agricultural practice, thus seeds of invasive species might be spread to distant locations.

Therefore, gathering of litter must be followed by proper composting techniques. For agricultural weeds, several experiments demonstrated that successful weed seed destruction can be achieved if properly established conditions (first of all temperature and moisture level) prevail during windrow composting (Tompkins et al. 1998; Eghball and Lesoing 2000). Nevertheless, significant differences between species in tolerance to temperature, moisture and composting time were also reported (Grundy et al. 1998), and it underlines the need of species specific experiments to reveal circumstances under which seeds of the invasive species lost viability.



# **Conclusion for future biology**

Examples for successful control of widely established invasive species are rare, although their eradication campaign may work if there is sufficient dedication and persistence, as it has already happened with honey locust in Australia (Csurhes 2004). However, as it is well known, preventing an invasion is much cheaper; therefore, future activities should be focused on prevention. Our findings highlight two kinds of threats. The first is the formation of long-term persistent soil seed bank under the canopy of planted honey locust trees. It can be prevented by perfect maintenance of parks that includes regular collection of fallen pods. The second threat is the mistreatment of litter containing seeds of honey locust. If honey locust's litter is removed in the simplest way into waste sites, and especially when it is deposited illegally in a nearby abandoned area, it involves a significant risk because the alien species can easily grow up on these sites and its established stands may become starting points for further spread (Rusterholz et al. 2012). A recent survey showed that the gardens of property owners who were moving from downtown to suburban areas, consist of predominantly exotic taxa (Ward and Amatangelo 2018). This trend is coupled with parallel shrinking of traditional gardening: the cultivation of fruit trees and vegetable gardens are in decline. Gaggini et al. (2017) demonstrated that the proximity of modern suburban gardens had positive effect on the species number and density of invasive plants in nearby forests. The risk of spread of invasive species is further enhanced by the urban sprawl—an increasingly prevalent process that is typical of modern cities—because it increases the urban edge, i.e., the zone where the city and nature are in contact. Therefore, ecologists must increase awareness of the general public and particularly the garden owners about the problems caused by invasive species in connection with illegal garden waste dumping (Rusterholz et al. 2012). Popularization and promotion of proper composting methods are also pivotal for the future maintenance of suburban gardens. Seeds of different species respond differently to the duration and method of composting; therefore, direct seed biological investigations are recommended for seeds of Gleditsia triacanthos, as well as for seeds of other alien species which tend to become invasive. Finally, it should be introduced as a rule that non-native species should only

be planted in ornamental gardens, where aesthetic reasons make sense their use, while native species should be preferred in other urban green areas (e.g., tree alleys planted to reduce the noise associated with roadway traffic).

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Authors' contributions PC, formulated the research concept, has leading role in planning and implementation, was engaged in field and laboratory work and participated in drafting the manuscript; TK added substantially to the planning and implementation of the research and participated in drafting the manuscript; TF carried out the majority of field and laboratory work and took part in data analysis and interpretation of results; AL was involved in the analysis of data; TH was engaged in the interpretation of results and in drafting the manuscript; JT participated in the analysis of data and interpretation of results.

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**Data accessibility** Data are included in the article's appendices. These appendices are given above the References.

## Compliance with ethical standards

Conflict of interest The authors declare that they have no conflict of interest.

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# **Appendices**



Appendix 1. Geographical location and soil seed bank data of sampled honey locust trees growing in urban parks of Budapest, Hungary

Sample GPS north		GPS east	Park treatment	Trunk size	Location	No. of seeds found	No. of seeds germinated	Seed bank density (seeds/ m <sup>2</sup> )
1	47° 28.838	19° 3.319	Perfect	Big	BME Campus	0	0	0
2	47° 29.126	19° 2.893	Moderate	Medium	Gellért Hill	1	1	41.66
3	47° 31.105	19° 4.754	Perfect	Big	Hősök tere area	0	0	0
4	47° 31.879	19° 4.823	Perfect	Medium	Floridsdorf sétány	0	0	0
5	47° 31.897	19° 4.813	Perfect	Medium	Floridsdorf sétány	0	0	0
6	47° 32.843	19° 4.653	Perfect	Big	Tripolisz tér	1	1	41.66
7	47° 32.814	19° 4.602	Perfect	Big	Tripolisz tér	2	2	83.32
8	47° 28.672	19° 6.183	Neglected	Big	Népliget park	7	7	291.62
9	47° 28.513	19° 6.427	Moderate	Big	Népliget park	2	2	83.32
10	47° 28.564	19° 6.500	Neglected	Small	Népliget park	3	3	124.98
11	47° 28.602	19° 6.557	Neglected	Big	Népliget park	9	9	374.94
12	47° 28.635	19° 4.813	Neglected	Big	Haller utca	7	6	249.96
13	47° 31.077	19° 4.803	Perfect	Big	Városliget city park	1	1	41.66
14	47° 30.902	19° 5.126	Perfect	Big	Városliget city park	0	0	0
15	47° 30.907	19° 5.122	Perfect	Big	Városliget city park	0	0	0
16	47° 29.463	19° 2.445	Perfect	Medium	Esperanto park	0	0	0
17	47° 29.630	19° 2.187	Perfect	Small	Esperanto park	0	0	0
18	47° 32.728	19° 2.368	Neglected	Medium	Boldog Sándor István park	5	4	166.64
19	47° 32.728	19° 2.359	Perfect	Medium	Boldog Sándor István park	1	1	41.66
20	47° 32.711	19° 2.365	Moderate	Small	Boldog Sándor István park	0	0	0
21	47° 27.967	19° 1.972	Moderate	Medium	Bikás Park	0	0	0
22	47° 29.174	19° 5.618	Neglected	Medium	Golgota út	10	10	416.6
23	47° 29.174 47° 29.152	19° 5.708	_	Medium	=	7	7	291.62
23 24	47° 29.132 47° 29.081	19° 5.588	Neglected Moderate	Small	Golgota út	1	1	41.66
24 25	47° 28.971	19° 5.433	Moderate		Biró Lajos utca	0	0	0
25 26	47° 30.766	19 3.433 19° 0.694		Big Medium	Orczy park Herman Ottó utca		5	208.3
20 27			Neglected			6	2	83.32
	47° 30.967	19° 12.523	Neglected	Small	Mátyásföld	2		
28	47° 30.574	19° 0.732	Neglected	Big	Szilágyi E. fasor	13	12	499.92
29	47° 28.975	19° 5.104	Neglected	Small	Ludovika tér	4	4	166.64
30	47° 29.086	19° 5.570	Neglected	Small	Orczy utca	4	4	166.64
31	47° 29.110	19° 5.577	Neglected	Small	Orczy utca	2	2	83.32
32	47° 29.176	19° 5.557	Neglected	Medium	Orczy utca	7	7	291.62
33	47° 30.288	19° 8.987	Moderate	Medium	Kerepesi út	5	5	208.3
34	47° 30.288	19° 8.988	Moderate	Small	Kerepesi út	0	0	0
35	47° 30.277	19° 8.937	Moderate	Small	Kerepesi út	1	1	41.66
36	47° 30.276	19° 8.902	Moderate	Medium	Kerepesi út	7	7	291.62
37	47° 30.279	19° 8.965	Moderate	Medium	Kerepesi út	3	3	124.98
38	47° 30.289	19° 8.915	Neglected	Medium	Kerepesi út	5	4	166.64
39	47° 30.247	19° 8.485	Neglected	Big	Kerepesi út	10	10	416.6
40	47° 30.259	19° 8.606	Perfect	Small	Kerepesi út	0	0	0
41	47° 28.513	19° 4.744	Moderate	Big	Haller Park	1	1	41.66
42	47° 28.492	19° 4.856	Moderate	Big	Haller Park	1	1	41.66
43	47° 28.782	19° 4.238	Moderate	Big	Tinódi Park	0	0	0
44	47° 28.844	19° 2.304	Perfect	Medium	Horticultural University	0	0	0
45	47° 29.367	19° 4.485	Perfect	Small	Horváth Mihály tér	0	0	0
46	47° 29.505	19° 4.767	Moderate	Small	Mátyás tér	0	0	0
47	47° 29.475	19° 2.400	Perfect	Small	Kereszt utca	0	0	0
48	47° 29.493	19° 2.468	Perfect	Small	Kereszt utca	0	0	0



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## undefined(continued)

Sample	GPS north	GPS east	Park treatment	Trunk size	Location	No. of seeds found	No. of seeds germinated	Seed bank density (seeds/ m <sup>2</sup> )
Total number of seeds found, germinated and average seed density				128	123	106.75		

Appendix 2. Mean soil seed bank densities and standard deviations (SD) under the canopy of honey locust trees in urban parks of Budapest, Hungary, considering three groups of park management intensity (perfect, moderate and neglected) and three groups of tree size category (small: circumference < 0.60 m, medium: cf. 0.61–1.20 m, and big: cf. > 1.20 m)

	Perfect	Moderate	Neglect	Management types jointed
Small	Mean value: 0	Mean value: 16.66	Mean value: 124.98	Mean value: 47.21
	SD: 0	SD: 22.82	SD: 41.66	SD: 62.72
	Min: 0.00	Min: 0.00	Min: 83.32	Min: 0.00
	Median: 0.00	Median: 0.00	Median: 124.98	Median: 0.00
	Max: 0.00	Max: 41.66	Max: 166.64	Max: 166.64
	n: 5	n: 5	n: 5	n: 15
Medium	Mean value: 8.33	Mean value: 133.31	Mean value: 256.90	Mean value: 140.60
	SD: 18.63	SD: 119.30	SD: 96.51	SD: 135.1
	Min: 0.00	Min: 0.00	Min: 166.64	Min: 0.00
	Median: 0.00	Median: 124.98	Median: 249.96	Median: 145.81
	Max: 41.66	Max: 291.62	Max: 416.60	Max: 416.60
	n: 5	n: 5	n: 6	n: 16
Big	Mean value: 23.81	Mean value: 33.32	Mean value: 366.61	Mean value: 127.43
	SD: 32.78	SD: 34.85	SD: 99.46	SD: 168.88
	Min: 0.00	Min: 0.00	Min: 249.96	Min: 0.00
	Median: 0.00	Median: 41.66	Median: 374.94	Median: 41.66
	Max: 83.32	Max: 83.32	Max: 499.92	Max: 499.92
	n: 7	n: 5	n: 5	n: 17
Tree size categories jointed	Mean value: 12.25	Mean value: 61.10	Mean value: 249.96	
	SD: 24.49	SD: 86.05	SD: 126.36	
	Min: 0.00	Min: 0.00	Min: 83.32	
	Median: 0.00	Median: 41.66	Median: 229.13	
	Max: 83.32	Max: 291.62	Max: 499.92	
	n: 17	n: 15	n: 16	

n, number of sampled trees

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