

This manuscript is contextually identical with the following published paper:

Mozsár A; Specziár A; Battonyai I; Borics G; Görgényi J; Horváth H; Présing M; G.-
Tóth L; Vitál Z; Boros G (2017) **Influence of environmental factors and individual
traits on the diet of non-native hybrid bigheaded carp (*Hypophthalmichthys
molitrix* × *H. nobilis*) in Lake Balaton, Hungary.** - *Hydrobiologia*, 794 (1) pp. 317–
332. <https://doi.org/10.1007/s10750-017-3137-z>

The original published PDF available in this website:

<https://link.springer.com/article/10.1007%2Fs10750-017-3137-z>

**Influence of environmental factors and individual traits on the diet of non-
native hybrid bigheaded carp (*Hypophthalmichthys molitrix* × *H. nobilis*) in
Lake Balaton, Hungary**

Attila Mozsár^{1*a}, András Specziár¹, Izabella Battonyai¹, Gábor Borics², Judit Görgényi²,
Hajnalka Horváth¹, Mátyás Présing¹, László G.-Tóth¹, Zoltán Vitál¹, Gergely Boros¹

¹Balaton Limnological Institute, MTA Centre for Ecological Research, Klebelsberg K. str. 3.,
H-8237 Tihany, Hungary

²Tisza Research Department, MTA Centre for Ecological Research, Bem sq. 18/C, H-4026
Debrecen, Hungary

*Corresponding author: Tel.: +36 66515300; email: mozsar.attila@haki.naik.hu

^aCurrent address: Research Institute for Fisheries and Aquaculture, National Agricultural
Research and Innovation Centre, Anna-liget 8., H-5540 Szarvas, Hungary

Abstract Planktivorous silver carp and bighead carp (collectively, the bigheaded carps) have
been stocked worldwide and their invasion has caused severe impacts on many freshwater
ecosystems. Exploiting the chance provided by the specific hybrid bigheaded carp stock in
Lake Balaton (Hungary) covering the entire morphological range between the two species
(including gill raker morphology), we implemented a comprehensive study (1) to reveal the

feeding habits of hybrid bigheaded carps living in a mesotrophic, lacustrine habitat; and (2) to assess how biotic and abiotic environmental factors and gill raker morphology affect diet composition. We found that all bigheaded carps utilized primarily zooplankton and neglected the scarce and inefficiently digestible phytoplankton, irrespective of gill raker morphology. Moreover, we observed strikingly high levels of inorganic debris consumption, but the proportion of inorganic matter in the guts was not associated directly with the concentration of suspended inorganic particles. Variance in the diet composition of bigheaded carps was related mostly to environmental factors, including the wind-induced resuspension of inorganic particles and seasonally variable availability of food resources. In conclusion, the effects of abiotic environmental factors and available food resources could overwhelm the effect of gill raker morphology in shaping the feeding habits of bigheaded carps.

Keywords: Asian carp, filter-feeding, hybrid fishes, introduced fish species, planktivory, zooplankton

Introduction

Silver carp (*Hypophthalmichthys molitrix*) and bighead carp (*H. nobilis*), collectively known as bigheaded carps, are cyprinid fishes native to the large rivers and lakes of eastern Asia (Kolar et al., 2007). From the early 1950s these filter-feeding fish species have been introduced worldwide to improve the water quality (because bigheaded carps were considered as effective biological control agents for algal blooms; Cremer & Smitherman, 1980; Xie & Liu, 2001), and to increase fishery yields (Kolar et al., 2007). However, recent studies demonstrated that bigheaded carps can adversely affect water quality, both by accelerating nutrient turnover and by consuming zooplankton which decreases the top-down control on phytoplankton (Yang et al., 1999; Borics et al., 2000). Moreover, bigheaded carps can cause a decline in fitness and condition factor in native fish populations (Irons et al., 2007; Sampson et al., 2009) and exert strong effects on community structure (Solomon et al., 2016). Therefore, the presence of bigheaded carps outside their native range is now considered a serious ecological threat (Cooke et al., 2009) and their stocking to natural waters has been prohibited or regulated in several countries (e.g., USA, Hungary; Kolar et al., 2007; Boros et al., 2014). In spite of the strict regulations, the spread of these invasive species is ongoing and their biomass is still high in many invaded habitats (Hayer et al., 2014 a,b).

Several previous studies dealing with the ecological impacts of non-native fish species emphasized that the detrimental effects of invaders are exerted mostly via food-web alterations or direct resource competition with the native fishes (e.g., Kleef et al., 2008; Khan & Panikkar, 2009; Britton et al., 2010; Sass et al., 2014). Bigheaded carps consume predominantly planktonic organisms (both phyto- and zooplankton) and thus they may compete for food with nearly all fish species at early life stages (Sass et al., 2014). Occupying a key trophic position in aquatic ecosystems, bigheaded carps can unfavourably affect the whole native fish community (Calkins et al., 2012). Accordingly, reliable assessment of the

ecological effects of bigheaded carps requires a better understanding of their feeding habits and their interspecific, interindividual and habitat-related variability.

Bigheaded carps use their filtering apparatus (gill rakers) to harvest plankton or any other suspended particles that overlap in size with potential food resources. The general notion is that, in filter-feeding fishes, the filtering efficiency (i.e., size range of consumed food items) is primarily determined by the morphology of gill rakers (Lieberman, 1996; Kolar et al., 2007), which shows substantial differences between bighead carp and silver carp. Bighead carp have long, thin gill rakers which form a comb-like structure, while the gill rakers of silver carp have a spongy appearance due to the fusion of gill filaments (Kolar et al., 2007). The mesh-size of this fused, sponge-like apparatus ranges between 12 and 26 μm (Hampl et al., 1983; Lu et al., 2002), while the comb-like gill raker is characterised with larger mesh-sizes and is specialised to harvest particles larger than 50 μm (Kolar et al., 2007). Thus, silver carp is able to retain smaller particles more effectively than bighead carp. Consequently, silver carp is considered to be primarily a phytoplankton-feeder species (Smith, 1989; Vörös et al., 1997), while bighead carp is thought to be primarily zooplankton-feeder (Dong & Li, 1994; Kolar et al., 2007). However, recent investigations on Lake Balaton's (Hungary) bigheaded carp stock suggested that under certain environmental conditions the influence of gill raker morphology on the size distribution of the consumed food items may be less important than it was assumed earlier, presumably it is overwhelmed by the resource availability (Battonyai et al., 2015). Some studies also have argued that food selectivity of bigheaded carps may also be influenced by the mucus produced by the epibranchial organ, enabling fish to capture particles smaller than the mesh size of their gill rakers (Kolar et al., 2007 and references therein).

The relationship between the feeding habits of bigheaded carps and the characteristics of invaded habitat has been widely studied in the past, but the vast majority of these studies have focused only on how the feeding habit-related effects can induce alterations in the plankton

community (Fukushima et al., 1999; Domaizon & Dévaux, 1999; Lu et al., 2002) and change the trophic state of ambient water (Lin et al., 2014). Although the influence of several environmental factors (e.g., transparency, temperature) on the diet composition of bigheaded carps is supposable, our knowledge on the effects of habitat attributes (i.e., environmental factors) is limited.

According to the regulation of the Hungarian governmental authorities, bigheaded carps had been stocked into Lake Balaton until the early 1980s. However, these fish still form a massive stock in the lake and exhibit high individual growth rates and condition factor, despite the low planktonic productivity (Boros et al., 2014). For a better understanding of factors influencing feeding habits, gut contents of hybrid bigheaded carps from Lake Balaton were examined and their compositions were evaluated in this study to reveal the relationship between consumed food and available food resources, abiotic environmental factors and individual traits of fish. Among the various individual traits (body size, gender, gill raker morphology), we paid special attention to the gill raker morphology. Lake Balaton's bigheaded carp stock consists mainly of hybrid (bighead carp \times silver carp) individuals (Tátrai et al., 2009; Kovács et al., 2016), and gill rakers of these fish cover the entire morphological and functional range between the comb-like and sponge-like filtering apparatus types. The specific objectives of this study were: (i) to provide detailed data on the diet composition of introduced hybrid bigheaded carps living in a mesotrophic lake; and (ii) to assess how biotic and abiotic environmental factors and gill raker morphology affect diet composition in bigheaded carps.

Material and methods

Study area

Lake Balaton is the largest natural, shallow lake (surface area: 593 km²; mean depth: 3.2 m) in Central Europe, situated at 46° 42' - 47° 04' N, 17° 15' - 18° 10' E (Hungary) and 104.8 m above sea level. The lake is mesotrophic with mean annual chlorophyll-*a* concentrations of 3.6-18.7 mg m⁻³ (Istvánovics et al., 2007). Due to strong sediment resuspension, the lake is generally turbid with a Secchi depth varying between 0.2 m and 0.8 m (Specziár et al., 2013). Oxygen deficiency has never been recorded in the lake, and concentrations of pollutants are low or insignificant. Forty-seven percent of the lake shore is covered by native reed grass *Phragmites australis* (Cav.) Trin. ex Steud., while the remaining part of the lake shore was stabilized with stones and concrete. Submerged macrophytes occur sparsely in the littoral zone. The most abundant fishes in the lake are bleak *Alburnus alburnus* (Linnaeus, 1758), common bream *Abramis brama* (Linnaeus, 1758), razor fish *Pelecus cultratus* (Linnaeus, 1758) and the hybrid bigheaded carps. Detailed information on the limnology and fish fauna of the lake can be found in studies of Herodek et al. (1988), Istvánovics et al. (2007) and Specziár et al. (2009, 2013).

Assessment of abiotic environmental parameters and food resources

We measured a number of environmental variables that are believed to influence feeding efficiency and diet composition of bigheaded carps. At each field sampling occasion, we recorded water temperature (°C), conductivity (µS cm⁻¹) and Secchi depth (cm). We took water column samples with a tube sampler to determine total suspended matter concentration (TSM, mg L⁻¹), inorganic suspended matter concentration (IOSM, mg L⁻¹), chlorophyll-*a* concentration (µg L⁻¹), phytoplankton percentage taxonomic composition by biovolume, zooplankton total dry biomass (g L⁻¹) and percentage taxonomic composition by dry biomass. Samples and data were collected around fishing nets, at times when fish samplings were conducted.

TSM was assessed by filtrating lake water samples through 1.2 μ m Whatman GF/C glass fiber filters and filters were subsequently dried to constant weight at 60°C (ca. for 72 hours), whereas IOSM was estimated from the ash content of samples obtained by ignition at 550°C for 1 hour. Chlorophyll-*a* was extracted by acetone method (Aminot & Rey, 2000) and measured spectrophotometrically (Shimadzu UV-1601 spectrophotometer). Lake water subsamples for phytoplankton and zooplankton assemblage analysis were processed similarly as described in the gut content analysis (see below).

Gill raker morphology evaluation

Five morphotypes of gill rakers were assigned subjectively including comb-like type (type 1 gill raker; GR1), sponge-like type (type 5 gill raker; GR5) and intermediate (hybrid) structures (Fig. 1). The three intermediate gill raker morphotype classes represented transition between comb-like and sponge-like structures but to a different degree. GR3 category represented the completely intermediate type between the comb- and sponge-like structures, while hybrid gill raker areas closer to comb-like structure were classified into GR2 and those closer to sponge-like structure into GR4 (Fig. 1). The filtering apparatus of each bigheaded carp (N = 60) was then characterized based on the proportional area of GR1 to GR5 segments on the first left and right gill arches. Further, the filtering-to-respiratory part ratios (the relative width of gill raker to the width of gill filaments; RGRA) were also measured on the first left and right gill arches, because this is an important species-specific attribute in silver carps and bighead carps, and gill rakers of hybrids are generally intermediate in their development between the two species (Kolar et al., 2007).

Fish sampling and gut content analysis

170 Bigheaded carps were captured from the eastern basin of Lake Balaton by professional
171 fishermen (Balaton Fish Management Non-Profit Ltd.) using 12 cm knot-to-knot mesh-size
172 gillnets. Sampling was conducted in 2011 and 2013 at monthly intervals between March and
173 October, except July and August when fishing was banned in the lake by local regulations.

174 Diet composition was assessed using gut contents collected from the anterior segment of
175 the intestines, close to the pharynx. Although Vitál et al. (2015) recently argued that the
176 analysis of the filtrate samples collected from the inner surface of the gill-rakers would likely
177 provide more reliable picture of the food composition of bigheaded carps, we still decided to
178 use gut content samples because we wanted to compare our results explicitly with preceding
179 studies on bigheaded carp feeding, and the results of most of these studies are based on gut
180 content analyses. We found a useful amount of freshly ingested food in the gut of altogether
181 60 adult specimens, ranging between 78-118 cm in standard body length (SL) and 10.7-35.0
182 kg in body mass (M). Each gut content sample was divided into three identical portions for
183 phytoplankton, zooplankton and inorganic matter content analyses. Samples for
184 phytoplankton analyses were preserved in Lugol's solution and were stored at 4°C until
185 processing, while samples for zooplankton analyses were preserved in 70% ethanol. For
186 phytoplankton counting and identification, we used a Zeiss Axiovert-40 CFL inverted
187 microscope (400-fold magnification) and followed the method of Utermöhl (1958).
188 Biovolumes of algae were assessed using taxon-specific measurements and relationships
189 (Hillebrand et al., 1999). Identified phytoplankton organisms were classified as:
190 Cyanobacteria, Centrales, Pennales, Chlorococcales, Desmiales, Cryptophyta, Dinophyta,
191 Euglenophyta, Chrysophyceae, and Xanthophyceae. Zooplankton items and their fragments
192 were identified and counted under binocular microscope at 40-fold magnification and their
193 dry biomasses were assessed according to relevant length-mass relationships (Dumont et al.,
194 1975). Zooplankton organisms were classified into the following categories: *Dreissena* larvae,

rotifers, nauplius larvae of copepods, *Eudiapthomus* spp., *Cyclops* spp., harpacticoid copepods, *Bosmina* spp., *Daphnia* spp., *Diaphanosoma* spp., *Leptodora kindtii* and ostracods. Gut content subsamples for inorganic matter determination were measured for wet weight, then dried to constant weight at 60°C (ca. for 72 hours) and finally ignited at 550°C for 1 hour to assess their ash contents.

Stable isotope analysis

Samples for stable isotope analysis (SIA) were collected to complement the microscopic gut content analysis and reveal the relative contribution of different food resources (i.e., phytoplankton and zooplankton) in the diet of bigheaded carps. For this aim, seston (phytoplankton), zooplankton and fish muscle samples from each sampling month were analysed using a SERCON Integra 2 Stable Isotope Analyser to determine their stable nitrogen ($\delta^{15}\text{N}$) and carbon ($\delta^{13}\text{C}$) isotope signatures.

For phytoplankton SIA, water samples were first filtered through a 60 μm mesh-size plankton net to remove zooplankton (Baranyai & G.-Tóth, 2010; G.-Tóth et al., 2011) and then the filtrate samples were filtered onto pre-combusted GF/C glass fiber filters (1.2 μm pore-size). We assumed that the seston samples collected on the filters consisted mainly of algae. However, it must be noted that samples might contain a certain amount of resuspended sediment particles in addition to algae, because wind-driven turbulence in Lake Balaton results in a high concentration of resuspended inorganic sediment (rich in carbonates minerals) in the water column. Thus, evaluation of $\delta^{13}\text{C}$ measurements should be done with caution, because seston samples for SIA might contain some carbonate-derived inorganic carbon.

The suspensions (planktonic masses) retained by the 60 μm mesh-size net were collected separately and were treated as zooplankton samples. Moreover, dorsal muscle samples were

excised and collected from bigheaded carps. All samples were dried to a constant weight at 60 °C and were homogenised prior to SIA.

During the calculations, we focused on estimating the contribution of zooplankton as a food resource for bigheaded carps, as the microscopic analyses showed the apparent dominance of this food item in the gut contents. The following mixing model was used to assess the relative contribution of zooplankton to the nutrition of bigheaded carp:

$$\% \text{ contribution of zooplankton} = 100 \times (\delta^{15}\text{N}_{\text{BC}} - F - \delta^{15}\text{N}_{\text{S}}) / (\delta^{15}\text{N}_{\text{Z}} - \delta^{15}\text{N}_{\text{S}})$$

where the $\delta^{15}\text{N}_{\text{BC}}$ is δ N value of bigheaded carp, the $\delta^{15}\text{N}_{\text{S}}$ is δ value of seston and $\delta^{15}\text{N}_{\text{Z}}$ is δ of zooplankton, while F is the fractionation factor between successive trophic levels.

Preceding studies reported fractionation factor values of 3‰ to 5‰ (Peterson & Fry, 1987) for N between trophic levels. In our calculations, we used an average 4‰ enrichment value for N.

Due to the extremely wide range of $\delta^{13}\text{C}$ signatures (see Results section) in seston samples (most probably arising from the presence of inorganic carbonates), it was assumed that $\delta^{13}\text{C}$ values were not appropriate for the assessment of carbon fluxes and to trace dietary contributions. Hence, here we rely on $\delta^{15}\text{N}$ values to evaluate the contribution of potential food resource to the diet of bigheaded carps.

Data analysis

Diet composition and feeding strategy of bigheaded carps were inspected with the graphical method proposed by Costello (1990) and modified by Amundsen et al. (1996). In this analysis, the prey-specific percentage abundance (PSA) of each food component was plotted in relation to its percentage frequency of occurrence (FO) in all fish studied. The PSA of a prey taxon is defined as its percentage of all prey items in only those predators (i.e., bigheaded carps) in which the taxon occurs as prey. The product of PSA and the

corresponding FO value equals the mean percentage abundance of the given prey taxa in the diet of predators or their specified subset under study. Terms of individual and stock level specialization and generalization in respect to different prey taxa were used according to Amundsen et al. (1996) to describe the origin of the diet diversity.

We performed partial direct gradient analysis followed by a variance partitioning approach (Cushman & McGarigal, 2002; Peres-Neto et al., 2006) to evaluate the role of food resources (i.e., chlorophyll-a concentration, total zooplankton dry biomass, phyto- and zooplankton percentage taxonomic composition by biovolume and biomass, respectively), other environmental factors (i.e., water temperature, conductivity, Secchi depth, TSM, IOSM), seasonality (i.e., sampling months) and individual features of fish (i.e., gender, SL, RGRA and percentage area of different GR morphotypes) in gut content variability of hybrid bigheaded carps. Because of their extremely low representation in the gut content (<0.05% in abundance for all fish), four algae groups – Chrysophyceae, Cryptophyta, Desmidiaceae and Xanthophyceae – were excluded from response variables to reduce their disproportionate effect in multivariate analyses (Legendre & Legendre, 2012). For analyses, percentage gut content data were $\arcsin\sqrt{x}$ transformed to improve their normality. Of the potential explanatory variables, phyto- and zooplankton percentage taxonomic composition and percentage GR morphotype data were $\arcsin\sqrt{x}$ transformed, and temperature, conductivity, Secchi depth, TSM, IOSM, SL, chlorophyll-a concentration, total zooplankton dry biomass and RGRA were $\ln x$ transformed prior to analysis. Month of sampling and gender were recoded into binary dummy variables. A detrended correspondence analysis (DCA) with down weighting rare taxa indicated relatively long gradient length (3.27 in standard deviation units) in our data, therefore we chose canonical correspondence analysis (CCA) for further analysis (Lepš & Šmilauer, 2003). Potential explanatory variables were filtered for collinearity at $r > 0.7$ (i.e., Pearson correlation analysis in Statistica 8.0 software package; www.statsoft.com)

and subjected to a forward stepwise selection procedure (at $P < 0.05$) based on Monte Carlo randomization test with 9,999 unrestricted permutations under the full model. This selection resulted in six effective explanatory variables for the final overall CCA model. Beside these variables, to improve readability of the graphical output of the analysis, we included some passive supplementary factors during this final analysis as well, such as males, total zooplankton abundance, chlorophyll-a concentration, GR1 and GR5. Supplementary factors were not used during the construction of the model, but based on the ordination results, their positions can be projected into the ordination space, and their meaning can be interpreted. Then, a series of CCA and partial CCAs were conducted to partition the effects of significant explanatory variables on gut content of hybrid bigheaded carps (Cushman & McGarigal, 2002). DCA and CCA analyses were performed using CANOCO version 4.5 software (ter Braak & Šmilauer, 2002).

Results

Abiotic environmental parameters and food resource

Measured values of abiotic environmental parameters such as water temperature, conductivity, Secchi depth, TSM and IOSM are summarized in the Table 1.

Chlorophyll-a concentration ranged between 1.5 and 7.3 $\mu\text{g L}^{-1}$ (Fig. 2a) indicating quite low total phytoplankton biomass. The phytoplankton assemblage was dominated by pennate and centric diatoms (Fig. 2a). Other abundant algae were Chlorococcales and larger abundance of Dinophyta was occasionally observed.

The total zooplankton biomass varied considerably between sampling dates and ranged from 0.107 to 0.659 mg L^{-1} (Fig. 2b). In most sampling dates, copepods (i.e., *Eudiaptomus gracilis* and *Cyclops* spp.) represented the highest bulk of the zooplankton biomass, except in

the warmest months, when *Diaphanosoma mongolianum*, a large-bodied cladoceran predominated.

Gill raker variability

In Lake Balaton, almost all bigheaded carps have gill rakers intermediate in morphology between those typical for bighead carp (GR1) and silver carp (GR5), but the proportion of different morphotype segments in the gill rakers varied considerably among individuals (Fig. 3). The most abundant gill raker morphotype segment was the GR3 (i.e., the intermediate type between bighead carp and silver carp), represented in 93% of the bigheaded carps with a median area of 58%. GR1 and GR5 morphotype segments occurred in 5% and 37% of individuals, respectively, and GR5 generally occurred only in minor proportions (median: 15%) of the total area of gill rakers. In most bigheaded carps, segments of two or three gill raker morphotypes occurred on the same gill arch, indicating complex gill raker morphology and various filtering capacity (i.e., complex food-size selectivity) even at the individual level.

Diet composition and feeding strategy

Zooplankton dominated over phytoplankton in the ingested food of bigheaded carps; the mean proportion of zooplankton ranged from 12.4% to 74.6%, whereas phytoplankton amounted between 0.0% and 42.8% (Fig. 4). Among zooplankters, rotifers, *Cyclops* spp. and *Bosmina* spp. were consumed in the largest quantity. However, beside planktonic crustaceans the occurrence of ostracods and some harpacticoid copepods in the gut contents indicated occasional role of benthic food resources in the diet of bigheaded carps as well. The most abundant phytoplankton taxa in the gut contents were diatoms, mainly taxa of Pennales. Moreover, 25.4-56.6% of the monthly mean gut content samples, proved to be inorganic matter, suggesting a significant amount of ballast feeding.

Graphical analysis of feeding strategy revealed that the most consistent component of the gut contents was the inorganic matter, which comprised a substantial proportion of the filtered matter in all bigheaded carps (Fig. 5). Both the frequency of occurrence and the prey-specific abundance of zooplankton taxa varied among seasons, but also indicated some taxon-specific tendencies. Namely, the consumption of rotifers was generally frequent at a moderate to low significance. On the other hand, *Bosmina* cladocerans, predacious cladoceran *Leptodora kindtii*, and harpacticoid copepods and ostracods were less frequently preyed by bigheaded carps but sometimes at substantial individual specialization. Diatoms were stable components of the gut contents in some periods but their prey-specific abundance never exceeded 24.2% and 8.6% regarding Pennales and Centrales taxa, respectively. Chlorococcales, cyanobacteria, Dynophyta and Euglenophyta occurred also frequently in the gut content but their prey-specific abundances were negligible.

Influence of environment, season and individual traits on the diet composition

The CCA model explained 26.5% of the variance in the diet of bigheaded carps and indicated statistically significant, but moderate roles of IOSM, food resource (i.e. relative abundance by biomass of *Cyclops* spp. and Chlorococcales), gender, water temperature and individual variability in morphology of the filtering apparatus (Table 2; Fig. 6). Variance partitioning revealed that influences of the six explanatory variables retained for the final CCA model were mostly independent (i.e., majority of their explanatory power came from pure effects).

The first CCA axis accounted for the 12.7% of the variance in the diet data and positively correlated with *Cyclops* spp. relative abundance in the lake and negatively correlated with IOSM and water temperature (Fig. 6). All algae taxa found in the gut content received negative scores along this axis, while zooplankton taxa dispersed more in the ordination range and generally positioned in the positive range. The second CCA axis captured 5.8% of the

total variance and positively correlated with females and negatively correlated with Chlorococcales abundance in the lake water and the percentage area of silver carp-like (GR4) hybrid gill raker morphotype. Inorganic matter content of the gut content seemed to be highly independent from the considered explanatory variables and positioned in the centre of the ordination space. Although with no identified significant effect in the CCA model, positioning of the supplementary variables indicated some positive tendencies between the consumption of algae and the chlorophyll-a concentration (i.e., total phytoplankton density) in the lake and the proportion of silver carp gill raker (GR5) morphotype, as well as between the consumption of zooplankton and the proportion of bighead carp gill raker (GR1) morphotype.

Stable isotope analysis

The average (\pm SD) $\delta^{13}\text{C}$ value of seston was $-12.7 (\pm 4.6)\text{‰}$, while the average $\delta^{13}\text{C}$ values of zooplankton and bigheaded carp muscle samples were substantially lower than of seston, with average values of $-26.4 (\pm 1.34)\text{‰}$ and $-25.5 (\pm 0.6)\text{‰}$, respectively (Fig. 7). The difference between average $\delta^{13}\text{C}$ value of zooplankton and bigheaded carps (1.1‰) were close to the value of enrichment between successive trophic levels, suggesting direct carbon flow from zooplankton to bigheaded carps. The average (\pm SD) $\delta^{15}\text{N}$ values of seston, zooplankton and bigheaded carps were $2.3 (\pm 1.4)\text{‰}$, $4.7 (\pm 1.5)\text{‰}$ and $8.9 (\pm 0.4)\text{‰}$, respectively (Fig. 7). The differences between average $\delta^{15}\text{N}$ values of the three sample type indicated predator-prey interactions between bigheaded carps and zooplankton, and did not show direct trophic interaction between seston and bigheaded carps. Calculations based on the mixed model revealed that the average contribution of zooplankton to the diet of bigheaded carps was $104.2 (\pm 22.4)\%$.

Discussion

In Lake Balaton's hybrid-dominated bigheaded carp stock we found various types of gill rakers, including comb-like (typical for bighead carp and effective in filtering zooplankton), sponge-like (typical for silver carp and effective in harvesting phytoplankton) structures, and various types of complex, intermediate (hybrid) gill rakers. It is likely that major differences in gill raker morphology among individuals may be accompanied by a high inter-individual variability in food composition and trophic role (Spataru et al., 1983; Jayasinghe et al., 2015), because mesh size of the rakers determines the smallest size of consumable food items (Dong & Li, 1994; Vörös et al., 1997). Nevertheless, food composition of individuals with various types of gill rakers was quite similar, and zooplankton predominated in the gut contents of all examined bigheaded carps. It turned out that gill raker morphology exerts only minor influence on the consumed food, but the effects of environmental factors (inorganic suspended matter, resource availability) have a decisive role in shaping the food composition of bigheaded carps in Lake Balaton (see also Battonyai et al., 2015).

Most bigheaded carps possess hybrid-type (intermediate in development between comb-like and sponge-like structures) filtering apparatus in Lake Balaton. The different gill raker morphotypes are often represented on the same gill arch of a single individual. Sponge-like structures (i.e., GR5) were observed on the gill arches of most individuals, providing a theoretical chance for the vast majority of the stock to capture phytoplankton effectively. However, it appears that phytoplankton has only a negligible contribution to the diet of bigheaded carps in Lake Balaton, which could be explained at least in part by the low biomass of algae in Lake Balaton (monitored via the chlorophyll-a concentration in the water column). Previous laboratory experiments on the feeding habits of silver carp revealed that the intensity of grazing on phytoplankton was primarily determined by the density of algae in the ambient water (Herodek et al., 1989). Thus, in oligotrophic and mesotrophic habitats (such as Lake Balaton), the relative importance of phytoplankton in the food is supposed to be lower

394 compared to that in highly productive, eutrophic or hypertrophic ecosystems. However, our
395 knowledge on the importance of phytoplankton consumption by silver carp and more
396 generally by bigheaded carps in habitats of low productivity is limited. Feeding habits of
397 bigheaded carps were mainly studied in eutrophic water bodies and in fertilized aquaculture
398 operations, particularly in relation with their use as tools of biomanipulation (Zhang et al.,
399 2008; Xie & Liu, 2001) and to boost aquaculture yields by implementing polyculture
400 technologies based on a more direct utilization of primary production (Kolar et al., 2007).

401 The wide range and extremely high $\delta^{13}\text{C}$ values of seston did not facilitate the traditional,
402 stable carbon-isotope-based evaluation of dietary interactions. Such high $\delta^{13}\text{C}$ values (e.g., –
403 10 to –15) for phytoplankton have been reported from shallow, eutrophic lakes (Gu &
404 Schelske, 1996). Thus, in productive ecosystems the phytoplankton can be enriched in ^{13}C
405 due to the high assimilation rate for ^{13}C -rich dissolved inorganic carbon (Gu & Schelske,
406 1996). However, in the case of Lake Balaton the observed high $\delta^{13}\text{C}$ values very likely were
407 consequence of the presence of carbonate- derived inorganic carbon in seston samples, i.e.,
408 the $\delta^{13}\text{C}$ values did not represent reliably the phytoplankton. However, using $\delta^{15}\text{N}$ signatures,
409 the complementary stable isotope analysis supported the findings of gut content analysis and
410 suggested that ingested phytoplankton does not contribute substantially to the nutrition of
411 bigheaded carps in Lake Balaton.

412 Beside the low abundance of algae in the water, the low contribution of phytoplankton in
413 the nutrition of bigheaded carps might also be a consequence of restrained capability of
414 bigheaded carps to digest and utilize most algae found in the lake. Görgényi et al. (2016)
415 studied the consumption and digestion of algae by bigheaded carps in Lake Balaton and
416 showed that cells or colonies of several phytoplankton species can survive the passage
417 through the alimentary canal and can be found in viable form in the hindguts (i.e., in the
418 faeces). Bitterlich (1985) and Gerking (1994) explained the low efficiency of bigheaded carps

in utilizing some phytoplankton taxa with the lack of cellulase enzyme in the gut fluids and the relatively high pH in their digestive tract. Accordingly, the stable isotope analyses in this study confirmed that the vast majority of the metabolised nutrients was zooplankton-derived in bigheaded carps.

In the light of these findings, it seems that the fundamental benefit provided by the diverged gill raker morphologies of bigheaded carps (i.e., comb-like and sponge-like rakers) to escape interspecific diet overlap can diminish under certain environmental conditions, for instance in phytoplankton-poor environments like Lake Balaton is. According to Ke et al. (2008), this phenomenon is the matter of abundance and quality of available food resources. The same authors found that both silver and bighead carp showed preference for zooplankton and shared this higher quality food resource when it was abundant. However, substantial diet overlap was also observed when alternative food resources were depleted (Ke et al., 2008; Chen et al., 2011). Thus, in the mesotrophic Lake Balaton where the availability of digestible phytoplankton is low, bigheaded carps have no alternatives to feeding on zooplankton regardless of the differences of their individual gill raker morphology and the likewise moderate zooplankton abundance (G.-Tóth et al., 2011). On the other hand, under eutrophic conditions, which is actually the typical environment of bigheaded carps (Kolar et al., 2007), food resources are more abundant and diverse, and therefore, species-specific differences in the filtration capacity (i.e., utilizable food size spectra) may be more important in grazing on the most profitable food resource and avoiding interspecific competition.

Recent hydroacoustic surveys revealed that non-native bigheaded carps constitute about 20 – 30 % of the total fish biomass in the Lake Balaton (Tátrai et al., 2009; Boros, 2015). Because this massive bigheaded carp stock primarily feeds on zooplankton, it certainly can thereby exert a considerable ecological effect on the whole ecosystem of Lake Balaton (e.g., through top-down control). Through their intense grazing, bigheaded carps can alter the

abundance and assemblage composition of zooplankton and consequentially indirectly the phytoplankton community as well (Lu et al., 2002; Zhou et al., 2011). Mass consumption of zooplankton especially limits the availability of food for several native species (e.g., bleak, razor fish and common bream) and the earliest, zooplanktivorous life stages of almost all fish species (Specziár & Rezsú, 2009). Moreover, bigheaded carps may affect the whole nutrient cycle of the lake through trophic cascades (Lieberman, 1996). For instance, by altering the structure of plankton community, these fish species affect the utilization patterns and the turnover time of nutrients in water column (Domaizon & Devaux, 1999; Mátyás et al., 2003).

The unintentional ingestion of inorganic particles is common in the case of filter-feeding animals, like herbivorous zooplankton (e.g., G.-Tóth et al., 1986; Rellstab & Spaak, 2007); however, this phenomenon is still poorly documented in filter-feeding fishes. The sampled hybrid bigheaded carps exhibited an unusually high level of inorganic debris consumption. It is also surprising that the proportion of the ingested inorganic debris in the gut content was highly independent of the environmental circumstances (considering both abiotic and food resource-related variables), and it did not correlate with gill raker morphology. Because neither the bighead carp-type, nor the silver carp-type gill rakers proved to be effective in avoiding inorganic debris consumption, it is likely that the entrapment of these small-sized suspended inorganic particles was facilitated by the mucus coating on the gill rakers (Sanderson, 1996; Gophen, 2014). The hydro-morphological attributes of Lake Balaton probably contributed to the high inorganic debris consumption of bigheaded carps. This large but relatively shallow lake is highly exposed to wind-generated turbulence, resulting in very high suspended sediment concentrations. Nevertheless, the statistical analysis failed to find any association between the amount of inorganic suspended material (IOSM) in the lake water and proportional contribution of inorganic matter in the gut. Normally fishes tend to avoid the ingestion of inorganic particles. However, because bigheaded carps seem not to be able to

469 separate food from inorganic particles during their filtering activity, they presumably decrease
470 the rate of filtration under unfavourable turbid conditions to set a limit for inorganic debris
471 feeding. In contrast to other filter feeders such as herbivorous zooplankters (Rellstab & Spaak,
472 2007), bigheaded carps' inorganic debris consumption did not result in poor somatic
473 conditions in Lake Balaton. The observed condition factor (expresses the "plumpness" of fish)
474 and growth rates of fish were high in Lake Balaton compared to other bigheaded carp
475 populations (Boros et al., 2014). Solving the contradiction between the seemingly poor food
476 resources and the ideal condition factor and growth rate of bigheaded carps remains the task
477 of future researches.

478 Abiotic environmental factors, namely IOSM and water temperature accounted for a
479 remarkable proportion of the explained variance in the diet of bigheaded carps. In our
480 opinion, the effect of these factors is mainly indirect and they may designate certain patterns
481 of food availability. Specifically, IOSM indicates the strength of wind-generated turbulence of
482 the water which can alter the distribution, abundance and assemblage structure of planktonic
483 organisms (Baranyai & G.-Tóth, 2010; Baranyai et al., 2011). For instance, some zooplankton
484 taxa are sensitive to turbulence and respond to windy weather with decreased abundance
485 either due to increased mortality (O'Brien et al., 2004) or because of they migrate into the
486 benthic zone (Baranyai & G.-Tóth, 2010). On the other hand, turbulence can displace benthic
487 taxa into the water column (Goździewska et al., 2006) creating a food resource for
488 planktivores such as bigheaded carps. The occasional occurrence of benthic crustaceans (e.g.,
489 *Ostracoda* and *Harpacticoida* species) in the gut contents of bigheaded carps supports this
490 hypothesis. Moreover, because fishes are able to effectively discern the inorganic
491 contaminants in food (e.g., Callan & Sanderson, 2002; Finger, 2008), the high amount of
492 suspended inorganic material may alter taste-sensitive foraging performance of bigheaded
493 carps, resulting in decreased feeding intensity and altered food selection. Water temperature

correlated strongly with the binary dummy variables describing seasonality of samples (i.e., name of months) and also represented seasonal patterns of food resources. Therefore, considering all the direct and indirect effects signified by the explanatory factors retained for the final CCA model to describe the background of dietary variability of bigheaded carps in Lake Balaton, we can submit that variability of the food resource availability and water turbidity are the factors which influence the feeding of these fish principally, whereas individual gill raker morphology has little importance in this particular environment.

Conclusions

Our findings emphasize the importance of environmental factors (especially turbidity and seasonally dissimilar availability of food resources) in shaping the feeding habits of bigheaded carps. However, results of this study also reveal some uncommon patterns in the feeding behavior of these filter-feeding fishes in this particular mesotrophic environment, where the ingested and especially the utilized food was mainly zooplankton, irrespective of the gill raker morphology and hybrid status of individuals. This study contradicts the general assumption that silver carp with its sponge-like gill raker consume mainly phytoplankton and shows that in certain habitats the nutritional role of this food resource could be negligible. Consequently, bigheaded carps can be considered direct food competitors of the zooplanktivorous native fishes and early, zooplanktivorous life stages of nearly all fish species in Lake Balaton. Further research is needed to explore what circumstances can cause the diminishing dietary and functional differences among fish species with markedly different filtering organs and to investigate the impact of invasive bigheaded carps on the native biota and nutrient cycling in low productivity ecosystems.

Acknowledgements

This study was supported by GINOP – 2.3.2-15-2016-00004 and OTKA K-83893 projects.

We thank I. Mecs-nóbel and Z. Poller for their skilful assistance in the laboratory. We also acknowledge the useful comments of J. Syväranta and the helpful contribution of L. Antal to the figure preparation. Bigheaded carps for this study were provided by the Balaton Fish Management Non-Profit Ltd.

References

- Aminot, A. & F. Rey, 2000. Standard procedure for the determination of chlorophyll-a by spectroscopic methods. International Council for the Exploration of Sea, Denmark: 1–11.
- Amundsen, P. A., H. M. Gabler, & F. J. Staldivik, 1996. A new approach to graphical analysis of feeding strategy from stomach contents data: modification of the Costello (1990) method. *Journal of Fish Biology* 48: 607–614.
- Baranyai, E. & L. G.-Tóth, 2010. The influence of turbulence on vertical distribution of zooplankton in shallow, kinetic Lake Balaton (Hungary). *Verhandlungen der Internationale Vereinigung für Theoretische und Angewandte Limnologie* 30: 1505–1508.
- Baranyai, E., L. G.-Tóth, Á. Vári, & Z.G. Homonnay, 2011. The effect of variable turbulent intensities on the distribution of zooplankton in the shallow, large Lake Balaton (Hungary). *Knowledge and Management of Aquatic Ecosystems* 400: article No. 07.
- Battonyai, I., A. Specziár, Z. Vitál, A. Mozsár, J. Görgényi, G. Borics, L. G.-Tóth, & G. Boros, 2015. Relationship between gill raker morphology and feeding habits of hybrid bigheaded carps (*Hypophthalmichthys* spp.). *Knowledge and Management in Aquatic Ecosystems* 416: article No. 36.
- Bitterlich, G., 1985. Digestive enzyme pattern of two stomachless filter feeders, silver carp, *Hypophthalmichthys molitrix* Val., and bighead carp, *Aristichthys nobilis* Rich. *Journal of Fish Biology* 27: 103–112.

544 Borics, G., I. Grigorszky, S. Szabó, & J., Padisák, 2000. Phytoplankton associations under
 545 changing pattern of bottom-up vs. top-down control in a small hypertrophic fishpond in
 546 East Hungary. *Hydrobiologia* 424: 79–90.

547 Boros, G., A. Mozsár, Z. Vitál, S. A. Nagy, & A. Specziár, 2014. Growth and condition factor
 548 of hybrid (Bighead *Hypophthalmichthys nobilis* Richardson, 1845 × silver carp *H. molitrix*
 549 Valenciennes, 1844) Asian carps in the shallow, oligo-mesotrophic Lake Balaton. *Journal*
 550 *of Applied Ichthyology* 30: 546–548.

551 Boros, G., 2015: A Balatoni busaállomány kutatásának előzményei és fontosabb eredményei a
 552 2010 – 2015 közötti időszakban. Scientific report of Balaton Limnological Institute (CER,
 553 HAS) for The Ministry of Agriculture (in Hungarian).

554 ter Braak, C. J. F. & P. Šmilauer, 2002. CANOCO reference manual and CanoDraw for
 555 Windows user's guide: software for canonical community ordination (version 4.5).
 556 Microcomputer Power, Ithaca, New York.

557 Britton, J. R., G. D., Davies & C. Harrod, 2010. Trophic interactions and consequent impacts
 558 of the invasive fish *Pseudorasbora parva* in a native aquatic foodweb: a field investigation
 559 in the UK. *Biological Invasions* 12: 1533–1542.

560 Calkins, H. A., S. J. Tripp & J. E. Garvey, 2012. Linking silver carp habitat selection to flow
 561 and phytoplankton in the Mississippi River. *Biological Invasions* 14: 949–958.

562 Callan, W. T. & S. L. Sanderson, 2002. Feeding mechanisms in carp: cross filtration, palatal
 563 protrusion and flow reversals. *The Journal of Experimental Biology* 206: 883–892.

564 Chen, G., Z. Wu, B. Gu, D. Liu, X. Li & Y. Wang, 2011. Isotopic niche overlap of two
 565 planktivorous fish in southern China. *Limnology* 12: 151–155.

566 Cooke, S. L., W. R. Hill & K. P. Meyer, 2009. Feeding at different plankton densities alters
 567 invasive bighead carp (*Hypophthalmichthys nobilis*) growth and zooplankton species
 568 composition. *Hydrobiologia* 625: 185–193.

569 Costello, M. J., 1990. Predator feeding strategy and prey importance: a new graphical
570 analysis. *Journal of Fish Biology* 36: 261–263.

571 Cremer, M. C. & R. O. Smitherman, 1980. Food habits and growth of silver and bighead carp
572 in cages and ponds. *Aquaculture* 20: 57–64.

573 Cushman, S. A. & K. McGarigal, 2002. Hierarchical, multi-scale decomposition of species-
574 environment relationships. *Landscape Ecology* 17: 637–646.

575 Domaizon, I. & J. Dévaux, 1999. Impact of moderate silver carp biomass gradient on
576 zooplankton communities in a eutrophic reservoir. Consequences for the use of silver carp
577 in biomanipulation. *Comptes Rendus de l'Académie des Sciences Paris, Life Sciences* 322:
578 621–628.

579 Dong, S. & D. Li, 1994. Comparative studies on the feeding selectivity of silver carp
580 *Hypophthalmichthys molitrix* and bighead carp *Aristichthys nobilis*. *Journal of Fish*
581 *Biology* 44: 621–626.

582 Dumont H. J., I. Van de Velde & S. Dumont, 1975. The dry weight estimate of biomass in a
583 selection of Cladocera, Copepoda and Rotifera from the plankton, periphyton and benthos
584 of continental waters. *Oecologia* 19: 75–97.

585 Finger T. E., 2008. Sorting food from stones: the vagal taste system in Goldfish, *Carassius*
586 *auratus* *Journal of Comparative Physiology A* 194: 135–143.

587 Fukushima, M., N. Takamura, L. Sun, M. Nakagawa, K. Matsushige & P. Xie, 1999. Changes
588 in the plankton community following introduction of filter-feeding planktivorous fish.
589 *Freshwater Biology* 42: 719–735.

590 G.-Tóth, L., K. V.-Balogh & N. Zánkai, 1986. Significance and degree of abioseston
591 consumption in the filter-feeder *Daphnia galeata* Sars am. Richard (Cladocera) in Lake
592 Balaton. *Archiv für Hydrobiologie* 106: 45–60.

593 G.-Tóth, L., L. Parpala, Cs. Balogh, I. Tátrai & E. Baranyai, 2011. Zooplankton community
 594 response to enhanced turbulence generated by water-level decrease in Lake Balaton, the
 595 largest shallow lake in Central Europe. *Limnology and Oceanography* 56: 2211–2222.

596 Gerking, S. D., 1994. Feeding ecology of fish. Academic Press, San Diego.

597 Gophen, M., 2014. Competitive consumption of the Lake Kinneret (Israel) plankton by
 598 *Hypophthalmichthys molitrix* and *Sarotherodon galilaeus*. *Open Journal of Ecology* 4:
 599 532–542

600 Goździewska, A., E. Paturej & A. Krajewska-Sołtys, 2006. Effect of environmental factors
 601 on the occurrence of the Harpacticoida in the zooplankton and in the diet of smelt *Osmerus*
 602 *eperlanus* (L.) in the Vistula Lagoon. *Electronic Journal of Polish Agricultural Universities*
 603 9: article No. 16.

604 Görgényi, J., G. Boros, Z. Vitál, A. Mozsár, G. Várbiro, G. Vasas & G. Borics, 2016. The role
 605 of filter feeding Asian carps in algal dispersion. *Hydrobiologia* 764: 115–126.

606 Gu, B. & C. L. Schelske, 1996. Temporal and spatial variations in phytoplankton carbon
 607 isotopes in a polymictic subtropical lake. *Journal of Plankton Research* 18: 2081–2092.

608 Hampl, A., J. Jirásek & D. Sirotek, 1983. Growth morphology of the filtering apparatus of
 609 silver carp (*Hypophthalmichthys molitrix* Val.) - II. microscopic anatomy. *Aquaculture* 31:
 610 153–158.

611 Hayer, C.-A., J. J. Breuggemann, R. A. Klumb, B. D. S. Graeb & K. N. Bertrand, 2014a.
 612 Population characteristics of bighead and silver carp on the northwestern front of their
 613 North American invasion. *Aquatic Invasions* 9: 289–303.

614 Hayer, C.-A., B. D. S. Graeb & K. N. Bertrand, 2014b. Adult, juvenile and young-of-year
 615 bighead, *Hypophthalmichthys nobilis* (Richardson, 1845) and silver carp, *H. molitrix*
 616 (Valenciennes, 1844) range expansion on the northwestern front of the invasion in North
 617 America. *BioInvasions Records* 3: 283–289.

618 Herodek, S., L. Laczkó & Á. Virág, 1988. Lake Balaton: Research and Management. Nexus,
 619 Budapest.
 620 Herodek, S., I. Tátrai, J. Oláh & L. Vörös, 1989. Feeding experiments with silver carp
 621 (*Hypophthalmichthys molitrix* Val.) fry. Aquaculture 83: 331–344.
 622 Hillebrand, H., C. D. Dürselen, D. Kirchtel, U. Pollinger & T. Zohary, 1999. Biovolume
 623 calculation for pelagic and benthic microalgae. Journal of Phycology 35: 403–424.
 624 Irons, K.S., G.G. Sass, M.A. McClelland & M.A. Stafford, 2007. Reduced condition factor of
 625 two native fish species coincident with invasion of nonnative Asian carps in the Illinois
 626 River, U.S.A. Is this evidence for competition and reduced fitness? Journal of Fish Biology
 627 71: 258–273.
 628 Istvánovics, V., A. Clement, L. Somlyódy, A. Specziár, L. G.-Tóth & J. Padisák, 2007.
 629 Updating water quality targets for shallow Lake Balaton (Hungary), recovering from
 630 eutrophication. Hydrobiologia 581: 305–318.
 631 Jayasinghe, U. A. D., E. García-Berthou, Z. Li, W. Li, T. Zhang & J. Liu, 2015. Co-occurring
 632 bighead and silver carps show similar food preference but different isotopic niche overlap
 633 in different lakes. Environmental Biology of Fishes 98: 1185–1199.
 634 Ke, Z., P. Xie & L. Guo, 2008. In situ study on effect of food competition on diet shifts and
 635 growth of silver and bighead carps in large biomanipulation fish pens in Meiliang Bay,
 636 Lake Taihu. Journal of Applied Ichthyology 24: 263–268.
 637 Khan, M. F. & P. Panikkar, 2009. Assessment of impacts of invasive fishes on the food web
 638 structure and ecosystem properties of a tropical reservoir in India. Ecological Modelling
 639 220: 2281–2290.
 640 van Kleef, H., G. van der Velde, R. S. E. W. Leuven & H. Esselink, 2008. Pumpkinseed
 641 sunfish (*Lepomis gibbosus*) invasions facilitated by introductions and nature management

642 strongly reduce macroinvertebrate abundance in isolated water bodies. Biological
643 Invasions 10: 1481–1490.

644 Kolar, C. S., D. C. Chapman, W. R. Jr Courtenay, C. M. Housel, J. D. Williams & D. P.
645 Jennings, 2007. Bigheaded carps – a biological synopsis and environmental risk
646 assessment. Bethesda.

647 Kovács, B., G. Boros, Z. Vítál, A. Mozsár, A. Specziár, V. Józsa, B. Urbányi & I. Lehoczky,
648 2016. Genetic analysis of filter-feeding Asian carps (*Hypophthalmichthys* spp.) in Lake
649 Balaton, Hungary. Aquaculture Europe 2016 – Conference of the European Aquaculture
650 Society, 20–23 September, Edinburgh, Scotland. (Abstract 251;
651 <https://www.was.org/EasOnline/AbstractDetail.aspx?i=6754>)

652 Legendre, P. & L. Legendre, 2012. Numerical Ecology. Elsevier, Amsterdam.

653 Lepš, J. & P. Šmilauer, 2003. Multivariate analysis of ecological data using CANOCO.
654 Cambridge University Press, New York.

655 Lieberman, D. M., 1996. Use of silver carp (*Hypophthalmichthys molitrix*) and bighead carp
656 (*Aristichthys nobilis*) for algae control in a small pond: Changes in water quality. Journal
657 of Freshwater Ecology 11: 391–397.

658 Lin, Q., X. Jiang, B. P. Han & E. Jeppesen, 2014. Does stocking of filter-feeding fish for
659 production have a cascading effect on zooplankton and ecological state? A study of
660 fourteen (sub)tropical Chinese reservoirs with contrasting nutrient concentrations.
661 Hydrobiologia 736: 115–125.

662 Lu, M., P. Xie, H. Tang, Z. Shao & L. Xie, 2002. Experimental study of trophic cascade
663 effect of silver carp (*Hypophthalmichthys molitrix*) in a subtropical lake, Lake Donghu: On
664 plankton community and underlying mechanisms of changes of crustacean community.
665 Hydrobiologia 487: 19–31.

666 Mátyás, K., I. Oldal, J. Korponai, I. Tátrai & G. Palulovits, 2003. Indirect effect of different
 667 communities on nutrient chlorophyll relationship in shallow hypertrophic water quality
 668 reservoirs. *Hydrobiologia* 504: 231–239.

669 O'Brien, K. R., C. L. Meyer, A. M. Waite, G. N. Ivey & D. P. Hamilton, 2004.
 670 Disaggregation of *Microcystis aeruginosa* colonies under turbulent mixing: laboratory
 671 experiments in a drid-stirred tank. *Hydrobiologia* 519: 143–152.

672 Peres-Neto, P. R., P. Legendre, S. Dray & D. Borcard, 2006. Variation partitioning of species
 673 data matrices: estimation and comparison of fractions. *Ecology* 87: 2614–2625.

674 Rellstab, C. & P. Spaak, 2007. Starving with a full gut? Effect of suspended particles on the
 675 fitness of *Daphnia hyaline*. *Hydrobiologia* 594: 131–139.

676 Peterson, B. J. & B. Fry, 1987. Stable isotopes in ecosystem studies. *Annual Review of*
 677 *Ecology and Systematics*. 18: 293–320.

678 Sanderson, S. L., M. C. Stebar, K. L. Ackerman, S. H. Jones, I. E. Batjakas & L. Kaufman,
 679 1996. Mucus entrapment of particles by a suspension-feeding tilapia (Pisces: Cichlidae).
 680 *The Journal of Experimental Biology* 199: 1746–1756.

681 Sass, G. G., C. Hinz, A. C. Erickson, N. N. McClelland, M. A. McClelland & J. M. Epifanio,
 682 2014. Invasive bighead and silver carp effects on zooplankton communities in the Illinois
 683 River, Illinois, USA. *Journal of Great Lakes Research* 40: 911–921.

684 Sampson, S. J., J. H. Chick & M. A. Pegg, 2009. Diet overlap among two Asian carp and
 685 three native fishes in backwater lakes on the Illinois and Mississippi rivers. *Biological*
 686 *Invasions* 11: 483–496.

687 Solomon, L. E., R. M. Pendleton, J. H. Chick & A. F. Casper, 2016. Long term changes in
 688 fish community structure in relation to the establishment of Asian carps in a large
 689 floodplain river. *Biological Invasions*: online first.

690 Spataru, P, G. W. Wohlfarth & G. Hulata, 1983. Studies on the natural food of different fish
 691 species in intensively manured polyculture ponds. *Aquaculture* 35: 283–298.

692 Specziár, A. & E. T. Rezsü, 2009. Feeding guilds and food resource partitioning in a lake fish
 693 assemblage: an ontogenetic approach. *Journal of Fish Biology* 75: 247–267.

694 Specziár, A., T. Erős, Á. I. György, I. Tátrai & P. Bíró, 2009. A comparison between the
 695 Nordic gillnet and whole water column gillnet for characterizing fish assemblages in the
 696 shallow Lake Balaton. *Annales de Limnologie – International Journal of Limnology* 45:
 697 171–180.

698 Specziár, A., Á. I. György & T. Erős, 2013. Within-lake distribution patterns of fish
 699 assemblages: the relative role of spatial, temporal and random environmental factors in
 700 assessing fish assemblages using gillnets in a large and shallow temperate lake. *Journal of*
 701 *Fish Biology* 82: 840–855.

702 Smith, D. W., 1989. The feeding selectivity of silver carp, *Hypophthalmichthys molitrix* Val.
 703 *Journal of Fish Biology* 34: 819–828.

704 Tátrai, I., G. Paulovits, V. Józsa, G. Boros, Á.I. György & J. Héri, 2009. Halállományok
 705 eloszlása és a betelepített halfajok állománya a Balatonban In: Bíró, P. & J. Banczerowski
 706 (eds), *A Balatonkutatások fontosabb eredményei 1999–2009*. MTA, Budapest: 129–141
 707 (in Hungarian).

708 Utermöhl, H., 1958. Zur Vervollkommnung der quantitativen Phytoplankton-Methodik.
 709 *Mitteilungen der Internationale Vereinigung für theoretische und angewandte Limnologie*
 710 9: 1–38.

711 Vitál, Z., A. Specziár, A. Mozsár, P. Takács, G. Borics, J. Görgényi, L. G.-Tóth, S. A. Nagy
 712 & G. Boros, 2015. Applicability of gill raker filtrates and foregut contents in the diet
 713 assessment of filter-feeding Asian carps. *Fundamental and Applied Limnology* 187: 79–86.

714 Vörös, L., I. Oldal, M. Présing & K. V.-Balogh, 1997. Size-selective filtration and taxon-
 715 specific digestion of plankton algae by silver carp (*Hypophthalmichthys molitrix* Val.).
 716 Hydrobiologia 342/343: 223–228.

717 Xie, P. & J. Liu, 2001. Practical Success of Biomanipulation using Filter-feeding Fish to
 718 Control Cyanobacteria Blooms. A Synthesis of Decades of Research and Application in a
 719 Subtropical Hypereutrophic Lake. The Scientific World Journal 1: 337-356.

720 Yang, Y., X. Huang & J. Liu, 1999. Long-term changes in crustacean zooplankton and water
 721 quality in a shallow eutrophic Chinese lake densely stocked with fish. Hydrobiologia 391:
 722 195–203.

723 Zhang, X., P. Xie & X. Huang, 2008. Review of Nontraditional Biomanipulation. The
 724 Scientific World Journal 8: 1184–1196.

725 Zhou G., X. Thao, Y. Bi & Z. Hu, 2011. Effects of silver carp (*Hypophthalmichthys molitrix*)
 726 on spring phytoplankton community structure of Three-Gorges Reservoir (China): results
 727 from an enclosure experiment. Journal of Limnology 70: 26–32.

Figure captions

Fig. 1 Morphological variability of gill rakers of hybrid bigheaded carps in Lake Balaton, Hungary. Five morphotypes of gill raker structures were defined ranging from the comb like, bighead carp type (type GR1; a) through transitional forms (type GR2-4; b-d) to the sponge like, silver carp type (type GR5; e).

Fig. 2 Mean chlorophyll-a concentration ($\mu\text{g L}^{-1}$) and percentage taxonomic composition of phytoplankton by biovolume (b) as well as mean total dry biomass (mg L^{-1}) and percentage taxonomic composition of zooplankton by dry biomass (b) in Lake Balaton, Hungary by sampling dates. Time series of chlorophyll-a concentration is indicated by continuous line for monthly samples and by dotted line for loose sampling periods.

Fig. 3 Gill raker structure demonstrated by the relative area of different morphotype segments of each bigheaded carp analysed for food composition in Lake Balaton, Hungary. Each column represents a fish in the sample. GR1 to GR5 gill raker morphotypes are specified in Fig. 1.

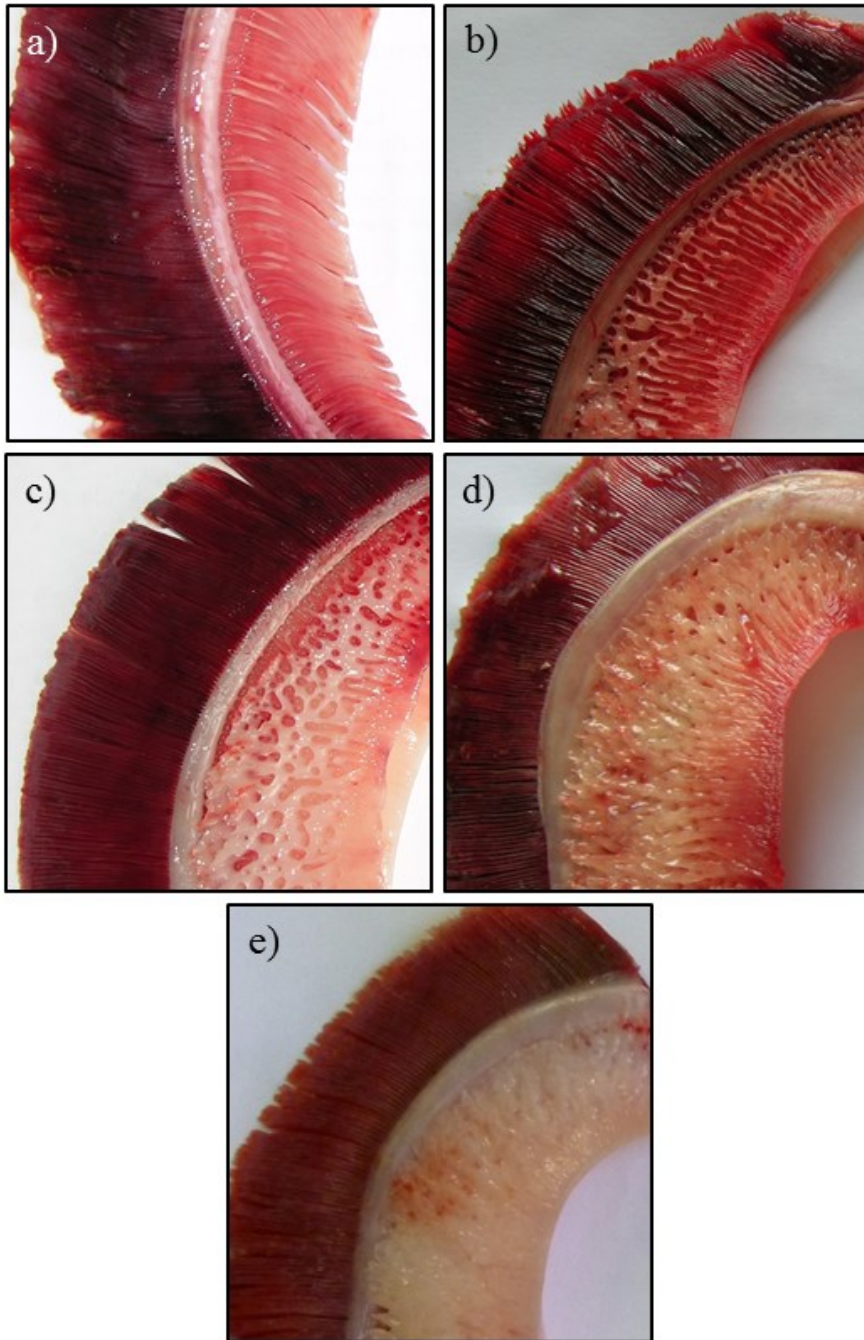
Fig. 4 Percentage contribution of zooplankton, phytoplankton and inorganic matter by sampling occasions to the gut content of bigheaded carps in Lake Balaton, Hungary.

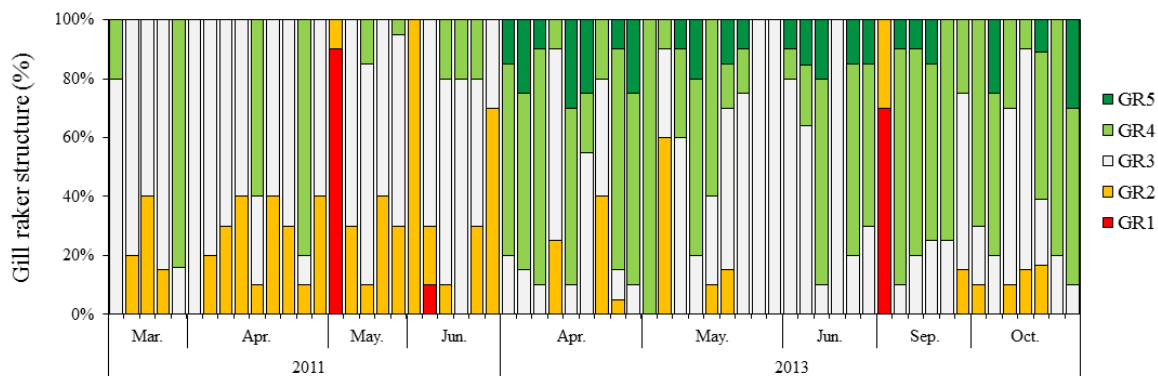
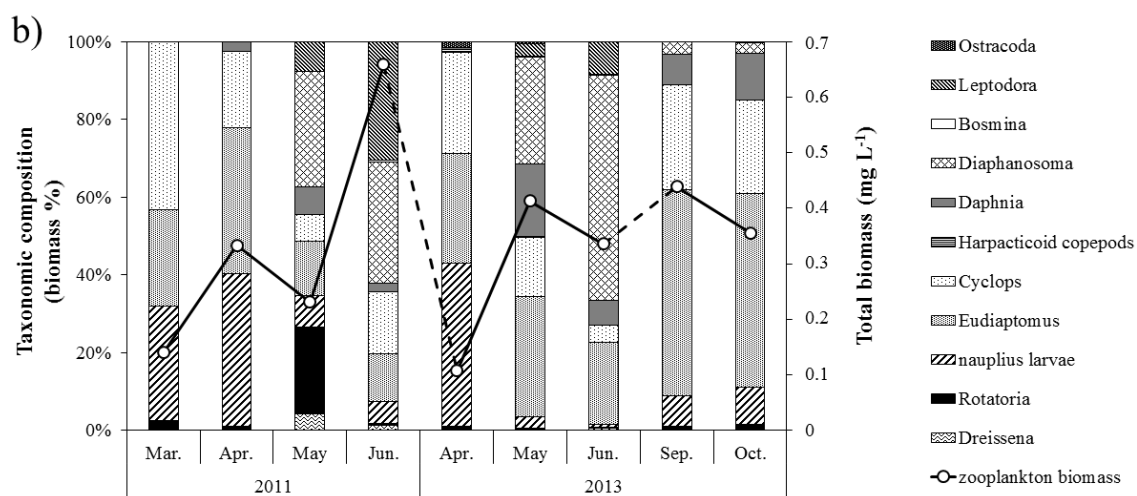
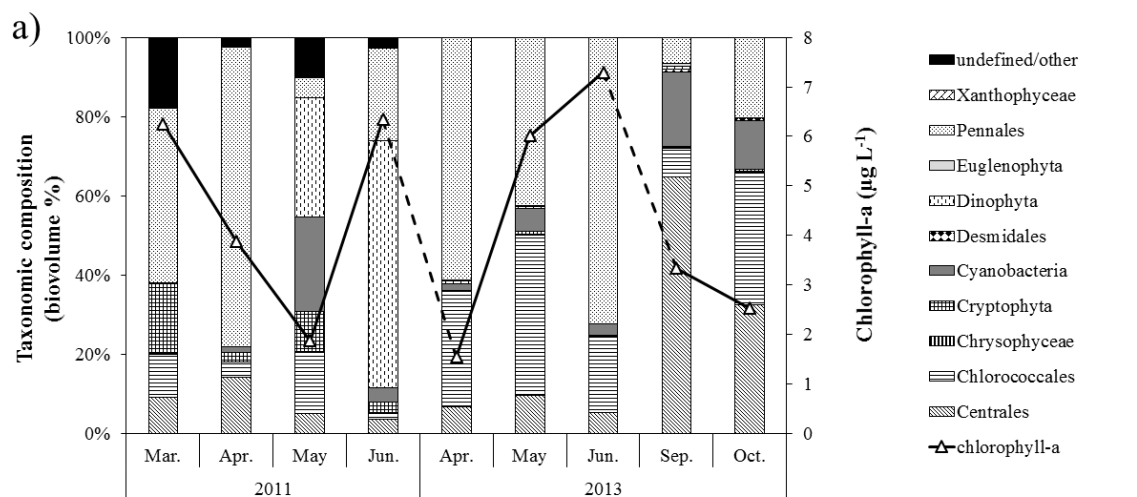
Fig. 5 Feeding strategy of bigheaded carps by sampling occasions plotted according to Amundsen et al. (1996). The horizontal axis represents the relative role of individual to stock level utilization of particular food components (o, zooplankton groups written in italic letters; Δ , algae groups written in normal letters; ■, inorganic matter written in underlined letters). Whereas, the vertical axis represents the rate of specialization with low scores indicating

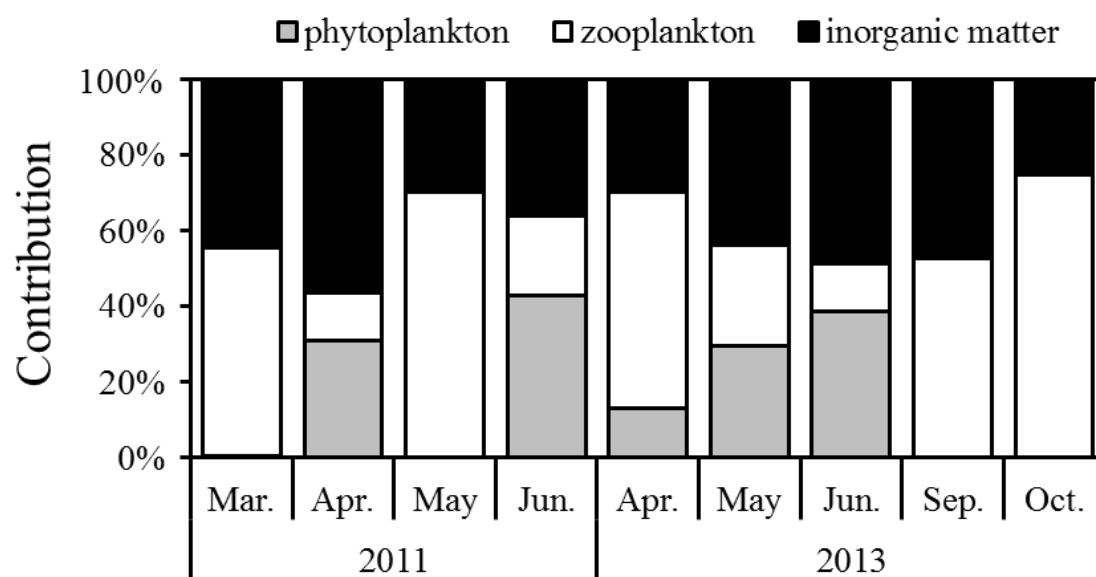
generalization and scores close to 100% indicating strong specialization. Thus, the overall importance of a food component increases from the lower left to the upper right corner of the plot and the sum of products of x and y axis scores for all food components equals one. Centr., Centrales; Penn., Pennales; Dreis., *Dreissena* larvae; Rota., Rotatoria; naupl., nauplius larvae of copepods; Eudia., *Eudiapthomus* spp.; Cycl., *Cyclops* spp.; Harp., Harpacticoid copepods; Bosm., *Bosmina* spp.; Daph., *Daphnia* spp.; Lept., *Leptodora kindtii*; Ostr., Ostracoda; inorg., inorganic matter.

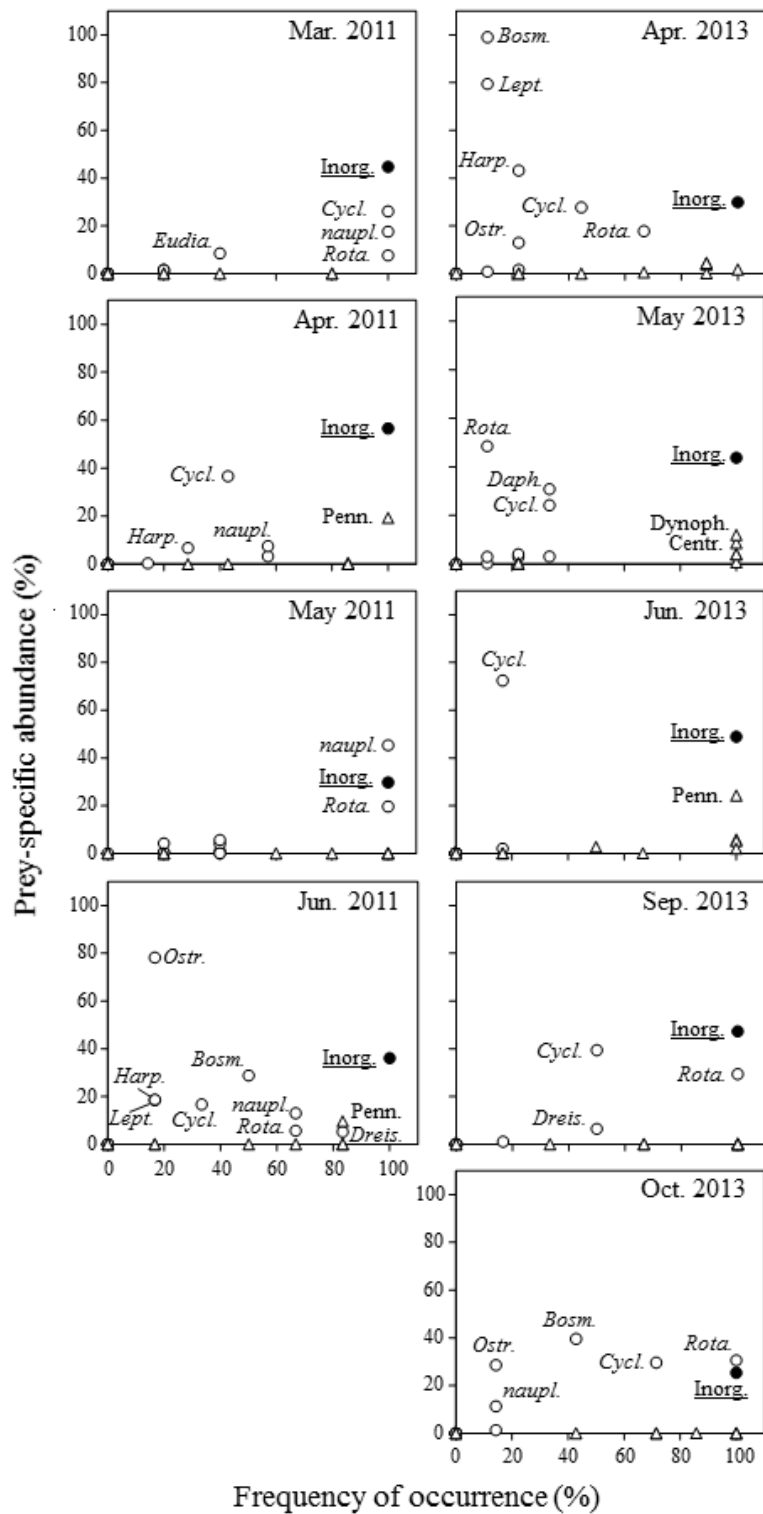
Fig. 6 Canonical correspondence analysis plot describing the relationship between the percentage gut content composition (o, zooplankton groups written in small italic letters; Δ, algae groups written in small normal letters; ■, inorganic matter written in small underlined letters) of hybrid bigheaded carps and forward selected, significant (at $P < 0.05$) environmental factors, seasonality and certain individual traits of fish (→, continuous explanatory variables; and ●, binary dummy explanatory variables written in large letters) in Lake Balaton, Hungary. Some passive, supplementary variables are also plotted in grey. Percentage variance represented by axes are indicated in brackets (of diet data; of diet-explanatory variables relation) after the axis name (for a more detailed statistics see Table 2). Scale factor for plotting is 4.9.

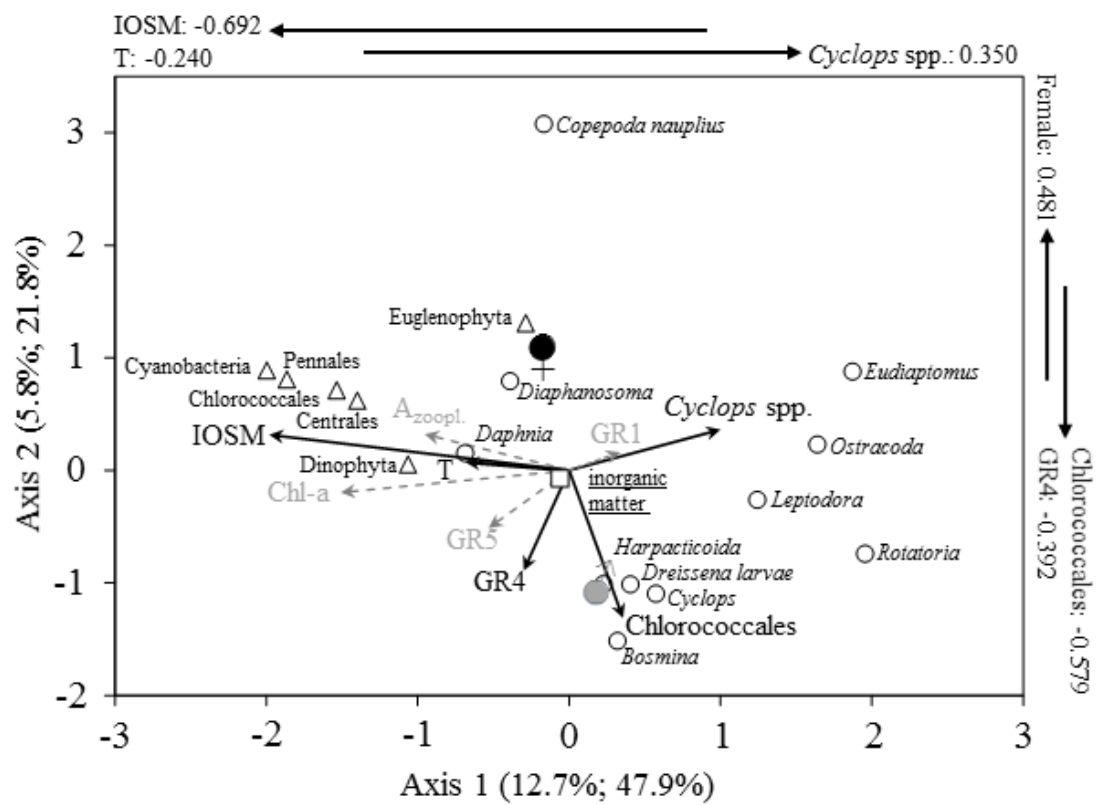
Fig. 7 Stable isotope values (‰) of bigheaded carps (◆) and their potential preys (●, zooplankton; ▲, seston/phytoplankton). Each symbol shows average value of one sampling month. The empty symbols and whiskers represent the mean and the SD of a given sample type.



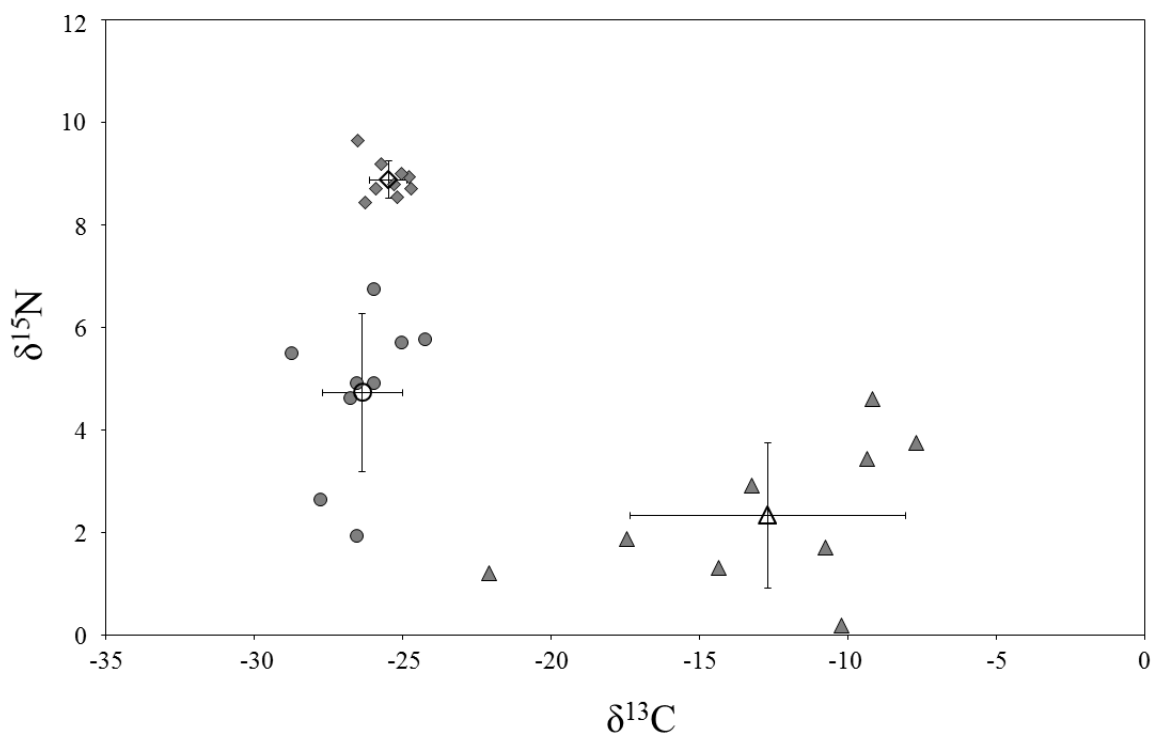








784



785

786

787

Table 1. Means and ranges of water temperature (T, °C), conductivity (Cond., $\mu\text{S cm}^{-1}$), Secchi depth (cm), total suspended matter concentration (TSM, mg L^{-1}) and inorganic suspended matter concentration (IOSM, mg L^{-1}) at bigheaded carp sampling sites in Lake Balaton, Hungary.

		Temperature		Conductivity		Secchi depth		TSM		IOSM	
		mean	SD	mean	SD	mean	SD	mean	SD	mean	SD
2011	Mar.	8.9	0.4	0.677	0.002	79	5.5	9.49	0.40	9.23	0.51
	Apr.	14.2	0.3	0.688	0.003	51	3.2	16.27	0.50	12.87	0.45
	May	21.3	1.9	0.725	0.087	157	2.9	3.06	0.28	2.20	1.18
	Jun.	20.4	0.1	0.920	0.000	36	1.2	33.17	1.26	22.33	0.60
2013	Apr.	18.2	0.2	0.711	0.002	61	5.9	11.30	0.10	8.03	0.23
	May	15.1	0.1	0.695	0.002	34	3.0	16.57	7.13	12.20	6.41
	Jun.	20.3	0.2	0.696	0.004	31	2.1	39.40	0.69	18.40	4.16
	Sep.	12.7	0.3	0.762	0.003	92	2.5	7.33	0.96	3.87	0.68
	Oct.	14.9	0.4	0.758	0.003	81	1.2	12.89	1.42	6.47	0.74

Table 2. Forward selected significant (F and P are indicated) explanatory variables and their percentage explanatory power in total and from pure effect in the canonical correspondence analysis of the percentage gut content data of hybrid bigheaded carps in Lake Balaton, Hungary.

Explanatory variables in the final model	F	P	Total explained variance (%)	Explained variance as pure effect (%)
Inorganic suspended matter	6.82	<0.001	10.51	8.33
Share of Chlorococcales in the phytoplankton	3.04	0.002	4.84	3.32
Female	2.42	0.011	4.14	2.92
Share of <i>Cyclops</i> spp. in the zooplankton	1.97	0.038	4.32	3.71
Water temperature	2.10	0.028	3.49	3.45
Area of gill rake type 4	1.90	0.054	2.62	2.18
Full model (all axes)	3.18	<0.001	26.48	-