# 1 Comparing two methods for estimating floral resource availability for

# 2 insect pollinators in semi-natural habitats

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17	Running title: Pollinator floral resource estimates
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19	Abstract
20	Pollinator and flowering plant interactions play a critical role in maintaining most terrestrial
21	ecosystems, including agroecosystems. Although estimates of floral resource availability are
22	essential to understand plant-pollinator relationships, no generally accepted methodology
23	exists to date. We compared two methods for sampling floral resources in a single meadow.

About every three days, we recorded species lists of insect-pollinated plants with abundance

25 categories assigned to each species (hereafter referred to as scanning) and we counted the 26 flowering shoots in 36  $2 \times 2$  m quadrats (hereafter quadrat sampling). These methods were 27 compared with respect to (i) the number of species detected, (ii) estimated floral resource 28 abundance, and (iii) temporal changes in flowering. With scanning, we found more potential nectar-plant species and species were found earlier than with quadrat sampling. With the 29 30 latter, abundant species were found with higher probability than the scarce. Flower 31 abundances were correlated between the two methods. We predicted that a cover of  $6.3\pm3.6\%$ 32 should be used for an appropriate estimate of flower abundance in our study site, although the 33 optimal cover probably varies across different habitats. Furthermore, flower abundance 34 changed 6% per day compared to the flowering peak. Overall, scanning seems to be more 35 appropriate for detecting presence and the timing of species, while quadrats may provide higher resolution for abundance estimates. Increased sampling coverage and frequency may 36 37 enhance research accuracy and using scanning and quadrat sampling simultaneously may help 38 to optimize research effort. We encourage further development of sampling protocols.

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### 40 Résumé

41 Comparaison de deux méthodes pour estimer la disponibilité des ressources florales 42 pour les insectes pollinisateurs dans des milieux semi-naturels. Les interactions entre 43 pollinisateurs et plantes à fleurs jouent un rôle capital dans le maintien de la plupart des 44 écosystèmes terrestres, y compris les agroécosystèmes. Bien que des estimations de la 45 disponibilité des ressources florales soient essentielles pour comprendre les relations plantespollinisateurs, aucune méthodologie largement acceptée n'existe actuellement. Nous avons 46 47 comparé deux méthodes pour échantillonner les ressources florales dans une prairie. Tous les 48 trois jours environ, nous avons établi des listes de plantes pollinisées par les insectes, avec des

49 catégories d'abondance pour chaque espèce (méthode appelée ici "balayage") et nous avons 50 compté les tiges portant des fleurs dans 36 quadrats de 2 m  $\times$  2 m (méthode de 51 l'échantillonnage par quadrat). Ces deux méthodes ont été comparées pour trois paramètres 52 (1) le nombre d'espèces comptabilisées, (ii) l'abondance estimée de la ressource florale, et (iii) les variations temporelles de la floraison. Dans la méthode du "balayage", nous avons 53 54 trouvé plus de plantes produisant potentiellement du nectar, et les espèces ont été détectées 55 plus tôt qu'avec la méthode de l'échantillonnage par quadrat. Avec cette dernière méthode, les espèces abondantes ont été détectées avec une probabilité plus élevée que les espèces plus 56 57 rares. L'abondance des fleurs était corrélée entre les deux méthodes. Nous avons prédit qu'une 58 couverture de  $6.3 \pm 3.6$  % de la surface pouvait être utilisée pour une estimation correcte de l'abondance des fleurs dans notre site d'étude, bien que la couverture optimale varie 59 probablement selon les milieux. De plus, l'abondance en fleurs a changé de 6% par jour en 60 61 comparaison au pic de floraison. Dans l'ensemble, la méthode du "balayage" semble être plus 62 appropriée pour détecter la présence et la phénologie des espèces, alors que les quadrats 63 peuvent permettre d'avoir une meilleure résolution dans les estimations d'abondance. Une 64 hausse de la surface couverte par l'échantillonnage et de sa fréquence peut améliorer la 65 précision des résultats et l'utilisation combinée des deux méthodes évoquées peut aider à 66 optimiser les efforts de recherche. Nous encourageons vivement au développement de 67 méthodes d'échantillonnage.

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69 Key words: plant-animal interactions, flower food-resource estimate, nectar resources,
70 vegetation sampling methods, scanning, quadrat

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#### 73 Introduction

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75 Plant-pollinator interactions play a critical role in maintaining most terrestrial ecosystems, 76 including agroecosystems. Pollinators consume floral nectar, pollen and oils, while sexual reproduction of plants depends on pollen transfer by pollinators (Goulson 1999; Nicolson et 77 78 al. 2007; Patiny 2012). In natural circumstances, plant-pollinator interactions may rapidly 79 change at the spatio-temporal scale (Kubo et al. 2008; Fründ et al. 2011; Bagella et al. 2013), 80 and food resources can be highly aggregated in space or time (Elzinga et al. 1998; Hatfield & 81 Lebuhn 2007). Floral food resource quality, quantity and production rates show a huge 82 variation with plant species, time of the day, age of flowers and competitors' consumption 83 (Galetto & Bernardello 2004; Nicolson et al. 2007; Hicks et al. 2016). The number of floral 84 resource species, quantity and density of flowers and the amount of food in flowers at least 85 partly determine pollinator abundance, diversity and resource-visit frequency, and are the 86 strongest factors structuring pollinator communities (Potts et al. 2004; Iserbyt et al. 2008; 87 Dennis 2010; Curtis et al. 2015). Floral resources offered to pollinators are advertised via different traits (Goulson 1999; Nicolson et al. 2007; Chartier et al. 2011), and insects use a 88 89 wide range of sensory systems and various cues to navigate at different spatial scales, thus 90 flowers may be perceived very differently by pollinators (Dennis 2010; Dauber et al. 2010; 91 Clarke et al. 2013).

Most field studies reviewed by Szigeti et al. (2016) measure floral resource availability to investigate the relationship between resource availability and pollinator population size or diversity (e.g. Kovács-Hostyánszki et al. 2013); the relationship between resource availability and flower preferences (e.g. Goulson & Darvill 2004); the effect of resource availability on pollinator flight distance between floral patches (e.g. Wolf & Moritz 2008); or the effects of

97 temporal changes in floral composition on the structure of pollination networks (e.g. Bosch et 98 al. 2009). Further studies investigated the link between pollinator conservation and floral 99 resource availability (e.g. Roger et al. 2016), while others investigated long-term 100 compositional changes and their impact on evolutionary processes (e.g. Miller-Struttmann et 101 al. 2015) or developed methods for estimating floral resource availability (e.g. Frankl et al. 102 2005). In order to design sampling, the spatial and temporal scale of the study must be 103 considered and this is ultimately derived from the research question. For example, rather 104 different spatial scales are required if the research question targets foraging strategies such as 105 learning resource cues (Laverty 1994; Chittka et al. 1997) or the diversity of the plant-106 pollinator networks at the landscape scale (Potts et al. 2006; Weiner et al. 2011). Working at 107 different scales requires different allocation of sampling investment and resolution. Szigeti et al. (2016) suggested to specify three research approaches: the *focus*, the *spatial and temporal* 108 109 scale of the study and the unit type of the count variables i.e. the count units to estimate flower resource amounts available for pollinators. Sampling depends on these three 110 111 approaches, e.g. the spatio-temporal scale of measurement has to be adjusted to the foraging ranges and life cycles of focal pollinators (Szigeti et al. 2016). Furthermore, sampling design 112 113 should also be adjusted to the type of the investigated biotopes, e.g. for homogeneous 114 agricultural landscapes less sampling investment, i.e. smaller sampling cover is required than 115 for heterogeneous semi-natural habitats. Taking all the factors influencing pollinator foraging 116 into account is challenging and sampling protocols are rather difficult to design. 117 In order to understand the mechanisms of plant-pollinator interactions, or to establish restoration management, estimates of resource availability are mandatory (Dennis 2010). 118 119 Sampling methods to estimate floral resource availability in the field are important key

120 procedures of plant-pollinator studies (Dicks et al. 2013). Currently no generally accepted

121 methodology exists to estimate floral resource availability (Szigeti et al. 2016). Although 122 there is a vast amount of studies investigating plant sampling protocols for botanical studies. 123 including the comparison and evaluation of different methods (Walker 1970; Everson & 124 Clarke 1987; Vittoz & Guisan 2007; Symstad et al. 2008), recommendations on how to 125 measure floral resource availability for pollinators are scarce, and rather different methods have been applied to estimate resource availability (Frankl et al. 2005; Hegland et al. 2010; 126 127 Hicks et al. 2016; Szigeti et al. 2016). Some studies neglect or do not refer to existing 128 vegetation sampling protocols or the description of methodology is insufficient (Szigeti et al. 129 2016). Sampling floral resource availability is entirely missing in many pollinator studies, 130 often only species lists, i.e. presence-absence data are recorded as an estimate (Kitahara et al. 131 2008), or resource availability is concluded from indirect proxies such as consumption rates (Bakowski & Boroń 2005; Hinners & Hjelmroos-Koski 2009; Aronne et al. 2012). Further 132 133 studies estimate the amount of floral resource, and the count variables can be the abundances 134 of flowers as a proxy (Goulson & Darvill 2004; Kovács-Hostyánszki et al. 2013) or even 135 nectar and pollen amounts (Potts et al. 2004; Hicks et al. 2016). Many pollination studies apply quadrat (sensu Gibson 2002) or transect sampling, commonly used by botanists for 136 137 plant community studies (Elzinga et al. 1998; Gibson 2002; Bonham 2013), but these cover 138 only a small proportion of the area of the study sites (Szigeti et al. 2016). Furthermore, 139 different aspects of sampling, such as study site area and sampling coverage are traded off, 140 reflecting to a limited research investment (Szigeti et al. 2016). These findings show that 141 further field work on optimising sampling techniques, including traditional sampling methods is mandatory to find efficient and feasible methods to provide a sound basis for understanding 142 143 plant-pollinator interactions.

144 We aimed to compare the benefits and pitfalls of two existing methods for sampling floral

resource availability. We repeatedly scanned a single meadow to list insect-pollinated 145 146 flowering plants and assigned an abundance category to each species. In parallel, we used 147 repeated quadrat sampling in the meadow. Data collected with these methods were compared from the following aspects: (i) the number of species detected, since species richness is widely 148 used as a proxy of floral resource availability, (ii) floral resource abundance, as abundance is 149 an estimate for the importance of each species as a potential floral source, and (iii) *temporal* 150 151 changes in flowering, estimated with first and peak flowering and the percent of daily changes 152 in flowering.

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#### 155 Methods

156 Study site

157 We carried out field work in a 0.6 hectare meadow in the Visegrádi-hegység, Hungary (47°44'23"N, 19°03'33"E, at 300 m a.s.l.), between late April to early June in 2011–2013. Our 158 159 study period was adjusted to the flight period of the Clouded Apollo butterfly Parnassius 160 mnemosyne (L. 1758), since we were also interested in its nectar consumption patterns 161 (Szigeti et al. *in prep.*). Therefore, our sampling period is limited from the point of view of 162 other pollinators or flowering plants in general. The vegetation of this colline meadow is 163 characterised as the Arrhenatheretalia association of the Pannonian floristic region. It is rich in 164 insect-pollinated flowering plant species heterogeneously distributed in the meadow. It had 165 previously been regularly mown once a year, but mowing had been abandoned at least for 20 years. The meadow is surrounded by a Turkey oak Quercus cerris (L. 1753) forest. 166 167

## 168 Sampling floral resource availability

We monitored the occurrence and the abundance of insect-pollinated flowering plants. We identified plant species according to Simon (1994). Sampling included *scanning* the entire field and using *quadrats*. Sampling was carried out by JK (scanning in all years) and VS (quadrats in all years and scanning in 2013), every three days on average, depending on weather (sampling dates: Tables 1 and 2), between 9:00-17:00h. We used pathways to avoid destructing the vegetation by trampling, and pathways were evenly scattered over the entire meadow (Figure 1).

176 First, we scanned insect-pollinated flowering plants by walking through the entire meadow 177 along the pathways every three days in about one hour per sampling. We recorded a species 178 list, and estimated the abundance for each species. Estimates were categories upon our overall 179 impression of the meadow's vegetation during the one hour walk. We estimated the levels of flower abundance categories for each species only for open, non-wilted flowers on a rank 180 181 scale for the entire meadow: 1: very scarce; 2: scarce; 3: more or less scarce; 4: more or less 182 abundant; 5: abundant; 6: extremely abundant. We chose these categories because we felt that 183 they could be reliably distinguished in the field and yielded the maximum achievable 184 resolution. The definition and the used number of such categories may vary across studies. We 185 aimed to use approximately equidistant categories. Similar sampling protocols were used in a 186 few studies investigating food availability for insect pollinators (Stefanescu 1997; Goulson & 187 Darvill 2004).

Second, we placed 36 2×2 m permanent quadrats more or less homogeneously in space, by
distributing them along the pathways at random distances from the starting point of the
respective path, placed 30 cm from the edge of the paths (Figure 1). We recorded plant
abundance for each species by counting flowering shoots with open, non-wilted flowers every
three days. Quadrats covered 2.4% of the total meadow area and we assessed ~20 quadrats per

hour. We investigated 36 quadrats because it required the time we could just fit in the
schedule of our other field studies such as observing pollinator behaviour. The 2×2 m quadrat
size is the minimal area generally used for mowed meadows (Lengyel et al. 2016) and the size
should not exceed 2×2 m, since small flowers in a larger quadrat can hardly be detected
without stepping in (Kearns & Inouye 1993).

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#### 199 Number of species

200 We compared the annual number of flowering species recorded with scanning and quadrat 201 sampling. We summed the number of species for each year and we presented the percent of 202 differences found with the two sampling procedures. We also calculated pseudoturnover, the 203 percentage of species overlooked by one observer or method but not the other, for both 204 methods for each year as well as for scanning between the two observers for 2013 (Nilsson & 205 Nilsson 1985; Morrison 2015). Pseudoturnover  $[\%] = (A + B) / (N_a + N_b) \times 100$ ; where A and B = the number of species recorded exclusively by one of the two observers or methods; N<sub>a</sub> 206 207 and  $N_b$  = the total number of species recorded by each observer or method. We compared 208 annual presence-absence data between the two methods by computing Jaccard-indices 209 (Jaccard 1912). This index is a similarity coefficient showing the proportion of the species, 210 found by the two methods, to the number of available species. Furthermore, we applied a 211 binomial Generalized Linear Mixed model (Zuur et al. 2009; Bates et al. 2015) to investigate 212 if species detection per quadrat sampling event was affected by flower abundance, by using 213 the presence or absence of a floral species during an observation event as the response variable. Flower abundance, estimated by scanning, was the explanatory variable, and we 214 215 included year and sampling event as nested random factors into the model to take the dependence of observation events into consideration. We applied this model only for the 216

217 species detected by scanning at a given sampling event.

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#### 219 Floral resource abundance

For abundance analyses, we assigned zero to species abundance when a species was not 220 221 recorded on a given sampling event, but it was detected at least once in a specific year. 222 Furthermore, for quadrat sampling, we averaged the number of flowering shoots of each plant 223 species across the quadrats for each sampling event. We calculated Kendall's correlation 224 coefficients (tau) between the mean number of flowering shoots and *flower abundance* 225 categories provided by scanning.

226 We estimated the reliability of scanning between two recorders (JK & VS). We analysed

227 flower abundance categories with squared weighted kappa (Graham & Jackson 1993). We

give only *kappa*-values, omitting *p*-values, since our data points are not independent. 228

To see how many quadrats would be sufficient to properly estimate the amount of 230 flowering shoots, i.e. flower density, we used the 36 quadrats as a pilot sampling, and we

231 applied the method proposed by Kupper & Hafner (1989) following Elzinga et al. (1998): n =

232  $((Z_{\alpha})^2 \times (s)^2) / (B)^2$ , where n = uncorrected sample size estimate;  $Z_{\alpha} = Z$ -score from standard

normal distribution corresponding to the desired confidence level (Z = 1.96 for 95% CI,  $\alpha$  = 233

234 0.025); s = standard deviation; B = the desired precision expressed as half of the maximum of

235 the acceptable confidence interval width. We set B equal to the mean, since our study site was

236 patchy, providing large variation among quadrats, and we did not aim to reduce this variation

237 because it characterises the entire study site. We corrected sample size (the number of

quadrats) by using the correction table in Elzinga et al. (1998). We also estimated the 238

239 sufficient proportion of the study site area covered with quadrats: we multiplied the estimated

number of quadrats with 4 m<sup>2</sup> (area of one quadrat) and divided it by the area of the whole 240

241 study site (6000 m<sup>2</sup>).

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#### 243 Temporal changes

We compared the days of first flowering and peak flowering between the two sampling 244 245 methods using March dates (the number of days since 1st March). We smoothed species' flowering dynamics with a kernel smoother (bandwidth = 3; Wand 2013). We defined the date 246 247 of first flowering of a species as the 1st day when the estimated flower abundance was higher than 5% of the species' maximum flower abundance. Peak flowering date was the day when 248 249 the estimated value was the maximum. We excluded species that started flowering prior to 250 sampling or that were likely to have peak flowering after the sampling period. We did not use 251 the length of the flowering period and the last day of flowering, since field work was finished 252 before some of the flowers would have started to wilt. We tested if the differences in the 253 number of days between the two methods or the two observers were distinct from zero (one-254 sample exact Wilcoxon-signed-rank test). From kernel smoothed data, we calculated mean 255 daily changes in the number of flowering shoots (quadrat) compared to its peak value for each 256 species.

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We analysed all data in the R statistical environment (R Core Team 2015), using the "sets"
package (Meyer & Hornik 2009) for calculating Jaccard indices, the "lme4" (Bates et al.
2015) for generalized linear mixed models, the "irr" package (Gamer & Lemon 2012) for
computing reliabilities, the "chron" package (James & Hornik 2013) for calculating March
dates, the "KernSmooth" package (Wand 2013) for kernel smoothing, and the "coin" package
(Hothorn et al. 2008) for the Wilcoxon-signed-rank test.

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#### 266 Results

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#### 268 *Number of species*

269 Scanning and quadrat sampling yielded different results finding insect-pollinated floral

270 species, and this difference varied from year to year. With scanning, we found more species,

but some species were found only by quadrat sampling. With the latter method, we found on

average only 60.6% of species found by scanning (2011: 69.0%, quadrat: 40 / scanning: 58

273 species; 2012: 48.9%, 23/47; 2013: 63.8%, 44/69), while scanning allowed us to find on

average 87.3% of the species found by quadrat sampling (2011: 90.9%, scanning: 40 /

275 quadrat: 44 species; 2012: 79.3%, 23/29; 2013: 91.7%, 44/48). Two researchers found almost

the same number of species by scanning (JK: 69 & VS: 73; 2013). Pseudoturnovers were

277 21.6% (2011); 39.5% (2012); 24.8% (2013) for the two methods, and 7.0% for scanning

between the two observers in 2013. The Jaccard index comparing the two sampling methods

279 for presence-absence data was on average 0.56 (2011: 0.65; 2012: 0.43; 2013: 0.60). The

280 probability to detect a species by quadrat sampling significantly increased with increasing

flower abundance (binomial model: exp(coefficients of flower abundance) is 2.34; P < 0.001;

i.e. the odds that quadrat sampling finds a species increased 2.34 times with one category

283 increase in flower abundance estimated with scanning).

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#### 285 Floral resource abundance

286 The number of flowering shoots (quadrat) and flower abundance categories (scanning) were

287 correlated in most cases (2011: *tau* = 0.59; 2012: *tau* = 0.49; 2013: *tau* = 0.66; Figure 2).

288 Correlation varied within and between years (Table 1). The estimates of scanning were

reliable between the two observers ( $\kappa = 0.82$ ; Figure 3). However, reliabilities varied during the sampling period (Table 2).

We predicted that  $94 \pm 54$  (mean  $\pm$  SD, averaged for all species), 11–161 (range) quadrats

with a cover of  $6.3 \pm 3.6\%$  (mean  $\pm$  SD), 0.7-10.7% (range) should be used for an appropriate

- estimate for the 0.6 ha study site.
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#### 295 *Temporal changes*

- 296 First flowers were detected earlier with scanning than quadrat sampling (mean  $\pm$  SD
- difference in days, one-sample Wilcoxon-signed-rank test; 2011:  $7.09 \pm 6.42$ , P < 0.001;
- 298 2012:  $6.17 \pm 3.24$ , P < 0.001; 2013:  $5.78 \pm 5.50$ , P < 0.001). Scanning and quadrat sampling
- gave similar estimates for the date of peak flowering (2011:  $-0.53 \pm 3.57$ , P = 0.781; 2012: -
- 300  $0.96 \pm 3.47$ , P = 0.363; 2013:  $1.83 \pm 6.24$ , P = 0.147). By scanning, JK found first flowers
- 301 slightly but significantly earlier than VS ( $0.83 \pm 4.18$ , P = 0.007), and both observers
- 302 estimated peak flowering similarly ( $-0.43 \pm 7.57$ , P = 0.749). The number of flowering shoots
- 303 changed  $5.8 \pm 5.4\%$  per day [mean  $\pm$  SD].
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306 Discussion

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#### 308 *Number of species*

309 We found more potential floral resource species with scanning than with the quadrat

- 310 method. Pseudoturnover and Jaccard-index values imply that the two methods found plant
- 311 species with different chances. Pseudoturnovers between the two methods were much larger
- than between the two observers for scanning, and for scanning, the inter-observer error was

313 smaller than values found in published botanical studies (Vittoz & Guisan 2007; Symstad et 314 al. 2008; Morrison 2015). Although we failed to detect a few species by scanning that were 315 found by quadrat sampling, this may be due to (i) the difference between the focus of the two methods, i.e. with scanning, researchers may overlook tiny plant species covered with 316 317 vegetation, and (ii) the difference between the sampling persons in their ability to distinguish similar species (Elzinga et al. 1998; Morrison 2015). The odds to find a species with quadrat 318 319 sampling deteriorated with species rarity. We suggest that even abundant species could be overlooked with quadrat sampling, if plants are highly aggregated in space, especially if only 320 321 a small proportion of the entire area is sampled thoroughly. Various methods may detect 322 different species with different probabilities (Walker 1970; Everson & Clarke 1987; Vittoz & 323 Guisan 2007). Since the number of flowering plant species are often increasing with pollinator species richness (Ebeling et al. 2008), some studies listed only the presence of 324 325 species to predict floral resource availability (Kitahara et al. 2008). In contrast, we 326 recommend using quantitative estimates such as flower abundance, because species lists alone 327 are not suitable estimates of floral resource availability (Hegland & Boeke 2006). For estimating the number of species of potential floral resources for a study site, scanning the 328 329 whole area, as we used in this study and similar to the method used by Goulson & Darvill 330 (2004) or using belt transects (Carvell et al. 2007) are more appropriate than quadrat 331 sampling. The presence of frequently visited, although rare floral sources may be noticed with 332 the help of pollinator behaviour.

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## 334 Floral resource abundance

335 Scanning and quadrat sampling yielded similar results, although with a large scatter.

336 Similarly, standard botanical sampling procedures also provide different values for abundance

337 (Walker 1970; Everson & Clarke 1987).

338 The *scanning* method we used, as well as similar methods (e.g. Goulson & Darvill 2004), 339 are subjective sampling techniques with low resolution and high coverage. They are fast and easy to apply at the landscape scale, no special skills or equipment are needed, except to be 340 341 able to identify all flowering plants at the required taxon level (Elzinga et al. 1998). Although identifying all the plant species is a strong skill in some regions, this is necessary for most 342 343 methods. Scanning yields more information than just listing presence/absence, since it 344 provides a rough estimate on abundance, thus enables detecting phenological trends. Studies 345 using similar estimates involved larger sampling areas than direct counts (Szigeti et al. 2016), 346 thus researchers have to decide on either using higher estimate accuracy or higher spatial 347 coverage. Due to higher coverage, scanning compared to quadrat sampling is less sensitive to measurement errors caused by habitat patchiness. Scanning was fairly reliable between the 348 349 recording persons, although reliabilities varied during the sampling period. Reliability could 350 be enhanced with experience. The qualitative measurement of abundance is usually influenced 351 by personal bias (Bonham 2013), e.g. for many botanical methods, there are lots of 352 differences between observers (Walker 1970; Vittoz & Guisan 2007) and this bias is expected 353 to increase with the method's subjectivity (Morrison 2015). Some source of bias may come 354 from ill-defined categories and lack of experience. An other important source of bias for this 355 method is that the detectability of different flower species is different (Bonham 2013; 356 Morrison 2015), e.g. tiny Myosotis discolor (Pers. 1798) flowers are less likely to be detected 357 than the large flowers of Inula hirta (L. 1753). A further drawback of scanning is its inability to estimate data scatter per sampling event per species, since it yields only one data point for 358 359 each species per sampling, although partitioning the trajectory scanned may refine this 360 procedure and provide more accurate data.

361 Overall, we think that scanning yields a reasonably good estimate for abundance. With 362 careful standardization of the sampling procedure, this method could be significantly 363 improved. Carvell et al. (2007) associated estimated density ranks to rough count estimates, using the following ranking: 0 shoots = 0 rank;  $\leq 25$  shoots = 1; 26–200 shoots = 2 etc. We 364 recommend using even-spaced or even-spaced logarithmic categories (Stefanescu 1997; 365 366 Elzinga et al. 1998), because such data can be handled easier mathematically compared to 367 uneven categories, and may considerably increase the reliability of scanning for a relative density estimate. The number of categories should be adapted to the study. Hahn & Scheuring 368 369 (2003) claim that the best estimates can be achieved at an optimum number of categories, 370 recommending e.g. ten categories. Although ten categories provide high resolution relative to 371 the method's simplicity, this requires even-spaced sampling, especially for small datasets, 372 since a variable with even-spaced categories may be used as a numerical covariate in 373 statistical modeling. We recommend using scanning for rough-scale estimates when timeconsuming quadrat sampling is not feasible (e.g. mapping good honey-plant meadows for 374 375 beekeepers). Further investment may be required due to the difficulties to estimate the number 376 of available flowers for species with different inflorescence structures, or nectar amount 377 estimates.

Using *quadrats* to estimate flowering plant abundance provides much higher resolution with the cost of requiring much more research investment than scanning. For example, quadrat or transect sampling need a lot of key sampling design decisions, such as the arrangement of plots, boundary definitions, plot size and shape, counting unit, and estimate precision, and presumes a preliminary survey. Most of these are thoroughly investigated in the botanical methodology literature (Elzinga et al. 1998; Gibson 2002; Bonham 2013) and should be used for floral resource estimates as well. Among these, the estimate on the

385 necessary number of quadrats is important in our case. Insufficient quadrat cover yields 386 biased data (Hicks et al. 2016) especially on rare and clumped species. On the one hand, 387 clumped species can be sampled with less bias if quadrat sizes are increased or their shape 388 varied, e.g. from square to elongated rectangle (Elzinga et al. 1998; Bonham 2013). Our 389 results indicate that more quadrats should be applied than generally used in most case studies. 390 Compared to the median 0.69% for 158 studies reviewed by (Szigeti et al. 2016), we covered 391 2.4% of our study site and this turned out to be insufficient to estimate the abundance of most 392 species. Approximately 6.3% of our heterogeneous study site should have been covered for a 393 reliable estimate, similar to the recommendations (5-10%) for standard botanical studies 394 (Bonham 2013). We did not find other recommendations, but the appropriate cover is 395 probably highly dependent on vegetation heterogeneity. Recommendations for different types of vegetation based upon thorough studies would be helpful to those having limited capacity 396 397 to assess floral composition. However, the proper procedure is to run a pilot study to estimate sampling cover prior to sampling. In natural biotopes, flowers are often aggregated (Elzinga et 398 399 al. 1998; Hatfield & Lebuhn 2007), thus the minimum number of quadrats required may be 400 rather large. The formula of Kupper-Haffner's method implies, as well as Hegland et al. 401 (2010) and Bonham (2013) suggest that the rare and/or aggregated species increase the 402 number of sufficient quadrats considerably. Since different aspects of research investment are 403 traded off (Szigeti et al. 2016), researchers need to find the overall optimum for achieving 404 sufficiently high resolution and coverage simultaneously. Nevertheless, in homogeneous 405 biotopes such as agricultural plots, especially in crops, even a smaller number of sampling 406 units may be sufficient and quadrat sampling is recommended.

407 Quadrat cover may also depend on the type of the count variables, not only on species
408 richness and biotope heterogeneity. Choosing the appropriate count variables are essential,

409 since if the estimated nectar and pollen production of flower units are known, measuring 410 proxies could provide reliable data of the total food production of an entire meadow, e.g the 411 number of flowers and flower size are related to nectar and/or pollen amount (Stanton & Preston, 1988; Galetto & Bernardello 2004; Potts et al. 2004; Torné-Noguera et al. 2014; 412 413 Hicks et al. 2016; but Wäckers 2004). Researchers use various count variables to assess resource availability (Szigeti et al. 2016), such as estimates on sugar and amino acid contents 414 415 of nectar and pollen (Zimmermann & Pleasants 1982; Kearns & Inouye, 1993), often using 416 proxies (Hegland & Totland 2005). Using a count variable simple to estimate, such as the 417 categorical estimates on the number of flowers, including floral traits like nectar and/or pollen 418 amounts were also measured for a couple of individuals in all plant species and the 419 measurements of these floral traits were extrapolated to study site (Hegland & Totland 2005; Hicks et al. 2016). This provide more reliable estimates on food availability than using solely 420 421 proxies such as flower units, although the effective sampling of very small volumes of nectar 422 or pollen in a sufficient amount for measurements is complicated and labour-intensive 423 (Tepedino & Stanton, 1982; Morrant et al. 2009; Hicks et al. 2016). 424 Walker (1970) found that with increasing time spent on a sampling event, precision 425 declined, and the limit of reliable botanical sampling time was 4–5 h per day in a tropical 426 study. Kearns & Inouye (1993) took 12 hours for counting the number of flowers in 25 2×2 m 427 quadrats. We needed one hour to sample 20 quadrats for counting flowering shoots. We used 428 36 quadrats since we were time-constrained, and the sufficient quadrat estimates predicted on 429 average 94 quadrats necessary to use, i.e. about 5 hours of sampling in a 0.6 hectare meadow. 430 In contrast, scanning of our 0.6 ha meadow took one hour. Scanning might be accelerated and 431 used also for much larger meadows effectively. We suggest that using both quadrats and 432 scanning may help to optimise the trade-off between spatio-temporal resolution and sample

433 coverage for a given amount of research investment (Szigeti et al. 2016). Because quadrat
434 sampling detects some species poorly and probably estimates the abundance with larger
435 scatter for rare species, relative densities for the more or less abundant species can be
436 estimated with quadrats whereas rare species can be detected with scanning.

437

### 438 Temporal changes

439 Scanning estimated the appearance of first flowers earlier than quadrat sampling, although 440 there was a large variation among species. We found that recording presence-absence of 441 flowering species in an entire meadow potentially detected most flowering species earlier than 442 quadrats. Earlier detection with scanning is due to small quadrat cover relative to the spatial heterogeneity of the vegetation. Furthermore, one observer found first flowers slightly earlier 443 than the other, and it may reflect how this method is influenced by field experience. 444 445 Nevertheless, we did not find differences in detecting peak flowering either between the 446 observers or the two methods. Investigating peak flowering might yield more robust results 447 than first flowering dates (see also Miller-Rushing et al. 2008). Scanning in small meadows potentially detects all flowering species even in case of extreme species heterogeneity. In 448 449 contrast, quadrat sampling is more appropriate to estimate the change over time in the relative 450 densities across species, due to its higher resolution. In order to get better estimates for first, 451 peak and last flowering dates, we recommend using kernel smoothing (Wand 2013), if 452 sampling is carried out frequently enough for data imputation. The daily changes in flower 453 abundances were 6% in our study, and it can be very different across various biotopes or seasons. Seasonal changes in flower abundances were also found by others (Kubo et al. 2008; 454 455 Bagella et al. 2013; Hicks et al. 2016). Pollinators necessarily follow the changes in the temporal distribution of flowering phenology (Goulson 1999; Potts et al. 2004; Kubo et al. 456

457 2008). Even, so rapid changes in floral resources are overlooked in many studies (Szigeti et
458 al. 2016), although this should be taken into account when planning sampling frequency. In
459 contrast, many studies used >30 days for sampling intervals (Szigeti et al. 2016).

460 Furthermore, time elapsed between sampling events often increases with the number of sites,

461 indicating that sampling frequency is determined by research effort constraints (Szigeti et al.

462 2016). We argue that this typical trade-off between spatial and temporal representativeness

463 could be overcome by combining different methods with either a high spatial or high temporal464 resolution.

We suggest that scanning presence-absence of flowering species in an entire meadow might detect some species to start blooming earlier than quadrats or transects, if these latter cover only a small proportion of the entire study area. In contrast, abundance estimates, e.g. by quadrat sampling, may be more suitable to estimate the change over time in relative densities across species, due to its higher resolution.

470

#### 471 Conclusion

We compared two sampling methods to estimate floral resource availability. Our data were 472 473 collected in one single meadow, thus conclusions have to be treated with caution. We found 474 that scanning and quadrat sampling yielded more or less similar results for estimating abundance, although the two methods were different in detecting species. We suggest that 475 476 increased sampling coverage and frequency may enhance research accuracy and using 477 scanning and quadrat sampling simultaneously may help to optimize research effort. Although a large variety of methods for botanical sampling is available (Elzinga et al. 1998; Gibson 478 479 2002; Bonham 2013), there are no appropriate, standardized, widely used methods to measure pollinators' floral food resources, even if some recommendations are available in Frankl et al. 480

481 (2005); Hegland et al. (2010); Szigeti et al. (2016); Hicks et al. (2016). With this case study
482 we would like to initiate further development of sampling protocols. Future studies should
483 address evaluating several sampling protocols in order to find or develop more appropriate
484 sampling techniques to estimate pollinators' floral resource availability.

485

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# 494 References

- 496 Aronne G, Giovanetti M, Guarracino MR, de Micco V. 2012. Foraging rules of flower
  497 selection applied by colonies of Apis mellifera: ranking and associations of floral
  498 sources. Functional Ecology. 26:1186–1196.
- Bagella S, Satta A, Floris I, Caria MC, Rossetti I, Podani J. 2013. Effects of plant community
   composition and flowering phenology on honeybee foraging in Mediterranean sylvo pastoral systems. Applied Vegetation Science. 16:689–697.
- Bąkowski M, Boroń M. 2005. Flower visitation patterns of some species of Lycaenidae
   (Lepidoptera). Biological Letters. 42:13–19.
- Bates D, Maechler M, Bolker B, Walker S. 2015. Fitting linear mixed-effects models using 
   {lme4}.
- 506 Bonham CD. 2013. Measurements for terrestrial vegetation. Fort Collins, Colorado: Wiley.
- Bosch J, González AMM, Rodrigo A, Navarro D. 2009. Plant-pollinator networks: adding the
   pollinator's perspective. Ecology Letters. 12:409–19.
- Carvell C, Meek WR, Pywell RF, Goulson D, Nowakowski M. 2007. Comparing the efficacy
   of agri-environment schemes to enhance bumble bee abundance and diversity on arable
   field margins. Journal of Applied Ecology. 44:29–40.
- 512 Chartier M, Pélozuelo L, Gibernau M. 2011. Do floral odor profiles geographically vary with
  513 the degree of specificity for pollinators? Investigation in two sapromyophilous Arum
  514 species (Araceae). Annales de la Société entomologique de France (N.S.). 47:71–77.
- 515 Chittka L, Gumbert A, Kunze J. 1997. Foraging dynamics of bumble bees: correlates of
   516 movements within and between plant species. Behavioral Ecology. 8:239–249.
- 517 Clarke D, Whitney H, Sutton G, Robert D. 2013. Detection and learning of floral electric
   518 fields by bumblebees. Science. 340:66–69.
- Curtis RJ, Brereton TM, Dennis RLH, Carbone C, Isaac NJB. 2015. Butterfly abundance is
   determined by food availability and is mediated by species traits. Journal of Applied
   Ecology. 52:1676–1684.
- 522 Dauber J, Biesmeijer JC, Gabriel D, Kunin WE, Lamborn E, Meyer B, Nielsen A, Potts SG,
  523 Roberts SPM, Sõber V, et al. 2010. Effects of patch size and density on flower visitation
  524 and seed set of wild plants: a pan-European approach. Journal of Ecology. 98:188–196.
- 525 Dennis RLH. 2010. A resource-based habitat view for conservation: butterflies in the British
   526 landscape. Oxford: Wiley-Blackwell.
- 527 Dicks L, Abrahams A, Atkinson J, Biesmeijer J, Bourn N, Brown C, Brown MJF, Carvell C,
   528 Connolly C, Cresswell JE, et al. 2013. Identifying key knowledge needs for evidence 529 based conservation of wild insect pollinators: a collaborative cross-sectoral exercise.
   530 Insect Conservation and Diversity. 6:435–446.
- Ebeling A, Klein A-M, Schumacher J, Weisser WW, Tscharntke T. 2008. How does plant
  richness affect pollinator richness and temporal stability of flower visits? Oikos.
  117:1808–1815.

- Elzinga CL, Salzer DW, Willoughby JW. 1998. Measuring & monitoring plant populations.
   Denver: U.S. Dept. of the Interior, Bureau of Land Management.
- Everson CS, Clarke GPY. 1987. A comparison of six methods of botanical analysis in the
   montane grasslands of Natal. Vegetatio. 73:47–51.
- Frankl R, Wanning S, Braun R. 2005. Quantitative floral phenology at the landscape scale: Is
  a comparative spatio-temporal description of "flowering landscapes" possible? Journal
  for Nature Conservation. 13:219–229.
- Fründ J, Dormann CF, Tscharntke T. 2011. Linné's floral clock is slow without pollinators flower closure and plant-pollinator interaction webs. Ecology Letters. 14:896–904.
- 543 Galetto L, Bernardello G. 2004. Floral nectaries, nectar production dynamics and chemical
   544 composition in six Ipomoea species (Convolvulaceae) in relation to pollinators. Annals
   545 of Botany. 94:269–280.
- 546 Gamer M, Lemon J. 2012. irr: Various coefficients of interrater reliability and agreement.
- 547 Gibson DJ. 2002. Methods in comparative plant population ecology. Gibson DJ, editor.
   548 Oxford: Oxford University Press.
- Goulson D, Darvill B. 2004. Niche overlap and diet breadth in bumblebees; are rare species
  more specialized in their choice of flowers? Apidologie. 35:55–63.
- Goulson D. 1999. Foraging strategies of insects for gathering nectar and pollen, and
   implications for plant ecology and evolution. Perspectives in Plant Ecology, Evolution
   and Systematics. 2:185–209.
- Graham P, Jackson R. 1993. The analysis of ordinal agreement data: beyond weighted kappa.
   Journal of Clinical Epidemiology. 46:1055–1062.
- Hahn I, Scheuring I. 2003. The effect of measurement scales on estimating vegetation cover: a
   computer-assisted experiment. Community Ecology. 4:29–33.
- Hatfield R, Lebuhn G. 2007. Patch and landscape factors shape community assemblage of
  bumble bees, Bombus spp. (Hymenoptera: Apidae), in montane meadows. Biological
  Conservation. 139:150–158.
- Hegland SJ, Boeke L. 2006. Relationships between the density and diversity of floral
  resources and flower visitor activity in a temperate grassland community. Ecological
  Entomology. 31:532–538.
- Hegland SJ, Dunne J, Nielsen A, Memmott J. 2010. How to monitor ecological communities
  cost-efficiently: the example of plant–pollinator networks. Biological Conservation.
  143:2092–2101.
- Hegland SJ, Totland Ø. 2005. Relationships between species' floral traits and pollinator
  visitation in a temperate grassland. Oecologia. 145:586–594.
- Hicks DM, Ouvrard P, Baldock KCR, Baude M, Goddard MA, Kunin WE, Mitschunas N,
  Memmott J, Morse H, Nikolitsi M, et al. 2016. Food for pollinators: quantifying the
  nectar and pollen resources of urban flower meadows. Plos One. 11: e0158117.
- Hinners S, Hjelmroos-Koski M. 2009. Receptiveness of foraging wild bees to exotic
   landscape elements. The American Midland Naturalist. 162:253–265.
- 574 Hothorn T, Hornik K, van de Wiel MA, Zeileis A. 2008. Implementing a class of permutation

- 575 tests: the {coin} package. Journal of Statistical Software. 28:1–23.
- 576 Iserbyt S, Durieux E-A, Rasmont P. 2008. The remarkable diversity of bumblebees
  577 (Hymenoptera: Apidae: Bombus) in the Eyne Valley (France, Pyrénées-Orientales).
  578 Annales de la Société entomologique de France (N.S.). 44:211–241.
- Jaccard P. 1912. The distribution of the flora in the alpine zone. New Phytologist. 11:37–50.
- 580 James D, Hornik K. 2013. chron: Chronological objects which can handle dates and times.
- Kearns CA, Inouye DW. 1993. Techniques for pollination biologists. Niwot, Colorado:
   University Press of Colorado.
- Kitahara M, Yumoto M, Kobayashi T. 2008. Relationship of butterfly diversity with nectar
  plant species richness in and around the Aokigahara primary woodland of Mount Fuji,
  central Japan. Biodiversity and Conservation. 17:2713–2734.
- Kovács-Hostyánszki A, Haenke S, Batáry P, Jauker B, Báldi A, Tscharntke T, Holzschuh A.
  2013. Contrasting effects of mass-flowering crops on bee pollination of hedge plants at different spatial and temporal scales. Ecological Applications. 23:1938–1946.
- Kubo M, Kobayashi T, Kitahara M, Hayashi A. 2008. Seasonal fluctuations in butterflies and
   nectar resources in a semi-natural grassland near Mt. Fuji, central Japan. Biodiversity
   and Conservation. 18:229–246.
- Kupper L, Hafner K. 1989. How appropriate are popular sample size formulas? The American
   Statistician. 43:101–105.
- Lengyel A, Illyés E, Bauer N, Csiky J, Király G, Purger D, Botta-Dukát Z. 2016.
  Classification and syntaxonomical revision of mesic and semi-dry grasslands in Hungary. Presila. 88:201–228.
- Laverty TM. 1994. Bumble bee learning and flower morphology. Animal Behaviour. 47:531–
  545.
- Meyer D, Hornik K. 2009. Generalized and customizable sets in {R}. Journal of Statistical
   Software. 31:1–27.
- Miller-Rushing AJ, Inouye DW, Primack RB. 2008. How well do first flowering dates
   measure plant responses to climate change? The effects of population size and sampling
   frequency. Journal of Ecology. 96:1289–1296.
- Miller-Struttmann N, Geib J, Franklin JD, Kevan PG, Holdo RM, Ebert-May D, Lynn AM,
  Kettenbach JA, Hedrick E, Galen C. 2015. Functional mismatch in a bumble bee
  pollination mutualism under climate change. Science. 349:75–78.
- Morrant DS, Schumann R, Petit S. 2009. Field methods for sampling and storing nectar from
   flowers with low nectar volumes. Annals of Botany. 103:533–542.
- Morrison LW. 2015. Observer error in vegetation surveys: a review. Journal of Plant Ecology.
   DOI: 10.1093/jpe/rtv077.
- 611 Nicolson SW, Nepi M, Pacini E. 2007. Nectaries and nectar. Dordrecht: Springer.
- 612 Nilsson IN, Nilsson SG. 1985. Experimental estimates of census efficiency and
- 613 pseudoturnover on islands: error trend and between-observer variation when recording 614 vascular plants. The Journal of Ecology. 73:65–70.

- Patiny S. 2012. Evolution of plant-pollinator relationships. Cambridge: Cambridge University
   Press.
- Potts SG, Vulliamy B, Roberts S, O'Toole C, Dafni A, Ne'eman G, Willmer PG. 2004. Nectar
  resource diversity organises flower-visitor community structure. Entomologia
  Experimentalis et Applicata. 113:103–107.
- Potts SG, Petanidou T, Roberts S, O'Toole C, Hulbert A, Willmer P. 2006. Plant-pollinator
  biodiversity and pollination services in a complex Mediterranean landscape. Biological
  Conservation. 129:519–529.
- 623 R Core Team. 2015. R: A language and environment for statistical computing.
- Roger N, Moerman R, Carvalheiro LG, Aguirre-Guitiérrez J, Jacquemart AL, Kleijn D,
  Lognay G, Moquet L, Quinet M, Rasmont P, et al. 2016. Impact of pollen resources drift
  on common bumblebees in NW Europe. Global Change Biology. DOI:
  10.1111/gcb.13373
- 628 Simon T. 1994. A magyarországi edényes flóra határozója: harasztok-virágos növények
  629 [Identification guide to the vascular plants of Hungary: ferns–flowering plants]. 5th ed.
  630 Budapest: Nemzeti Tankönyvkiadó Rt.
- 631 Stanton ML, Preston RE. 1988. Ecological consequences and phenotypic correlates of petal
  632 size variation in wild radish, Raphanus sativus (Brassicaceae). American Journal of
  633 Botany. 75:528–539.
- 634 Stefanescu C. 1997. Migration patterns and feeding resources of the Painted Lady butterfly,
  635 Cynthia cardui (L.)(Lepidoptera, Nymphalidae) in the northeast of the Iberian
  636 peninsula. Miscel·lània Zoològica. 20:31–48.
- 637 Symstad AJ, Wienk CL, Thorstenson AD. 2008. Precision, repeatability, and efficiency of two
  638 canopy-cover estimate methods in northern Great Plains vegetation. Rangeland Ecology
  639 & Management. 61:419–429.
- Szigeti V, Kőrösi Á, Harnos A, Nagy J, Kis J. 2016. Measuring floral resource availability for
   insect pollinators in temperate grasslands a review. Ecological Entomology. 41:231–
   240
- Tepedino V, Stanton N. 1982. Estimating floral resources and flower visitors in studies of
   pollinator-plant communities. Oikos. 38:384–386.
- Torné-Noguera A, Rodrigo A, Arnan X, Osorio S, Barril-Graells H, Da Rocha-Filho LC,
  Bosch J. 2014. Determinants of spatial distribution in a bee community: Nesting
  resources, flower resources, and body size. PLoS ONE. 9:e97255
- 648 Vittoz P, Guisan A. 2007. How reliable is the monitoring of permanent vegetation plots? A test
  649 with multiple observers. Journal of Vegetation Science. 18:413–422.
- Wäckers F. 2004. Assessing the suitability of flowering herbs as parasitoid food sources:
   flower attractiveness and nectar accessibility. Biological Control. 29:307–314.
- Walker BH. 1970. An evaluation of eight methods of botanical analysis on grasslands in
   Rhodesia. Journal of Applied Ecology. 7:403–416.
- 654 Wand M. 2013. KernSmooth: Functions for kernel smoothing for Wand & Jones (1995).
- 655 Weiner CN, Werner M, Linsenmair KE, Blüthgen N. 2011. Land use intensity in grasslands:

- changes in biodiversity, species composition and specialisation in flower visitor
   networks. Basic and Applied Ecology. 12:292–299.
- Wolf S., Moritz R. 2008. Foraging distance in Bombus terrestris L. (Hymenoptera: Apidae).
   Apidologie. 39:419–427.
- Zimmerman M, Pleasants J. 1982. Competition among pollinators: quantification of available
   resources. Oikos. 38:381–383.
- Zuur A, Ieno EN, Walker N, Saveliev AA, Smith GM. 2009. Mixed effects models and
   extensions in ecology with R. New York: Springer New York.

Table 1. Kendall's *tau* for the estimates between scanning (daily ranks) and quadrat sampling
(abundance) by year. We show March dates (the number of days since 1<sup>st</sup> March) of sampling events
for both methods. Dates may be slightly different between the methods.

2011			2012			2013		
date		tau	date		tau	date		tau
scanning	quadrat		scanning	quadrat		scanning	quadrat	
59	59	0.59	54	54	0.34	55	56	0.69
64	64	0.68	58	58	0.67	59	60	0.67
67	67	0.61	61	61	0.48	65	63	0.65
70	70	0.69	64	64	0.49	68	67	0.67
73	73	0.66	67	68	0.38	71	70	0.66
75	75	0.65	70	71	0.44	74	73	0.71
79	79	0.54	73	74	0.54	77	76	0.63
82	82	0.51	77	77	0.56	80	80	0.72
84	84	0.48	80	80	0.55	83	83	0.67
87	87	0.54				86	87	0.65
89	89	0.54				92	91	0.67
						95	94	0.64
						100	100	0.57

Table 2. Reliabilities for flower abundance estimates between the two sampling persons, JK and
VS: squared weighted *kappa* for categorical data. We show March dates (the number of days since
1<sup>st</sup> March) of sampling events for both recorders. Dates may be slightly different between recorders.

	2013	
da	kappa	
JK	VS	
55	56	0.58
59	60	0.68
65	63	0.82
68	67	0.81
71	70	0.89
74	73	0.89
77	76	0.84
80	80	0.87
83	83	0.83
86	87	0.86
92	91	0.81
95	94	0.85
100	100	0.82

674 Legends

675

Figure 1. Study site. White patches represent shrubs, forest or individual trees, dotted lines are thepathways, dark squares show quadrat positions.

678

**Figure 2.** Relationships between flower abundance categories (scanning) and the number of flowering shoots by years. Each dot represents one plant species at a single sampling event. Dots are jittered horizontally and vertically for better visualisation.

- **Figure 3.** Scatter plot to visualize reliability for estimating flower abundance categories for the two
- 684 observers for scanning. Each dot represents a plant species found at a specific sampling event. Dots
- are jittered horizontally and vertically for better visualisation.





