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4

5 **Abstract**

6 Worldwide reforestation has been recommended as a landscape restoration strategy to
7 mitigate climate change in areas where the climate can sustain forest. This approach may
8 threaten grassland ecosystems of unique biodiversity as such policies are based on **the** false
9 assumption that most grasslands are man-made. Here, we use multiple lines of evidence
10 (palaeoecological, pedological, phylogenetic, palaeontological) from Central Eastern Europe
11 and show that various types of grasslands have persisted in this area throughout postglacial
12 **i.e.** the past 11,700 years. A warm and dry climate, frequent **fires, herbivore pressure,** and
13 early Neolithic settlements kept forests open until widespread forest clearance **beginning**
14 4000-3000 years ago. Closed forest cover has been the exception for **the past** two million
15 years. **This long-term persistence has likely contributed to the high biodiversity of these**
16 **grasslands.** Consequently, we call for a more **cautious** prioritisation of the protection of what
17 may be erroneously considered natural, **i.e.** forests, by many environmental specialists and
18 managers. Instead we provide a new framework for a better understanding of the evolution
19 and persistence of different grassland types and their biodiversity, so that grasslands can be
20 better understood, valued and conserved.

21
22 Keywords: ancient grasslands, anthropogenic disturbance, climate change mitigation, fire,
23 fossil records, herbivores
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25
26 **1. Introduction**

27 The World Resources Institute (WRI) Atlas of Forest and Landscape Restoration was
28 designed to identify opportunities for landscape restoration worldwide, an initiative supported
29 by several international organisations concerned with land degradation, climate change
30 mitigation and biodiversity loss (WRI, 2015). It contrasts the potential extent of tree cover
31 based on climate conditions with the current distribution of forest globally. This map identifies
32 23 million km² of land worldwide suitable for tree planting, mostly being currently open
33 landscapes with grassland (<http://www.wri.org/applications/maps/flr-atlas>). However,
34 grassland experts oppose the offsetting of agricultural deforestation through the afforestation
35 of grassy ecosystems arguing that this approach ignores the unique biodiversity, cultural
36 significance and important ecosystem services provided by this **ecosystem** (Willis et al.,
37 2008; Parr et al., 2014; Veldman et al., 2015a,b; Bond et al., 2016; **Joshi et al., 2018**). Their
38 imperative is to map “old-growth” grasslands, where tree cover is naturally sparse and where
39 such afforestation would be detrimental. Another important debate is how to provide
40 strategies to reduce the impact of the ongoing abandonment of high biodiversity grasslands,
41 therefore **preventing** succession towards low biodiversity secondary shrub and forest
42 communities (Biró et al., 2010; Valkó et al., 2018a).

43 Here we go further and challenge the perception that treeless areas in temperate regions,
44 where the current climate would permit forest development, have all previously been forested
45 and therefore grasslands and open canopy woodlands are secondary habitat types in these
46 regions. This perception fails to consider the vital role of natural disturbances such as fire or

47 herbivores (Bond and Keeley, 2005). We illustrate this by analysing evidence from Central
48 Eastern Europe. The choice of this region is based on the following arguments: i) it hosts one
49 of the largest tracts of grasslands in Europe crucial for maintaining biodiversity in European
50 agricultural landscapes (Wilson et al., 2012); ii) has one of the highest small-scale species
51 diversities in the world (Dengler et al., 2014; Turtureanu et al., 2014; Chytrý et al., 2015); and
52 iii) is a transitional, complex region between closed forest and steppe biomes (Bohn et al.,
53 2003) and hence contains grasslands of diverse origin and history. Yet, despite these
54 features, grasslands are rarely highlighted as biodiversity hotspots. This is because the key
55 characteristics and ecological processes important for this classification, such as biodiversity
56 intactness and a lack of human disturbance, cannot be readily applied to them (Mittermeier
57 et al., 2011). Our goal is to better define grassland types based on their origin, age and the
58 drivers of their formation and maintenance so that grasslands can be better understood,
59 valued and conserved.

60

61 **2. A new framework for defining European grassland types**

62 Bohn et al. (2003) provided a geobotanical expert assessment, which maps Europe's
63 potential natural vegetation (PNV) *i.e.* the vegetation cover that would exist today in the
64 absence of human activity. It defines most areas that are currently covered by grasslands or
65 open woodlands in Central Eastern Europe as dominated by deciduous broadleaved forest
66 or mixed coniferous and broadleaved forest (Fig. 1). Open, or at least partly open vegetation
67 types, are only recognised in the lowlands of the Carpathian Basin. Fossil records show that
68 grasslands and open canopy woodlands covered extensive areas in Central Eastern Europe
69 during the Pleistocene (*i.e.*, the past 2 million years) when cold and dry climate conditions
70 prevailed (Kuneš et al., 2008; Ellenberg and Leuschner, 2010; Feurdean et al., 2014;
71 Magyari et al., 2014). Warmer climate conditions during the Holocene (*i.e.* the last 11,700
72 years) then greatly reduced the potential distribution and/or extension of grasslands (Birks
73 and Willis, 2008). An especially critical period for grassland persistence was the mid
74 Holocene period (9000-4500 cal yr BP), when moister climatic conditions triggered forest
75 expansion (Roberts et al., 2018). Identification of warm/moist stage refugia for grasslands
76 (*i.e.* locations where they persisted) during the mid Holocene is therefore of crucial
77 importance for understanding ancient grasslands.

78 It is widely accepted that natural grasslands growing on rocky skeletal and other poor soils
79 with a permanent or seasonal moisture deficit, *i.e.* outcrops, steeper slopes, gravel
80 riverbanks, salt and sandy soils the so-called primary grasslands have survived continuously
81 in small pockets throughout the Holocene in their current locations (Lang, 1994; Poschlod
82 and WallisDeVries, 2002). Currently, the extent of these grassland types is limited, apart
83 from grasslands growing on salt and dry sandy soils, which are more common in the
84 Carpathian Basin (Molnár and Borhidi 2003; Deák et al., 2014). It is therefore highly
85 improbable that these small, isolated grassland areas were the only refugia of the
86 extraordinarily rich grassland flora of so-called "semi-natural" grasslands. The high genetic
87 diversity of some grassland plant species in this region (Turtureanu et al., 2014) and the
88 remarkable species richness and endemic plant and animal species typical for grasslands
89 (Chytrý et al., 2015) suggest a wider extent of primary grasslands during the Holocene.

90 In contrast to the primary grasslands, open canopy woodlands and grasslands currently
91 found in areas where climate and soils would allow forest growth, and which are only
92 extensively managed (*i.e.*, no artificial fertiliser and pesticide application) are considered to
93 be semi-natural (Pärtel et al., 2005; Leuschner and Ellenberg, 2017). Here, we challenge the
94 view that most of these grasslands have replaced formerly naturally occurring forests within
95 recent centuries or millennia. We present multiple lines of evidence (palaeobotanical,
96 pedological, phylogenetic, palaeontological) from five countries in Central Eastern Europe, a
97 region with some of the highest-biodiversity grasslands of the world. Firstly, we review
98 published direct records of past grassland occurrence (pollen, plant macrofossils, charcoal)

99 from both natural (lakes, bogs) and archaeological archives, alongside other indirect fossil
100 (pedological, zoological) and recent genetic evidence from Central Eastern Europe.
101 Secondly, we review the characteristics of the environmental and disturbance factors
102 (climate, fire, herbivores and human impact) during the Holocene in this region. Finally, we
103 evaluate whether continuous grassland presence was possible under the Holocene climatic
104 conditions, fire and grazing regimes, and increasing anthropogenic impacts.
105 Based on their age and the drivers of their formation, we set out a new framework for three
106 types of grasslands in Central Eastern Europe: 1) Primary, natural grasslands on skeletal
107 and other poor soils, which have existed throughout the Holocene until the present; 2)
108 Primary, ancient grasslands on deeper soils, maintained by climate and disturbances during
109 the early Holocene and then predominantly by disturbances until the present; and 3) Semi-
110 natural grasslands, extensively managed grasslands, formed and maintained by
111 anthropogenic disturbances during the late Holocene (Table 1). We argue for the recognition
112 of the importance of previously overlooked ancient grasslands that have persisted throughout
113 the Holocene, maintained by natural and later also by anthropogenic disturbances.

114 **3. Multi-proxy evidence for grassland persistence during the Holocene**

115 **3.1 Fossil plant evidence**

116 Pollen and palaeobotanical records from natural archives (lakes and peatbogs) in currently
117 grassland-rich areas in Central Eastern Europe indicate the prevalence of a more open
118 landscape between 11,700 and 9000 cal yr BP and the maximum extent of forest cover
119 between 9000 and 4500 cal yr BP (Figs.1, 2; Table 2; Fig.S1). While these studies show a
120 reduction in grassland cover, especially of xerothermic and floodplain grasslands during the
121 mid Holocene, there is no evidence of their widespread disappearance. Rather, there is
122 strong support for their persistence, given the concurrent presence of many grasslands, i.e.
123 *Adonis* spp., *Artemisia* spp., *Centaurea* spp., *Festuca rubra*, *Festuca* spp., *Filipendula* spp.,
124 *Helianthemum* spp., *Potentilla erecta*, *Potentilla* spp., *Sanguisorba* spp., *Trifolium* spp.,
125 *Thymus* spp., and light-demanding tree and shrub taxa during this period (Fig. 1; Table 2).
126 Archaeobotanical reports from Hungary and Poland suggest an even greater proportion of
127 heliophilous taxa growing locally than pollen records indicate (Fig. 1; Table 2). Remains of
128 grassland species including steppe elements, i.e. *Asperula cynanchica*, *Phleum pratense*,
129 *Plantago media*, *Stipa pennata*, *Stipa* sp., *Silene vulgaris*, *Teucrium chamaedrys*, have been
130 reported from archaeobotanical records of early Neolithic sites from Germany (Fig. 1; Table
131 2) and Czech Republic (Archaeobotanical Database of the Czech Republic;
132 <http://www.arup.cas.cz>). Calcareous grasslands have also been identified in the Neolithic
133 lakeshore sites in the northern foothill of the Alps (Fig. 1; Table 2). The species found are
134 typical of closed xerophilous and mesophilous grasslands at sites where trees would have
135 been able to grow under the climatic conditions of the mid Holocene. Taken together, fossil
136 plant evidence from natural archives and archaeological sources suggest that grasslands
137 existed locally before the start of the Neolithic and therefore before marked human impacts.
138 These findings also demonstrate grassland persistence throughout the mid Holocene,
139 although archaeological records show the occurrence of higher grassland diversity than that
140 found in pollen records.

141 **3.2. Zoological evidence**

142 Indirect evidence for the persistence of open, or partly open landscapes throughout the mid
143 Holocene comes from the palaeontological remains of animal species restricted to extensive,
144 open habitats (Fig. 1). Results from the Carpathian Basin (Hungary and Romania) show that
145 several species typical of steppic environments e.g. *Asinus hydruntinus* (European Wild
146 Ass), *Equus ferus* subsp. *gmelini* (Eastern European Wild Horse), *Microtus gregalis* (Narrow
147 Headed Vole), *Ochotona pusilla* (Steppe Pika), *Otis tarda* (Great Bustard) and *Vipera ursinii*
148 subsp. *rakosiensis* (Meadow Adder) were abundant during the early Holocene (Németh et
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150

151 al., 2017). They became discontinuously present from 8000 cal yr BP and several of these
152 species disappeared between 5000 and 4000 cal yr BP at a time of increased anthropogenic
153 pressure, but in a grassland landscape. Fossil malacological records in currently grassland-
154 rich landscapes reveal a similar picture (Fig. 1); a continuous Holocene presence of strictly
155 open habitat molluscs (*Chondrula tridens*, *Helicopsis striata*, and *Vallonia pulchella*) unable
156 to survive in closed forests (Ložek, 2005; Horsák et al., 2009; Moskal-del Hoyo et al., 2018).
157 Overall, while there is evidence that many open habitat mollusc species contracted their
158 range during mid Holocene forest expansion, there is also robust support for their local long-
159 term persistence, and consequently also for the continuity of grassy ecosystems.

160

161 **3.3 Phylogeographic evidence**

162 Phylogeographic analysis represents a further source of data facilitating the interpretation of
163 the distribution of past grasslands. A pattern of genetic diversity decline from core
164 populations in southern Siberia towards the range periphery of smaller populations in
165 western Europe has been confirmed for several grassland plant species including *Adonis*
166 *vernalis* (Hirsch et al., 2015), *Iris aphylla* (Wroblewska 2008), *Stipa capillata* (Wagner et al.,
167 2011) and *Stipa pennata* (Wagner et al., 2012), These studies have also revealed a
168 surprisingly low genetic differentiation between central and peripheral intermediate
169 populations, or a complete lack of private alleles among peripheral populations (e.g. Wagner
170 et al. 2011; Hirsch et al., 2015), which may reflect the absence of any long-standing isolation
171 of these populations. This implies that these species must have had a more continuous past
172 distribution in Central and Eastern Europe allowing gene flow and interbreeding. Rapid
173 progress in the field of DNA analysis, especially environmental DNA, may shed further light
174 on the origin and past range distribution of grasslands (Thomsen and Willerslev, 2015)

175

176 **3.4 Pedological evidence**

177 A further line of evidence used in the interpretation of former vegetation distributions comes
178 from soil types (IUSS WRB 2006). In Central Eastern Europe, dark soils (chernozems) from
179 steppe and forest steppe zones are considered to have developed before the spread of
180 forests and to have persisted under open or semi-open vegetation (Pokorný et al., 2015).
181 Others, however, view these soils as having survived under forest development (Eckmeier et
182 al., 2007). A palaeo-pedological analysis from the Transylvanian Basin (Romania) shows the
183 occurrence of islands of dark soils of Pleistocene age (20,000-14,000 cal yr BP), which, in
184 drier areas, persisted until the present (Pendea et al., 2002) suggesting grassland
185 persistence throughout the Holocene. However, in other areas of the Transylvanian Basin,
186 dark soils were overlain by Luvisols, typical of nemoral forests, about 5000 cal yr BP (Timar
187 et al., 2010) when wetter climatic conditions prevailed, whilst the current vegetation is
188 predominantly grassland. Grassland occurrence on soils typically favouring forests may be
189 explained by the prevalence of open woodlands throughout the Holocene, allowing the long-
190 term persistence of dark soil, rather than the [post-deforestation](#) formation of this soil type.
191 Forest soils occur extremely rarely on chernozems developed on loess substrates in the
192 Hungarian Plain therefore suggesting the long-term existence of steppe grasslands (Máté
193 1957, Molnár 2010). Indeed, the continuous dominance of grasslands from the Late
194 Pleistocene on loess deposits in the southern Carpathian Basin has been recently
195 demonstrated on the basis of n-alkane biomarkers (Marković et al., 2018). Thus, pedological
196 evidence from chernozems, including those developed on loess, shows that chernozems
197 existence under open or semi-open vegetation.

198

199 **4. Drivers of grassland persistence during the Holocene**

200 **4.1 Climate conditions**

201 Proxy-based and climate simulations indicate warmer-than-present summer temperatures,

202 lower precipitation and soil moisture, and greater seasonality in the early Holocene (11700-
203 9000 cal yr BP) in Central and Eastern Europe at the time of maximum grassland extent
204 (Feurdean et al., 2013; 2014; Heiri et al., 2014). Palaeoclimatological records show a decline
205 in temperatures in this region from approximately 9000 to 4500 cal yr BP (Heiri et al., 2015;
206 Tóth et al., 2015; Hajkova et al., 2016). Climate simulations are consistent with this pattern of
207 mid Holocene cooling, but also marked precipitation and soil moisture increases in the mid
208 latitudes in Europe (Feurdean et al., 2013). Palaeoecological reconstructions show that
209 significant forest expansion occurred in response to cool and moist conditions in Central
210 Eastern Europe (Magyari et al., 2010; Feurdean et al., 2015; Kuneš et al., 2015; Novenko et
211 al., 2016; Pokorný et al. 2015; Jamrichová et al., 2017; Moskal-del Hoyo et al., 2018).
212 Therefore, both proxy and modelled palaeoclimatic and palaeoecological evidence from the
213 lowlands of Central Eastern Europe clearly show that grasslands were most extensive during
214 the warm and dry climatic conditions, with prolonged droughts, of the early Holocene
215 (11,700-9000 cal yr BP) and became restricted under wetter conditions during the mid
216 Holocene (9000-4500 cal yr BP; Fig. 2). This illustrates the stronger competitive advantage
217 of grasslands over trees when resources are limited i.e., lower moisture availability and
218 prolonged droughts.

219

220 **4.2 Fire**

221 The role of fire as one of the main drivers of the rise in grassy ecosystem during the Miocene
222 has been advocated (Osborne and Behling, 2006; Strömberg, 2011) and confirmed by
223 fossil records from C₄ dominated grasslands in Africa (Hoetzel et al., 2013). This is not
224 surprising as dominant grassland species have fine fuels with rapid curing and fast regrowth
225 rates as well as perennating buds near or below the soil surface; adaptations that ensure
226 regeneration after disturbances that damage the above-ground parts of the plant (He and
227 Lamont, 2018). Thus, whilst grass and herbs can withstand frequent fire, this shift in fire
228 regime may have harmed previously dominant tree species adapted to infrequent fire (He
229 and Lamont, 2018). Although, short-term, field-based burning experiments in Hungary found
230 conflicting results about the effect of fire on grassland biodiversity (Valkó et al., 2014, 2018b),
231 controlled laboratory and small-scale field experiments examining the effect of fire on seeds
232 found a predominantly negative effect of fire on seed germination in grassland species,
233 however, some positive effects (*Fabaceae*) also emerged (Ruprecht et al., 2013; 2015).
234 Disturbances by fire have recently been considered essential for increased grassland
235 [competitive advantage](#) over trees during the Holocene in Central Eastern Europe (Magyari et
236 al., 2010; Feurdean et al., 2015). For example, a positive effect of frequent fires on the
237 competitive advantage of grasses over trees and, by this means, on the extent of grassland
238 in Transylvania, Romania, has been inferred from sedimentary charcoal particles and pollen
239 (Feurdean et al., 2013). On the contrary, the decline in fire frequency during the mid
240 Holocene has been shown to be detrimental for grassland extent. [The significance of global
241 fire activity in grassy biomes during the early and late Holocene, based on charcoal datasets,
242 has recently been emphasised by Leys et al \(2018\)](#). Taken together, neo- and
243 palaeoecological evidence indicate that fire may have had a more important role in the shifts
244 between forest and grassland, and in grassland maintenance, than previously thought.
245 Exploring to what extent temperate grasslands are associated with frequent fires and which
246 grassland species/communities are most resilient or benefit mostly from fire could be useful
247 for the enhancement of management practices, i.e., preventing excessive dominance by
248 competitor grass species as well as the succession towards shrublands and forests.

249

250 **4.3 Herbivores**

251 Large herbivorous mammals influence the physiology and growth of plants and are
252 considered ecological keystones in maintaining tree–grass coexistence (Crawley, 1983;

253 Sankaran et al., 2005). Released from megaherbivore pressure and with a change in climate
254 at the Pleistocene-Holocene boundary, European lowlands witnessed substantial forest
255 regeneration and a reduction in vegetation openness (Vera, 2000; Svenning, 2002).
256 Megaherbivore extinction may also have had a cascading effect on the population size and
257 diversity of small mammals dependent on vegetation openness and indirectly on the fire
258 regime (Gill et al., 2014). However, comparatively, little attention has been given to the effect
259 of the declining population size or extinction of wild herbivores, or the subsequent role of
260 livestock, on landscape structure during the mid to late Holocene. We know from
261 palaeontological and archaeozoological records in the Carpathian Basin that several large
262 herbivores i.e., *Alces alces* (Eurasian Elk), *Bison bonasus* (European Bison), *Dama dama*
263 (Eurasian Fallow Deer), *Equus ferus* subsp. *gmelini* (Wild Horse) and *Equus hemionus*
264 (Asiatic Wild Ass) became discontinuously present from the mid Holocene, i.e. 8000 cal yr
265 BP and that many became extinct by 4000 cal yr BP (Németh et al., 2017; Bejenaru et al.,
266 2018). In contrast, livestock numbers increased from 6500 cal yr BP (Schumacher et al.,
267 2016). Domestic livestock could prevent forest encroachment in the absence or with a low
268 density of wild herbivores. Domestic animals can replace wild herbivores as dispersal agents
269 (Bruun and Fritzboeger, 2002; Cosyns et al., 2005), however, their movement is limited by
270 agricultural practices. Combined fossil records of fauna, pollen and coprophilous fungi that
271 reproduce exclusively on animal dung (*Sporormiella* spp., *Sordaria* spp., *Podospora* spp.)
272 can provide means of assessing the effects of herbivores on grassland dynamics and also
273 the timing of the shift in influence from grazing by wild herbivores to livestock grazing (Gill et
274 al., 2009). Such records are still scant in Europe, but the existing studies generally show the
275 increasing effects of grazing by domestic livestock from 5000 cal yr BP (Schumacher et al.,
276 2016). Understanding the responses of grasslands to different grazing animals (body size,
277 grazing intensity and height, foraging strategy and forage selectivity) will be essential in the
278 development of future grassland management strategies as various forms of livestock
279 grazing have been proposed to simulate the effects of grazing and browsing by wild
280 herbivores (Poschlod and WallisDeVries, 2002; Bakker et al., 2004; Tóth et al., 2016;
281 Poschlod, 2017).

282

283 **4.4. Early human impact counteracted the encroachment of forest onto primary** 284 **grasslands**

285 As the increase in forest cover from 8000 cal yr BP coincided with the spread of Neolithic
286 culture across South Eastern Europe (Bogaard et al., 2004; Kreuz 2008), a critical question
287 in respect to grassland extent is whether anthropogenic impacts could have counteracted the
288 climate-driven development of a closed forest (Pokorný et al., 2015). Archaeological datasets
289 from this part of Europe indicate that Neolithic settlements tended to be established in open
290 landscapes and that field sizes were small (Moskal-del Hoyo et al., 2013; Chapman, 2017;
291 Marinova and Ntinou, 2017). As people first settled in naturally open landscapes, this
292 tendency could explain the apparent lack of major deforestation at this time in the pollen
293 records from Central Eastern Europe (Fig. 2). Fire activity was naturally high during the early
294 Holocene (Magyari et al., 2010; Feurdean et al., 2013), and humans may have taken
295 advantage of wildfires to extend their agro-pastoral activities into freshly burned habitats. In
296 agreement with Pokorný et al. (2015) we hypothesise that early anthropogenic land
297 management may have slowed, or partially arrested, the development of closed forest
298 favoured by the wetter climatic conditions of the mid Holocene at locations with low biomass
299 productivity, contributing to the maintenance of landscape openness.

300

301 **4.5. Semi-natural grasslands replacing forests: when and how?**

302 Individual pollen records, as well as large-scale quantitative vegetation reconstructions from
303 Central Eastern Europe, show that the level of anthropogenic impact on forest remained low

304 until about 6000-5000 cal yr BP (Magyari et al., 2010; Feurdean et al., 2015; Kuneš et al.,
305 2015; Jamrichová et al., 2017; Fig. 2). Modelled vegetation and land use (arable and pasture
306 cover) changes across Europe suggest that open areas expanded gradually from previously
307 cleared forest after ca. 6000 cal yr BP (Kaplan et al., 2017). A noticeable increase in the
308 abundance and richness of grassland along with the decline in total forest cover but increase
309 for *Quercus*, a tree taxon typical for woodland and woody pasture, in Central Eastern Europe
310 occurred from 4700-3500 cal yr BP onwards (Jamrichová et al., 2017). This demonstrates a
311 growing anthropogenic role in the extension of grasslands and the formation of open
312 woodlands. These grasslands belong to the so-called semi-natural grasslands that
313 developed from forests and are maintained by land management (Pärtel et al., 2007; Pereira
314 et al., 2017). Technological advances in agriculture and the expansion of urban centres and
315 farms from the Late Bronze Age and Iron Age (3500 cal yr BP) have led to both an extension
316 and intensification of the land use in Central Europe (Poschold, 2015; Rösch et al., 2016). It
317 is therefore not surprising that from this time onwards, the richness and extent of grassland
318 has been found to correlate closely with prehistoric settlement density and land management
319 (Poschold and WallisDeVries, 2002; Pärtel et al., 2005; Hajkova et al., 2011; Hejcman et al.,
320 2013; Poschold, 2017). Later on *i.e.* from the 15th to 20th centuries, grassland expansion is
321 strongly linked to sheep flock migration. Livestock acted as dispersal vectors and their
322 mobility may be one of the reasons that ancient and older semi-natural grasslands may have
323 similar species diversity (Poschold and WallisDeVries, 2002; Molnár et al. 2012; Poschold,
324 2017). The sowing of hayseed and mowing may also have promoted grassland expansion in
325 many parts of Europe (Babai and Molnár, 2014). In summary, semi-natural grasslands
326 expanded into formerly forested sites and have subsequently been maintained by a variety of
327 land management practices including grazing, burning and mowing.

329 **5. A new framework for Holocene grassland persistence; conservation consequences**

330 We provide a new framework distinguishing three types of biodiversity-rich grasslands in
331 Central Eastern Europe. [These are: primary grasslands on skeletal and other poor soils](#)
332 [\(primary grasslands I\), ancient grasslands maintained by natural and anthropogenic](#)
333 [disturbances \(primary grasslands II\), and semi-natural grasslands developed as a result of](#)
334 [human activities replacing forests \(Fig. 3; Table 1\).](#) We have identified the reasons for the
335 continuous presence of primary grasslands during the Holocene including both natural *i.e.*
336 climate conditions and soils (primary I), climate and disturbance factors *i.e.* fire and grazing
337 (primary grasslands II). Neolithic people may have first settled in naturally open areas such
338 as grasslands or grassland-woodland mosaics arresting the development of a full forest
339 cover when the climate became wetter (mid Holocene), indirectly favouring the preservation
340 and expansion of grasslands. The intensification of human impact from 4700-3500 cal yr BP
341 onwards subsequently lead to considerable extension of semi-natural grassland on formerly
342 forested sites. A succession from grassland to forest after the cessation of land management
343 (e.g. grazing or burning) is not necessarily a proof against the primary or ancient aspect of
344 grasslands, but may indicate the lack of disturbances. Livestock grazing in the late Holocene
345 has replaced ancient grazing by megaherbivores prevailing until the early Holocene and that
346 by large herbivores throughout the mid Holocene.

347 Our findings also reveal misconceptions about the origin of Central Eastern European
348 grasslands and open canopy woodlands. The concept of a previously continuous, closed
349 forest in extant grassland-rich landscapes where climatic conditions are favourable for forest
350 fails to hold true, as grasslands are likely to have been continuously present throughout the
351 Holocene. The long-term persistence of grasslands at these locations is probably an
352 important reason for one of the highest small-scale species richness, many endemic,
353 worldwide in these habitats. Such species-rich plant communities can require millennia to
354 develop and only well-connected grassland patches can support genetically diverse plant

355 populations. These findings challenge the commonly held view that conservation activities
356 should primarily focus on the protection of forests in many areas of Central Eastern Europe.
357 Anthropogenic impacts tend to focus on forest clearance, but this perspective paper
358 highlights that conservationists and land managers need to carefully consider that, in many
359 cases, it is not primary forests that hold the highest biodiversity. Further, human-made, or
360 managed environments, such as extensively managed grasslands, are long-term landscape
361 features, contain unique plant and animal communities, and provide important ecosystems
362 services. Our findings support the recent wider acceptance of the notion that people and
363 nature should not be separated in the societal discourse of environmental science (Mace
364 2014). Finally, we advocate the need for a more detailed understanding of the role of
365 disturbances in grassland-forest dynamics, to avoid the overly simplistic assumption that
366 sparse tree cover is evidence of past deforestation. Fossil records provide such data and the
367 routine incorporation of palaeoecological investigations into environmental management is a
368 key step in developing science-based evidence for the conservation of the biodiversity of
369 grasslands. Thus, our regional case study supports the advocacy of Willis et al. (2010),
370 Barnosky et al. (2017) and Whitlock et al. (2018) for merging palaeobiology and conservation
371 biology as well as an appreciation of the dynamic history of species and ecosystems,
372 including the role of humans.
373

375 **Figures captions**

376

377 **Figure 1.** Location of the study area in Europe (A) and the distribution of the main vegetation
378 types in Central Eastern Europe based on the potential natural vegetation map of Europe (B;
379 Bohn et al., 2003). Colour symbols show location of various types of fossil records extracted
380 from literature (Table 2 and S1) indicating either continuous grassland presence throughout
381 the Holocene or during the afforestation phases of the mid Holocene (9000-4000 cal yr BP).
382 These fossil records reveal that grasslands were continuously present throughout the
383 Holocene in places where potential natural vegetation has been assumed to be forest.

384

385 **Figure 2.** Pollen based reconstruction of forest (green) versus open land cover (yellow) from
386 Central Eastern Europe during the Holocene using the pseudobiomisation method (Fyfe et
387 al., 2015). Cumulative land cover record was constructed by spatially aggregating 96 pollen
388 records extracted from the Pangaea Database and distributed across the region shown in
389 Fig.1 and Fig. S1. Forest cover includes both broadleaf and conifer trees, whereas open land
390 cover includes pastures/natural grasslands, and arable/disturbed land. Geological and
391 archaeological periods as well as the predominance of each grassland type throughout the
392 Holocene are also highlighted. Trends in simulated growing season temperature and
393 precipitation for Lake Stiucii, Romania after Feurdean et al. (2015).

394

395 **Figure 3.** The effect of climate, soils and disturbances by fire, herbivores and humans on the
396 three types of grasslands and forest. Blue line denote a positive effect, red line a negative
397 effect and grey both effects.

398

399

400 **References**

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661

662 **Supplementary Material**

663

664 **Figure S1** Location of sites extracted from the European Pollen Database (EPD) and used to
665 construct Figure 2.

666 **Table S1.** Location of sites extracted from the literature and used to construct Figure 1.

667

669 **Table 1.** Species-rich grassland types

670 Grassland type	Characteristics
671 Primary grassland I 672	Natural grasslands on skeletal and other poor soils with moisture deficit
673 674 Primary grassland II 675 676	Ancient grasslands formed and maintained mainly by climate conditions and natural fires, herbivores and, later, also influenced by anthropogenic disturbances
677 678 Semi-natural grassland 679 680 681	Secondary grasslands formed and maintained by anthropogenic disturbances (deforestation, livestock grazing, cultivation, use of fire) in areas suitable for forests during the late Holocene

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684

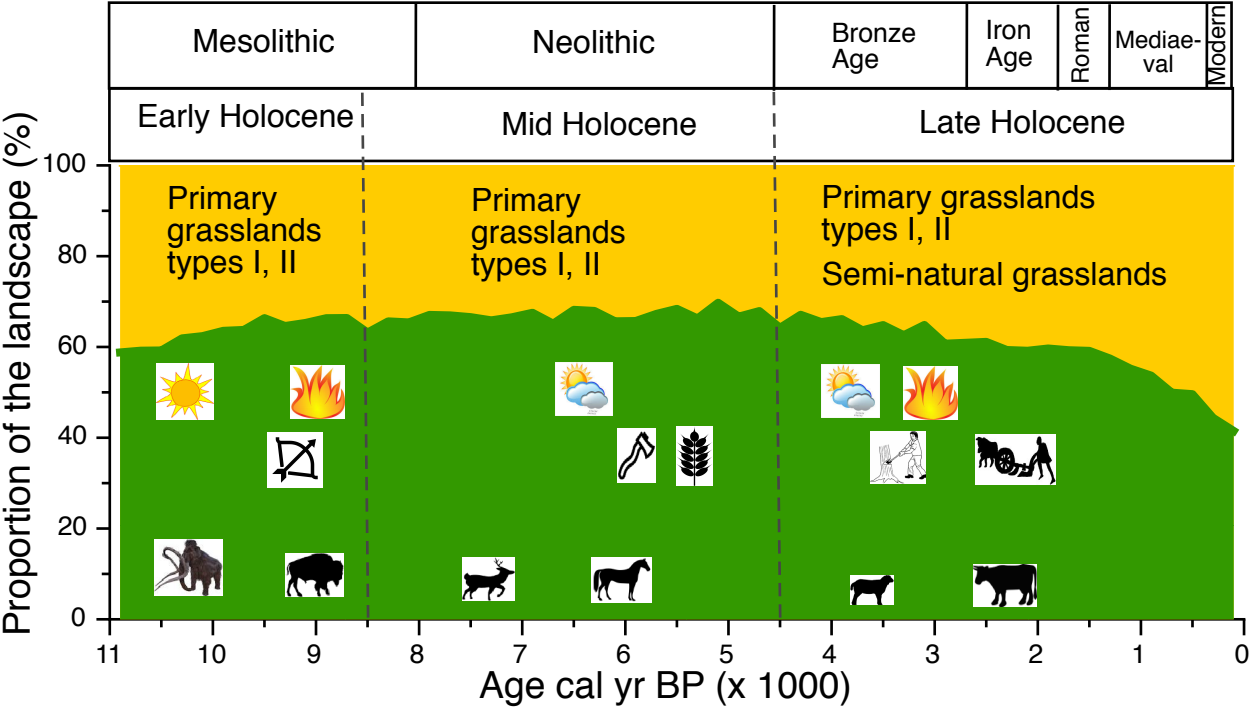
685 Table 2. Fossil pollen, plant macrofossil and charcoal evidence for the presence of various types of grasslands during the early and mid Holocene. Note:
 686 on the basis of pollen analysis differentiation between primary grassland and semi-natural grassland is not always possible.
 687

688	Location	Taxa	Period	References
689	Romania	<i>Artemisia</i> , <i>Aster</i> -type, Caryophyllaceae undiff., Compositae Cichorioideae,	Holocene	Feurdean et al., 2015
690		<i>Centaurea</i> , Chenopodiaceae, <i>Helianthemum</i> , <i>Filipendula</i> , <i>Potentilla</i> , Poaceae,		
691		<i>Plantago lanceolata</i> , <i>Plantago major</i> / <i>P. media</i> , <i>Rumex acetosa</i> , <i>Thalictrum</i> ,		
692		<i>Chenopodium</i> , <i>Suaeda maritima</i> , <i>Festuca</i> , <i>Lycopus</i> , <i>Thymus</i> , Leguminosae undiff.		
693				
694				
695	Hungary	<i>Artemisia</i> , <i>Achillea</i> , <i>Ajuga</i> , <i>Adonis</i> , <i>Astragalus</i> , <i>Allium</i> , <i>Aster</i> -type, Poaceae,	Holocene	Magyari et al., 2010
696		<i>Centaurea</i> , <i>Filipendula</i> , <i>Festuca</i> , <i>Euphorbia</i> , Caryophyllaceae undiff.,		
697		<i>Gagea</i> , Chenopodiaceae, Compositae Cichorioideae, <i>Dianthus</i> -type,		
698		<i>Genista</i> , <i>Helianthemum</i> , <i>Hypericum</i> , <i>Inula</i> , Leguminosae undiff., <i>Lotus</i> -type,		
699		<i>Plantago lanceolata</i> , <i>Plantago major</i> / <i>P. media</i> , <i>Potentilla</i> , <i>Trifolium</i> ,		
700	<i>Rumex acetosa</i> , <i>Teucrium</i> , <i>Thymus</i> , <i>Thalictrum</i> , <i>Verbascum</i> , <i>Sanguisorba</i>			
701				
702				
703	Hungary	<i>Chenopodium album</i> , <i>Echinochloa crus-galli</i> , <i>Fallopia convolvulus</i>	Neolithic	Moskal del Hoyo et al.,
704		<i>Galium spurium</i> , <i>Melandrium album</i> , <i>Plantago lanceolata</i> , <i>Polycneum arvense</i> ,		
705		<i>Polygonum aviculare</i> , <i>Polygonum mite</i> , <i>Polygonum minus</i> , <i>Rumex acetosa</i> ,		
706		<i>Setaria pumila</i> , <i>Setaria viridis</i> , <i>Bromus</i> sp., <i>Chenopodium</i> sp., <i>Galium</i> ,		
707		<i>Ononis</i> sp., <i>Polygonum</i> sp., <i>Rumex</i> sp., <i>Stipa</i> sp., <i>Trifolium</i> sp., <i>Vicia</i> sp.,		
708	Fabaceae, Poaceae, Polygonaceae, Caryophyllaceae			2018
709				
710	Poland (S)	<i>Artemisia</i> , <i>Stipa</i> sp., <i>Knautia arvensis</i> , <i>Hypericum</i> , <i>Plantago media</i> , <i>P. lanceolata</i> ,	Mid Holocene	Moskal del Hoyo et al., 2018
711		<i>Filipendula</i> , <i>Aster</i> -type, Compositae Cichorioideae, Poaceae, Chenopodiaceae		
712				
713	Czech Republic	<i>Artemisia</i> , <i>Achillea</i> , <i>Ajuga</i> , <i>Astragalus</i> , <i>Filipendula</i> , Chenopodiaceae,	Holocene	Kuneš et al., 2015 Jamrichova et al., 2017 Hajek et al., 2016
714		<i>Centaurea</i> , Caryophyllaceae undiff., Compositae Cichorioideae, <i>Genista</i> , <i>Filipendula</i> ,		
715		<i>Helianthemum</i> , <i>Hypericum</i> , <i>Inula</i> -type, <i>Plantago lanceolata</i> , <i>P. media</i> , Poaceae,		
716		<i>Potentilla</i> -type, <i>Ranunculus acris</i> -type, Rubiaceae, <i>Rumex acetosa</i> ,		
717		<i>Teucrium</i> , <i>Thymus</i> , <i>Thalictrum</i> , <i>Verbascum</i>		
718				

719	Austria	<i>Poa pratensis/trivialis, Festuca ovina/rubra, Brachypodium pinnatum,</i>	Neolithic	Körber-Grohne 1990
720		<i>Anthoxanthum odoratum, Stipa pennata</i>		
721				
722	Germany (S)	<i>Alchemilla vulgaris, Asperula cynanchica, Bupleurum falcatum, Carex muricata,</i>	Neolithic	Kreuz et al., 2005, 2008
723		<i>Centaurea, Daucus carota, Euphrasia, Phleum pratense, Poa annua, Polygonum aviculare,</i>		Kreuz and Schäfer, 2011
724		<i>Stipa sp., Stipa pennata, Rumex acetosella, Stellaria graminea, Trifolium campestre,</i>		
725		<i>Galium cf. verum, Galium molugo, Urtica dioica, Teucrium chamaedrys, Chenopodium spp.,</i>		
726		<i>Veronica arvensis, Plantago media</i>		
727				
728	Germany (NW)	<i>Achillea ptarmica, Angelica sylvestris, Anthriscus sylvestris, Artemisia cf. campestris,</i>	Preboreal & Boreal	Knörzer 1996
729		<i>Crepis biennis, Chaerophyllum hirsutum, Dianthus spp., Empetrum nigrum,</i>		
730		<i>Euphorbia cyparissias, Festuca cf. pratensis, Festuca rubra, Filipendula ulmaria, Galium mollugo,</i>		
731		<i>Heracleum sphondylium, Hypochaeris radicata, Juncus cf. effusus, Knautia arvensis,</i>		
732		<i>Lysimachia vulgaris, Laserpitium prutenicum, Poa pratensis / angustifolia,</i>		
733		<i>Polemonium coeruleum, Scabiosa columbaria, Plantago lanceolata,</i>		
734		<i>Potentilla tabernaemontani, Ranunculus acris, Rumex acetosa, Rumex tenuifolius,</i>		
735		<i>Sanguisorba officinalis, Scabiosa aff. columbaria, Silene vulgaris, Thalictrum flavum,</i>		
736		<i>Valeriana officinalis, Valeriana procurrens, Viola canina</i>		
737	Germany (NW)	<i>Achillea millefolium, Leontodon autumnalis, Agrostis tenuis, Melandrium rubrum,</i>	Neolithic	Knörzer 1996
738		<i>Phleum nodosum, Poa cf. pratensis, Artemisia cf. campestris, Poa cf. trivialis,</i>		
739		<i>Polygonum bistorta, Potentilla argentea, Carex ovalis, Prunella vulgaris,</i>		
740		<i>Centaurea cf. nigra, Ranunculus repens, Cerastium cf. semidecandrum, Rumex tenuifolius,</i>		
741		<i>Chrysanthemum leucanthemum, Deschampsia caespitosa, Selinum carvifolium,</i>		
742		<i>Festuca rubra, Silaum silaus, Heracleum sphondylium, Stachys recta, Hieracium pilosella,</i>		
743		<i>Trifolium arvense, Hypericum cf. maculatum, Trifolium dubium, Hypericum tetrapterum,</i>		
744		<i>Valerianella dentata, Juncus bufonius / conglomeratus, Veronica cf. arvensis,</i>		
745		<i>Trifolium repens, Juncus cf. effusus, Viola tricolor</i>		
746				
747				
748				
749				

Highlights

- We prove a new framework for understanding long-term dynamics of grasslands
- This framework facilitates the identification of ancient grasslands worldwide
- Evidence of the long-term grassland continuity [in potential forest areas](#)
- Grassland continuity explains the high current species richness of this ecosystem
 - Better classification highlights the importance of this overlooked environment



Biodiversity-rich European grasslands: ancient, forgotten **ecosystems**

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Conflict of Interest

There is no conflict of interest with any other people or organizations.

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4

5 **Abstract**

6 Worldwide reforestation has been recommended as a landscape restoration strategy to
7 mitigate climate change in areas where the climate can sustain forest. This approach may
8 threaten grassland ecosystems of unique biodiversity as such policies are based on a false
9 assumption that most grasslands are man-made. Here, we use multiple lines of evidence
10 (palaeoecological, pedological, phylogenetic, palaeontological) from Central Eastern Europe
11 and show that various types of grasslands have persisted in this area throughout the
12 postglacial i.e., the past 11,700 years. A warm and dry climate, frequent fires and herbivore
13 pressure, and early Neolithic settlements kept forests open until widespread forest clearance
14 from 4000-3000 years ago. Closed forest cover has been the exception for about two million
15 years. This long-term grasslands persistence is probably important for the high biodiversity of
16 grasslands. Consequently, we call for a more caution prioritisation of the protection of what
17 may be erroneously considered natural, i.e. forests, by many environmental specialists and
18 managers. Instead we provide a new framework for a better understanding of the evolution
19 and persistence of different grassland types and their biodiversity, so that grasslands can be
20 better understood, valued and conserved.

21
22
23 **Keywords:** fossil records, ancient grasslands, fire, herbivores, anthropogenic disturbance,
24 climate change mitigation
25

26
27 **1. Introduction**

28 The World Resources Institute (WRI) Atlas of Forest and Landscape Restoration was
29 designed to identify opportunities for landscape restoration worldwide, an initiative supported
30 by several international organisations concerned with land degradation, climate change
31 mitigation and biodiversity loss (WRI, 2015). It contrasts the potential extent of tree cover
32 based on climate conditions with the current distribution of forest globally. This map identifies
33 23 million km² of land worldwide suitable for tree planting, mostly being currently open
34 landscapes with grassland (Fig. S11). However, grassland experts oppose the offsetting of
35 agricultural deforestation through the afforestation of grassy ecosystems arguing that this
36 approach ignores the unique biodiversity, cultural significance and important ecosystem
37 services provided by this ecosystem (Willis et al., 2008; Parr et al., 2014; Veldman et al.,
38 2015a,b; Bond et al., 2016). Their imperative is to map “old-growth” grasslands, where tree
39 cover is naturally sparse and where such afforestation would be detrimental. Another
40 important debate is how to provide strategies to reduce the impact of the ongoing
41 abandonment of high biodiversity grasslands, therefore hampering succession towards low
42 biodiversity secondary shrub and forest communities (Biró et al., 2010; Valkó et al., 2018a).
43 Here we go further and challenge the perception that treeless areas in temperate regions,
44 where the current climate would permit forest development, have all previously been forested
45 and therefore grasslands and open canopy woodlands are secondary habitat types in these
46 regions. This perception fails to consider the vital role of natural disturbances such as fire or
47 herbivores (Bond and Keeley, 2005). We illustrate this by analysing evidence from Central

48 Eastern Europe. The choice of this region is based on the following arguments: i) it hosts one
49 of the largest tracts of grasslands in Europe crucial for maintaining biodiversity in European
50 agricultural landscapes (Wilson et al., 2012); ii) has one of the highest small-scale species
51 diversities in the world (Dengler et al., 2014; Turtureanu et al., 2014; Chytrý et al., 2015); and
52 iii) is a transitional, complex region between closed forest and steppe biomes (Bohn et al.,
53 2003) and hence contains grasslands of diverse origin and history. Yet, despite these
54 features, grasslands are rarely highlighted as biodiversity hotspots. This is because the key
55 characteristics and ecological processes important for this classification, such as biodiversity
56 intactness and a lack of human disturbance, cannot be readily applied to them (Mittermeier
57 et al., 2011). Our goal is to better define grassland types based on their origin, age and the
58 drivers of their formation and maintenance so that grasslands can be better understood,
59 valued and conserved.

60

61 **2. A new framework for defining European grassland types**

62 Bohn et al. (2003) provided a geobotanical expert assessment, which maps Europe's
63 potential natural vegetation (PNV) i.e., the vegetation cover that would exist today in the
64 absence of human activity. It defines most areas that are currently covered by grasslands or
65 open woodlands in Central Eastern Europe as dominated by deciduous broadleaved forest
66 or mixed coniferous and broadleaved forest (Fig. 1). Open, or at least partly open vegetation
67 types, are only recognised in the lowlands of the Carpathian Basin. Fossil records show that
68 grasslands and open canopy woodlands covered extensive areas in Central Eastern Europe
69 during the Pleistocene (i.e., the past 2 million years) when cold and dry climate conditions
70 prevailed (Kuneš et al., 2008; Ellenberg and Leuschner, 2010; Feurdean et al., 2014;
71 Magyari et al., 2014). Warmer climate conditions during the Holocene (i.e., the last 11,700
72 years) then greatly reduced the potential distribution and/or extension of grasslands (Birks
73 and Willis, 2008). An especially critical period for grassland persistence was the mid
74 Holocene period (9000-4500 cal yr BP), when moister climatic conditions triggered forest
75 expansion (Roberts et al., 2018). Identification of warm/moist stage refugia for grasslands
76 (i.e., locations where they persisted) during the mid Holocene is therefore of crucial
77 importance for understanding ancient grasslands.

78 It is widely accepted that natural grasslands growing on rocky skeletal and other poor soils
79 with a permanent or seasonal moisture deficit i.e., outcrops, steeper slopes, gravel
80 riverbanks, salt and sandy soils the so-called primary grasslands have survived continuously
81 in small pockets throughout the Holocene in their current locations (Lang, 1994; Poschlod
82 and WallisDeVries, 2002). Currently, the extent of these grassland types is limited, apart
83 from grasslands growing on salt and dry sandy soils, which are more common in the
84 Carpathian Basin (Molnár and Borhidi 2003; Deák et al., 2014). It is therefore highly
85 improbable that these small, isolated grassland areas were the only refugia of the
86 extraordinarily rich grassland flora of so-called "semi-natural" grasslands. The high genetic
87 diversity of some grassland plant species in this region (Turtureanu et al., 2014) and the
88 remarkable species richness and endemic plant and animal species typical for grasslands
89 (Chytrý et al., 2015) suggest a wider extent of primary grasslands during the Holocene.

90 In contrast to the primary grasslands, open canopy woodlands and grasslands currently
91 found in areas where climate and soils would allow forest growth, and which are only
92 extensively managed (i.e., no artificial fertiliser and pesticide application) are considered to
93 be semi-natural (Pärtel et al., 2005; Leuschner and Ellenberg, 2017). Here, we challenge the
94 view that most of these grasslands have replaced formerly naturally occurring forests within
95 recent centuries or millennia. We present multiple lines of evidence (palaeobotanical,
96 pedological, phylogenetic, palaeontological) from five countries in Central Eastern Europe, a
97 region with some of the highest-biodiversity grasslands of the world. Firstly, we review
98 published direct records of past grassland occurrence (pollen, plant macrofossils, charcoal)
99 from both natural (lakes, bogs) and archaeological archives, alongside other indirect fossil

100 (pedological, zoological) and recent genetic evidence from Central Eastern Europe.
101 Secondly, we review the characteristics of the environmental and disturbance factors
102 (climate, fire, herbivores and human impact) during the Holocene in this region. Finally, we
103 evaluate whether continuous grassland presence was possible under the Holocene climatic
104 conditions, fire and grazing regimes, and increasing anthropogenic impacts.
105 Based on their age and the drivers of their formation, we set out a new framework for three
106 types of grasslands in Central Eastern Europe: 1) Primary, natural grasslands on skeletal
107 and other poor soils, which have existed throughout the Holocene until the present; 2)
108 Primary, ancient grasslands on deeper soils, maintained by climate and disturbances during
109 the early Holocene and then predominantly by disturbances until the present; and 3) Semi-
110 natural grasslands, extensively managed grasslands, formed and maintained by
111 anthropogenic disturbances during the late Holocene (Table 1). We argue for the recognition
112 of the importance of previously overlooked ancient grasslands that have persisted throughout
113 the Holocene, maintained by natural and later also by anthropogenic disturbances.

114 **3. Multi-proxy evidence for grassland persistence during the Holocene**

115 **3.1 Fossil plant evidence**

116 Pollen and palaeobotanical records from natural archives (lakes and peatbogs) in currently
117 grassland-rich areas in Central Eastern Europe indicate the prevalence of a more open
118 landscape between 11,700 and 9000 cal yr BP and the maximum extent of forest cover
119 between 9000 and 4500 cal yr BP (Figs.1, 2; Table 2). While these studies show a reduction
120 in grassland cover, especially of xerothermic and floodplain grasslands during the mid
121 Holocene, there is no evidence of their widespread disappearance. Rather, there is strong
122 support for their persistence, given the concurrent presence of many grasslands and light-
123 demanding tree and shrub taxa i.e., *Artemisia* spp., *Thymus* spp., *Festuca rubra*, *Festuca*
124 spp., *Sanguisorba* spp., *Adonis* spp., *Helianthemum* spp., *Potentilla erecta*, *Potentilla* spp.,
125 *Centaurea* spp., *Filipendula* spp., *Trifolium* spp. during this period (Fig. 1; Table 2).
126 Archaeobotanical reports from Hungary and Poland suggest an even greater proportion of
127 heliophilous taxa growing locally than pollen records indicate (Fig. 1; Table 2). Remains of
128 grassland species including steppe elements i.e., *Stipa pennata*, *Stipa* sp., *Phleum pratense*,
129 *Teucrium chamaedrys*, *Asperula cynanchica*, *Plantago media*, *Silene vulgaris*, have been
130 reported from archaeobotanical records of early Neolithic sites from Germany (Fig. 1; Table
131 2) and Czech Republic (Archaeobotanical Database of the Czech Republic;
132 <http://www.arup.cas.cz>). Calcareous grasslands have also been identified in the Neolithic
133 lakeshore sites in the northern foothill of the Alps (Fig. 1; Table 2). The species found are
134 typical of closed xerophilous and mesophilous grasslands at sites where trees would have
135 been able to grow under the climatic conditions of the mid Holocene. Taken together, fossil
136 plant evidence from natural archives and archaeological sources suggest that grasslands
137 existed locally before the start of the Neolithic and therefore before marked human impacts.
138 These findings also demonstrate grassland persistence throughout the mid Holocene,
139 although archaeological records show the occurrence of higher grassland diversity than that
140 found in pollen records.

141 **3.2. Zoological evidence**

142 Indirect evidence for the persistence of open, or partly open landscapes throughout the mid
143 Holocene comes from the palaeontological remains of animal species restricted to extensive,
144 open habitats (Fig. 1). Results from the Carpathian Basin (Hungary and Romania) show that
145 several species typical of steppic environments e.g., *Ochotona pusilla* (Steppe Pika),
146 *Microtus gregalis* (Narrow Headed Vole), *Equus ferus* subsp. *gmelini* (Eastern European
147 Wild Horse), *Asinus hydruntinus* (European Wild Ass), *Otis tarda* (Great Bustard) and *Vipera*
148 *ursinii* subsp. *rakosiensis* (Meadow Adder) were abundant during the early Holocene
149 (Németh et al., 2017). They became discontinuously present from 8000 cal yr BP and
150
151

152 several of these species disappeared between 5000 and 4000 cal yr BP at a time of
153 increased anthropogenic pressure, but in a grassland landscape. Fossil malacological
154 records in currently grassland-rich landscapes reveal a similar picture (Fig. 1); a continuous
155 Holocene presence of strictly open habitat molluscs (*Chondrula tridens*, *Vallonia pulchella*,
156 and *Helicopsis striata*) unable to survive in closed forests (Ložek, 2005; Horsák et al., 2009;
157 Moskal-del Hoyo et al., 2018). Overall, while there is evidence that many open habitat
158 mollusc species contracted their range during mid Holocene forest expansion, there is also
159 robust support for their local long-term persistence, and consequently also for the continuity
160 of grassy ecosystems.

161

162 **3.3 Phylogeographic evidence**

163 Phylogeographic analysis represents a further source of data facilitating the interpretation of
164 the distribution of past grasslands. A pattern of genetic diversity decline from core
165 populations in southern Siberia towards the range periphery of smaller populations in
166 western Europe has been confirmed for several grassland plant species including *Adonis*
167 *vernalis* (Hirsch et al., 2015), *Stipa pennata* (Wagner et al., 2012), *Iris aphylla* (Wroblewska
168 2008) and *Stipa capillata* (Wagner et al., 2011). These studies have also revealed a
169 surprisingly low genetic differentiation between central and peripheral intermediate
170 populations, or a complete lack of private alleles among peripheral populations (e.g. Wagner
171 et al. 2011; Hirsch et al., 2015), which may reflect the absence of any long-standing isolation
172 of these populations. This implies that these species must have had a more continuous past
173 distribution in Central and Eastern Europe allowing gene flow and interbreeding. Rapid
174 progress in the field of DNA analysis, especially environmental DNA, may shed further light
175 on the origin and past range distribution of grasslands (Thomsen and Willerslev, 2015)

176

177 **3.4 Pedological evidence**

178 A further line of evidence used in the interpretation of former vegetation distributions comes
179 from soil types (IUSS WRB 2006). In Central Eastern Europe, dark soils (chernozems) from
180 steppe and forest steppe zones are considered to have developed before the spread of
181 forests and to have persisted under open or semi-open vegetation (Pokorný et al., 2015).
182 Others, however, view these soils as having survived under forest development (Eckmeier et
183 al., 2007). A palaeo-pedological analysis from the Transylvanian Basin (Romania) shows the
184 occurrence of islands of dark soils of Pleistocene age (20,000-14,000 cal yr BP), which, in
185 drier areas, persisted until the present (Pendea et al., 2002) suggesting grassland
186 persistence throughout the Holocene. However, in other areas of the Transylvanian Basin,
187 dark soils were overlain by Luvisols, typical of nemoral forests, about 5000 cal yr BP (Timar
188 et al., 2010) when wetter climatic conditions prevailed, whilst the current vegetation is
189 predominantly grassland. Grassland occurrence on soils typically favouring forests may be
190 explained by the prevalence of open woodlands throughout the Holocene, allowing the long-
191 term persistence of dark soil, rather than the post deforestation formation of this soil type.
192 Forest soils occur extremely rarely on chernozems developed on loess substrates in the
193 Hungarian Plain therefore suggesting the long-term existence of steppe grasslands (Máté
194 1957, Molnár 2010). Indeed, the continuous dominance of grasslands from the Late
195 Pleistocene on loess deposits in the southern Carpathian Basin has been demonstrated on
196 the basis of n-alkane biomarkers (Marković et al., 2018). Thus, pedological evidence from
197 chernozems, including those developed on loess, shows that chernozems existence under
198 open or semi-open vegetation.

199

200 **4. Drivers of grassland persistence during the Holocene**

201 **4.1 Climate conditions**

202 Proxy-based and climate simulations indicate warmer-than-present summer temperatures,

203 lower precipitation and soil moisture, and greater seasonality in the early Holocene (11700-
204 9000 cal yr BP) in Central and Eastern Europe at the time of maximum grassland extent
205 (Feurdean et al., 2013; 2014; Heiri et al., 2014). Palaeoclimatological records show a decline
206 in temperatures in this region from approximately 9000 to 4500 cal yr BP (Heiri et al., 2015;
207 Tóth et al., 2015; Hajkova et al., 2016). Climate simulations are consistent with this pattern of
208 mid Holocene cooling, but also marked precipitation and soil moisture increases in the mid
209 latitudes in Europe (Feurdean et al., 2013). Palaeoecological reconstructions show that
210 significant forest expansion occurred in response to cool and moist conditions in Central
211 Eastern Europe (Magyari et al., 2010; Feurdean et al., 2015; Kuneš et al., 2015; Novenko et
212 al., 2016; Pokorný et al. 2015; Jamrichová et al., 2017; Moskal-del Hoyo et al., 2018).
213 Therefore, both proxy and modelled palaeoclimatic and palaeoecological evidence from the
214 lowlands of Central Eastern Europe clearly show that grasslands were most extensive
215 (during the warm and dry climatic conditions, with prolonged droughts, of the early Holocene
216 (11,700-9000 cal yr BP) and became restricted under wetter conditions during the mid
217 Holocene (9000-4500 cal yr BP). This illustrates the stronger competitive advantage of
218 grasslands over trees when resources are limited i.e., lower moisture availability and
219 prolonged droughts.

220

221 **4.2 Fire**

222 The role of fire as one of the main drivers of the rise in grassy ecosystem during the Miocene
223 has been advocated (Osborne and Behling, 2006; Strömberg, 2011) and confirmed by
224 fossil records from C₄ dominated grasslands in Africa (Hoetzel et al., 2013). This is not
225 surprising as dominant grassland species have fine fuels with rapid curing and fast regrowth
226 rates as well as perennating buds near or below the soil surface; adaptations that ensure
227 regeneration after disturbances that damage the above-ground parts of the plant (He and
228 Lamont, 2018). Thus, whilst grass and herbs can withstand frequent fire, this shift in fire
229 regime may have harmed previously dominant tree species adapted to infrequent fire (He
230 and Lamont, 2018). Although, short-term, field-based burning experiments in Hungary found
231 conflicting results about the effect of fire on grassland biodiversity (Valkó et al., 2014, 2018b),
232 controlled laboratory and small-scale field experiments examining the effect of fire on seeds
233 found a predominantly negative effect of fire on seed germination in grassland species,
234 however, some positive effects (mainly from Fabaceae family) also emerged (Ruprecht et al.,
235 2013; 2015). Disturbances by fire have recently been considered essential for increased
236 grassland competition over trees during the Holocene in Central Eastern Europe (Magyari et
237 al., 2010; Feurdean et al., 2015). For example, a positive effect of frequent fires on the
238 competitive advantage of grasses over trees and, by this means, on the extent of grassland
239 in Transylvania, Romania, has been inferred from sedimentary charcoal particles and pollen
240 (Feurdean et al., 2013). On the contrary, the decline in fire frequency during the mid
241 Holocene has been shown to be detrimental for grassland extension (Fig. 3). Taken together,
242 neo- and palaeoecological evidence indicate that fire may have had a more important role in
243 the shifts between forest and grassland, and in grassland maintenance, than previously
244 thought. Exploring to what extent temperate grasslands are associated with frequent fires
245 and which grassland species/communities are most resilient or benefit mostly from fire could
246 be useful for the enhancement of management practices, i.e., preventing excessive
247 dominance by competitor grass species as well as the succession towards shrublands and
248 forests.

249

250 **4.3 Herbivores**

251 Large herbivorous mammals influence the physiology and growth of plants and are
252 considered ecological keystones in maintaining tree–grass coexistence (Crawley, 1983;
253 Sankaran et al., 2005). Released from megaherbivore pressure and with a change in climate
254 at the Pleistocene-Holocene boundary, European lowlands witnessed substantial forest

255 regeneration and a reduction in vegetation openness (Vera, 2000; Svenning, 2002).
256 Megaherbivore extinction may also have had a cascading effect on the population size and
257 diversity of small mammals dependent on vegetation openness and indirectly on the fire
258 regime (Gill et al., 2014). However, comparatively, little attention has been given to the effect
259 of the declining population size or extinction of wild herbivores, or the subsequent role of
260 livestock, on landscape structure during the mid to late Holocene. We know from
261 palaeontological and archaeozoological records in the Carpathian Basin that several large
262 herbivores i.e., *Bison bonasus* (European Bison), *Alces alces* (Eurasian Elk), *Equus ferus*
263 subsp. *gmelini* (Wild Horse), *Equus hemionus* (Asiatic Wild Ass) and *Dama dama* (Eurasian
264 Fallow Deer) became discontinuously present from the mid Holocene, i.e., 8000 cal yr BP
265 and that many became extinct by 4000 cal yr BP (Németh et al., 2017; Bejenaru et al., 2018).
266 In contrast, livestock numbers increased from 6500 cal yr BP (Schumacher et al., 2016).
267 Domestic livestock could prevent forest encroachment in the absence or with a low density of
268 wild herbivores. Domestic animals can act as dispersal vectors for plant propagules (Bruun
269 and Fritzboøger, 2002; Cosyns et al., 2005) and thereby increase the connectivity of
270 grassland patches in a heterogeneous landscape. Combined fossil records of fauna, pollen
271 and coprophilous fungi that reproduce exclusively on animal dung (*Sporormiella* spp.,
272 *Sordaria* spp., *Podospora* spp.) can provide means of assessing the effects of herbivores on
273 grassland dynamics (Fig. 3) and also the timing of the shift in influence from grazing by wild
274 herbivores to livestock grazing (Gill et al., 2009). Such records are still scant in Europe, but
275 the existing studies generally show the increasing effects of grazing by domestic livestock
276 from 5000 cal yr BP (Schumacher et al., 2016). Understanding the responses of grasslands
277 to different grazing animals (body size, grazing intensity and height, foraging strategy and
278 forage selectivity) will be essential in the development of future grassland management
279 strategies as various forms of livestock grazing have been proposed to simulate the effects
280 of grazing and browsing by wild herbivores (Poschlod and WallisDeVries, 2002; Bakker et
281 al., 2004; Tóth et al., 2016; Poschlod, 2017).

282

283 **4.4. Early human impact counteracted the encroachment of forest onto primary** 284 **grasslands**

285 As the increase in forest cover from 8000 cal yr BP coincided with the spread of Neolithic
286 culture across South Eastern Europe (Bogaard et al., 2004; Kreuz 2008), a critical question
287 in respect to grassland extent is whether anthropogenic impacts could have counteracted the
288 climate-driven development of a closed forest (Pokorný et al., 2015). Archaeological datasets
289 from this part of Europe indicate that Neolithic settlements tended to be established in open
290 landscapes and that field sizes were small (Moskal-del Hoyo et al., 2013; Chapman, 2017;
291 Marinova and Ntinou, 2017). As people first settled in naturally open landscapes, this
292 tendency could explain the apparent lack of major deforestation at this time in the pollen
293 records from Central Eastern Europe (Fig. 2). Fire activity was naturally high during the early
294 Holocene (Magyari et al., 2010; Feurdean et al., 2013), and humans may have taken
295 advantage of wildfires to extend their agro-pastoral activities into freshly burned habitats. In
296 agreement with Pokorný et al. (2015) we hypothesise that early anthropogenic land
297 management may have slowed, or partially arrested, the development of closed forest
298 favoured by the wetter climatic conditions of the mid Holocene at locations with low biomass
299 productivity, contributing to the maintenance of landscape openness.

300

301 **4.5. Semi-natural grasslands replacing forests: when and how?**

302 Individual pollen records, as well as large-scale quantitative vegetation reconstructions from
303 Central Eastern Europe, show that the level of anthropogenic impact on forest remained low
304 until about 6000-5000 cal yr BP (Magyari et al., 2010; Feurdean et al., 2015; Kuneš et al.,
305 2015; Jamrichová et al., 2017; Fig. 2). Modelled vegetation and land use (arable and pasture
306 cover) changes across Europe suggest that open areas expanded gradually from previously

307 cleared forest after ca. 6000 cal yr BP (Kaplan et al., 2017). A noticeable increase in the
308 abundance and richness of grassland along with the decline in total forest cover but increase
309 for *Quercus*, a tree taxon typical for woodland and woody pasture, in Central Eastern Europe
310 occurred from 4700-3500 cal yr BP onwards (Jamrichová et al., 2017). This demonstrates a
311 growing anthropogenic role in the extension of grasslands and the formation of open
312 woodlands. These grasslands belong to the so-called semi-natural grasslands that
313 developed from forests and are maintained by land management (Pärtel et al., 2007; Pereira
314 et al., 2017). Technological advances in agriculture and the expansion of urban centres and
315 farms from the Late Bronze Age and Iron Age (3500 cal yr BP) have led to both an extension
316 and intensification of the land use in Central Europe (Poschold, 2015; Rösch et al., 2016). It
317 is therefore not surprising that from this time onwards, the richness and extent of grassland
318 has been found to correlate closely with prehistoric settlement density and land management
319 (Poschold and WallisDeVries, 2002; Pärtel et al., 2005; Hajkova et al., 2011; Hejcman et al.,
320 2013; Poschold, 2017). Later on i.e., from the 15th to 20th centuries, grassland expansion is
321 strongly linked to sheep flock migration. Livestock acted as dispersal vectors and their
322 mobility may be one of the reasons that ancient and older semi-natural grasslands may have
323 similar species diversity (Poschold and WallisDeVries, 2002; Molnár et al. 2012; Poschold,
324 2017). The sowing of hayseed and mowing may also have promoted grassland expansion in
325 many parts of Europe (Babai and Molnár, 2014). In summary, semi-natural grasslands
326 expanded into formerly forested sites and have subsequently been maintained by a variety of
327 land management practices including grazing, burning and mowing.

328 **5. A new framework for Holocene grassland persistence; conservation consequences**

329 We provide a new framework distinguishing three types of biodiversity-rich grasslands in
330 Central Eastern Europe: small-scale primary grasslands on skeletal and other poor soils
331 (primary grasslands I); ancient grasslands that were not forested because of natural and
332 anthropogenic disturbances (primary grasslands II), and semi-natural grasslands developed
333 as a result of human activities replacing forests (Fig. 4; Table 1). We have identified the
334 reasons for the continuous presence of primary grasslands during the Holocene including
335 both natural i.e., climate conditions and soils (primary I), climate and disturbance factors e.g.,
336 fire and grazing (primary grasslands II). Neolithic people may have first settled in naturally
337 open areas such as grasslands or grassland-woodland mosaics arresting the development of
338 a full forest cover when the climate became wetter (mid Holocene), indirectly favouring the
339 preservation and expansion of grasslands. The intensification of human impact from 4700-
340 3500 cal yr BP onwards subsequently lead to considerable extension of semi-natural
341 grassland on formerly forested sites. A succession from grassland to forest after the
342 cessation of land management (e.g. grazing or burning) is not necessarily a proof against the
343 primary or ancient aspect of grasslands, but may indicate the lack of disturbances. Livestock
344 grazing in the late Holocene has replaced ancient grazing by megaherbivores prevailing until
345 the early Holocene and that by large herbivores throughout the mid Holocene.

346
347 Our findings also reveal misconceptions about the origin of Central Eastern European
348 grasslands and open canopy woodlands. The concept of a previously continuous, closed
349 forest in extant grassland-rich landscapes where climatic conditions are favourable for forest
350 fails to hold true, as grasslands are likely to have been continuously present throughout the
351 Holocene. The long-term persistence of grasslands at these locations is probably an
352 important reason for one of the highest small-scale species richness, many endemic,
353 worldwide in these habitats. Such species-rich plant communities can require millennia to
354 develop and only well-connected grassland patches can support genetically diverse plant
355 populations. These findings challenge the commonly held view that conservation activities
356 should primarily focus on the protection of forests in many areas of Central Eastern Europe.

357 Anthropogenic impacts tend to focus on forest clearance, but this perspective paper
358 highlights that conservationists and land managers need to carefully consider that, in many
359 cases, it is not primary forests that hold the highest biodiversity. Further, human-made, or
360 managed environments, such as extensively managed grasslands, are long-term landscape
361 features, contain unique plant and animal communities, and provide important ecosystems
362 services. Our findings support the recent wider acceptance of the notion that people and
363 nature should not be separated in the societal discourse of environmental science (Mace
364 2014). Finally, we advocate the need for a more detailed understanding of the role of
365 disturbances in grassland-forest dynamics, to avoid the overly simplistic assumption that
366 sparse tree cover is evidence of past deforestation. Fossil records provide such data and the
367 routine incorporation of palaeoecological investigations into environmental management is a
368 key step in developing science-based evidence for the conservation of the biodiversity of
369 grasslands. Thus, our regional case study supports the advocacy of Willis et al. (2010),
370 Barnosky et al. (2017) and Whitlock et al. (2018) for merging palaeobiology and conservation
371 biology as well as an appreciation of the dynamic history of species and ecosystems,
372 including the role of humans.
373

375 **Figures captions**

376

377 **Figure 1.** Location of the study area in Europe (A) and the distribution of the main vegetation
378 types in Central Eastern Europe based on the potential natural vegetation map of Europe (B;
379 Bohn et al., 2003). Colour symbols show location of various types of fossil records extracted
380 from literature (Table S1 and 2) indicating either continuous grassland presence throughout
381 the Holocene or during the afforestation phases of the mid Holocene (9000-4000 cal yr BP).
382 These fossil records reveal that grasslands were continuously present throughout the
383 Holocene in places where potential natural vegetation has been assumed to be forest.

384

385 **Figure 2.** Pollen based reconstruction of forest (green) versus open land cover (yellow) from
386 Central Eastern Europe during the Holocene using the pseudobiomisation method (Fyfe et
387 al., 2015). Cumulative land cover record was constructed by spatially aggregating 96 pollen
388 records extracted from the Pangaea Database and distributed across the region shown in
389 Fig.1 and Fig. S2. Forest cover includes both broadleaf and conifer trees, whereas open land
390 cover includes pastures/natural grasslands, and arable/disturbed land. Geological and
391 archaeological periods as well as the predominance of each grassland type throughout the
392 Holocene are also highlighted. Trends in simulated growing season temperature and
393 precipitation after Feurdean et al. (2015).

394

395 **Figure 3.** Reconstruction of tree, grassland and arable land cover from sedimentary pollen,
396 fire frequency from sedimentary charcoal and herbivore numbers based on sedimentary
397 coprophilous fungi from a grassland-rich landscape in Transylvania, Romania during the
398 Holocene (Feurdean et al., 2015). This multi-proxy approach illustrates the role of fire and
399 wild herbivores on vegetation openness during the early Holocene and the role of human
400 impact (deforestation, burning, livestock grazing) during the late Holocene. Note that the icons
401 only tentatively illustrate the evolution of grazing type, burning and cultivation during the
402 Holocene.

403

404 **Figure 4.** The effect of climate, soils and disturbances by fire, herbivores and humans on the
405 three types of grasslands and forest. Blue line denote a positive effect, red line a negative
406 effect and grey both effects.

407

408

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674 675 **Supplementary Material**

676
677 **Figure S11.** The World Resources Institute (WRI) Atlas of Forest and Landscape
678 Restoration, which identified areas currently occupied by open woody landscapes with
679 grasslands that could be suitable for various degrees of reforestation. Mosaic restoration
680 means that forests are embedded in a mosaic of land uses, whereas remote restoration

681 refers to the restoration of degraded forest in remote areas, mainly in the northern boreal
682 zone.

683

684

685 **Figure S2** Location of sites extracted from the European Pollen Database (EPD) and used to
686 construct Figure 2.

687 **Table S1.** Location of sites extracted from the literature and used to construct Figure 1.

688

690 **Table 1.** Species-rich grassland types

691 Grassland type	Characteristics
692 Primary grassland I 693	Natural grasslands on skeletal and other poor soils with moisture deficit
694 695 Primary grassland II 696 697	Ancient grasslands formed and maintained mainly by climate conditions and natural fires, herbivores and, later, also influenced by anthropogenic disturbances
698 699 Semi-natural grassland 700 701 702	Secondary grasslands formed and maintained by anthropogenic disturbances (deforestation, livestock grazing, cultivation, use of fire) in areas suitable for forests during the late Holocene

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706 Table 2. Fossil pollen, plant macrofossil and charcoal evidence for the presence of various types of grasslands during the early and mid Holocene. Note:
 707 on the basis of pollen analysis differentiation between primary grassland and semi-natural grassland is not always possible.
 708

709	Location	Taxa	Period	References
710	Romania	<i>Artemisia</i> , <i>Aster</i> -type, Caryophyllaceae undiff., Compositae Cichorioideae,	Holocene	Feurdean et al., 2015
711		<i>Centaurea</i> , Chenopodiaceae, <i>Helianthemum</i> , <i>Filipendula</i> , <i>Potentilla</i> , Poaceae,		
712		<i>Plantago lanceolata</i> , <i>Plantago major</i> / <i>P. media</i> , <i>Rumex acetosa</i> , <i>Thalictrum</i> ,		
713		<i>Chenopodium</i> , <i>Suaeda maritima</i> , <i>Festuca</i> , <i>Lycopus</i> , <i>Thymus</i> , Leguminosae undiff.		
714				
715				
716	Hungary	<i>Artemisia</i> , <i>Achillea</i> , <i>Ajuga</i> , <i>Adonis</i> , <i>Astragalus</i> , <i>Allium</i> , <i>Aster</i> -type, Poaceae,	Holocene	Magyari et al., 2010
717		<i>Centaurea</i> , <i>Filipendula</i> , <i>Festuca</i> , <i>Euphorbia</i> , Caryophyllaceae undiff.,		
718		<i>Gagea</i> , Chenopodiaceae, Compositae Cichorioideae, <i>Dianthus</i> -type,		
719		<i>Genista</i> , <i>Helianthemum</i> , <i>Hypericum</i> , <i>Inula</i> , Leguminosae undiff., <i>Lotus</i> -type,		
720		<i>Plantago lanceolata</i> , <i>Plantago major</i> / <i>P. media</i> , <i>Potentilla</i> , <i>Trifolium</i> ,		
721		<i>Rumex acetosa</i> , <i>Teucrium</i> , <i>Thymus</i> , <i>Thalictrum</i> , <i>Verbascum</i> , <i>Sanguisorba</i>		
722				
723				
724	Hungary	<i>Chenopodium album</i> , <i>Echinochloa crus-galli</i> , <i>Fallopia convolvulus</i>	Neolithic	Moskal del Hoyo et al.,
725		<i>Galium spurium</i> , <i>Melandrium album</i> , <i>Plantago lanceolata</i> , <i>Polycneum arvense</i> ,		2018
726		<i>Polygonum aviculare</i> , <i>Polygonum mite</i> , <i>Polygonum minus</i> , <i>Rumex acetosa</i> ,		
727		<i>Setaria pumila</i> , <i>Setaria viridis</i> , <i>Bromus</i> sp., <i>Chenopodium</i> sp., <i>Galium</i> ,		
728		<i>Ononis</i> sp., <i>Polygonum</i> sp., <i>Rumex</i> sp., <i>Stipa</i> sp., <i>Trifolium</i> sp., <i>Vicia</i> sp.,		
729		Fabaceae, Poaceae, Polygonaceae, Caryophyllaceae		
730				
731	Poland (S)	<i>Artemisia</i> , <i>Stipa</i> sp., <i>Knautia arvensis</i> , <i>Hypericum</i> , <i>Plantago media</i> , <i>P. lanceolata</i> ,	Mid Holocene	Moskal del Hoyo et al., 2018
732		<i>Filipendula</i> , <i>Aster</i> -type, Compositae Cichorioideae, Poaceae, Chenopodiaceae		
733				
734	Czech Republic	<i>Artemisia</i> , <i>Achillea</i> , <i>Ajuga</i> , <i>Astragalus</i> , <i>Filipendula</i> , Chenopodiaceae,	Holocene	Kuneš et al., 2015
735		<i>Centaurea</i> , Caryophyllaceae undiff., Compositae Cichorioideae, <i>Genista</i> , <i>Filipendula</i> ,		Jamrichova et al., 2017
736		<i>Helianthemum</i> , <i>Hypericum</i> , <i>Inula</i> -type, <i>Plantago lanceolata</i> , <i>P. media</i> , Poaceae,		Hajeck et al., 2016
737		<i>Potentilla</i> -type, <i>Ranunculus acris</i> -type, Rubiaceae, <i>Rumex acetosa</i> ,		
738		<i>Teucrium</i> , <i>Thymus</i> , <i>Thalictrum</i> , <i>Verbascum</i>		
739				

740	Austria	<i>Poa pratensis/trivialis, Festuca ovina/rubra, Brachypodium pinnatum,</i>	Neolithic	Körber-Grohne 1990
741		<i>Anthoxanthum odoratum, Stipa pennata</i>		
742				
743	Germany (S)	<i>Alchemilla vulgaris, Asperula cynanchica, Bupleurum falcatum, Carex muricata,</i>	Neolithic	Kreuz et al., 2005, 2008
744		<i>Centaurea, Daucus carota, Euphrasia, Phleum pratense, Poa annua, Polygonum aviculare,</i>		Kreuz and Schäfer, 2011
745		<i>Stipa sp., Stipa pennata, Rumex acetosella, Stellaria graminea, Trifolium campestre,</i>		
746		<i>Galium cf. verum, Galium molugo, Urtica dioica, Teucrium chamaedrys, Chenopodium spp.,</i>		
747		<i>Veronica arvensis, Plantago media</i>		
748				
749	Germany (NW)	<i>Achillea ptarmica, Angelica sylvestris, Anthriscus sylvestris, Artemisia cf. campestris,</i>	Preboreal & Boreal	Knörzer 1996
750		<i>Crepis biennis, Chaerophyllum hirsutum, Dianthus spp., Empetrum nigrum,</i>		
751		<i>Euphorbia cyparissias, Festuca cf. pratensis, Festuca rubra, Filipendula ulmaria, Galium mollugo,</i>		
752		<i>Heracleum sphondylium, Hypochaeris radicata, Juncus cf. effusus, Knautia arvensis,</i>		
753		<i>Lysimachia vulgaris, Laserpitium prutenicum, Poa pratensis / angustifolia,</i>		
754		<i>Polemonium coeruleum, Scabiosa columbaria, Plantago lanceolata,</i>		
755		<i>Potentilla tabernaemontani, Ranunculus acris, Rumex acetosa, Rumex tenuifolius,</i>		
756		<i>Sanguisorba officinalis, Scabiosa aff. columbaria, Silene vulgaris, Thalictrum flavum,</i>		
757		<i>Valeriana officinalis, Valeriana procurrens, Viola canina</i>		
758	Germany (NW)	<i>Achillea millefolium, Leontodon autumnalis, Agrostis tenuis, Melandrium rubrum,</i>	Neolithic	Knörzer 1996
759		<i>Phleum nodosum, Poa cf. pratensis, Artemisia cf. campestris, Poa cf. trivialis,</i>		
760		<i>Polygonum bistorta, Potentilla argentea, Carex ovalis, Prunella vulgaris,</i>		
761		<i>Centaurea cf. nigra, Ranunculus repens, Cerastium cf. semidecandrum, Rumex tenuifolius,</i>		
762		<i>Chrysanthemum leucanthemum, Deschampsia caespitosa, Selinum carvifolium,</i>		
763		<i>Festuca rubra, Silaum silaus, Heracleum sphondylium, Stachys recta, Hieracium pilosella,</i>		
764		<i>Trifolium arvense, Hypericum cf. maculatum, Trifolium dubium, Hypericum tetrapterum,</i>		
765		<i>Valerianella dentata, Juncus bufonius / conglomeratus, Veronica cf. arvensis,</i>		
766		<i>Trifolium repens, Juncus cf. effusus, Viola tricolor</i>		
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