

***Batrachochytrium dendrobatidis* in Hungary: an overview of recent and historical occurrence**

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2 **occurrence**

3
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24 **Abstract.** *Batrachochytrium dendrobatidis* (*Bd*) is a fungal pathogen which causes the
25 emerging infectious disease chytridiomycosis. *Bd* presents low host specificity and threatens
26 amphibians worldwide, thus systematic inventory is the key in order to detect and mitigate
27 the effects of the disease. Extensive data collection was conducted in Hungary in 2009-2015
28 from 14 different areas. Combined data – recent field sampling on 16 taxa and the
29 examination of archived *Bombina* spp. specimens – from 1360 individuals were analysed
30 with qPCR. Two sentinel taxa, *Bombina variegata* and the members of the *Pelophylax*
31 *esculentus* complex were marked to monitor the occurrence of *Bd* in two core areas (Bakony
32 Mts and Hortobágy National Park, respectively) of sampling. Climatic variables were also
33 examined in core areas to test their effect on prevalence and infection intensity. Among the
34 sixteen sampled amphibian taxa seven tested positive for *Bd* and the overall prevalence in
35 Hungary was 7.46%. Among the ethanol-fixed *Bombina* spp. individuals *Bd* was not
36 detected. In the first core area (Bakony Mts) the overall prevalence in *B. variegata* was
37 10.32% and juvenile individuals showed significantly higher prevalence than adults. On the
38 other hand there was a significant negative relationship between infection prevalence and
39 monthly mean air temperature. Finally, in the other core area (Hortobágy National Park) the
40 overall prevalence in *P. esculentus* complex was 13.00%, and no differences were found in
41 prevalence or infection intensity between sexes, sampling years or age classes.

42
43 **Key words.** chytridiomycosis, emerging infectious diseases, *Pelophylax esculentus* complex,
44 *Bombina variegata*, inventory, Central-Europe

45
46 Running title: Occurrence of *Bd* in Hungary

INTRODUCTION

47

48 Over the past decades several epidemics – caused by emerging infectious diseases –
49 resulted in the large-scale decline of numerous animal species globally (Dobson and
50 Foufopoulos, 2001). One such emerging disease is chytridiomycosis in amphibians caused by
51 the fungal pathogen *Batrachochytrium dendrobatidis* [hereafter, *Bd* (Longcore et al., 1999)].
52 *Bd* is a highly generalist, waterborne pathogen which is primarily transmitted through direct
53 contact with aquatic zoospores or infected individuals (Fisher et al., 2009). *Bd* is responsible
54 for population declines, mass mortalities and even extinction of species, and presents one of
55 the greatest threats to amphibians worldwide (Berger et al., 1998; Skerratt et al., 2007; Fisher
56 et al., 2009).

57 *Bd* is widespread on all continents where amphibians occur (Olson et al., 2013), but
58 the heaviest disease outbreaks were observed in the American Neotropics, Australia, North-
59 America and Western Europe (Fisher et al., 2009). In Europe, the first detection of *Bd* related
60 mass mortalities dates back to 1997 when the first recorded population decline as a result of
61 mass die-off after the emergence of chytridiomycosis was observed in Central Spain, in the
62 Guadarrama Mountain National Park, and targeted the Common midwife toad, *Alytes*
63 *obstetricans* (Bosch et al., 2001). Though, as a result of the increased attention in the
64 subsequent years, studies performed in the same region revealed that other species are highly
65 susceptible to the disease as well (e.g. *Salamandra salamandra*, *Bufo spinosus*; Bosch and
66 Martínez-Solano, 2006; Bosch et al., 2007). Moreover, the evidenced strong population
67 declines of *A. obstetricans*, *A. muletensis* and *A. dickhilleni* in the Iberian Peninsula (Bosch et
68 al., 2001; Walker et al., 2010; Bosch et al., 2013; Doddington et al., 2013; Rosa et al., 2013),

69 and the high susceptibility of these species made the midwife toads the “flagship” species of
70 European chytridiomycosis threat.

71 Central Europe harbours several amphibian species that might be susceptible to
72 chytridiomycosis, such as *S. salamandra*, *B. bufo*, *Bombina bombina* or *Bombina variegata*
73 (Baláž et al., 2014a,b). In the recent years *Bd* infection was detected in various areas of the
74 Czech Republic, as a result of a systematic inventory (Civiš et al., 2012). Furthermore, the
75 presence of the fungus was recently reported in low prevalence from Luxembourg (Wood et
76 al., 2009), Poland (Sura et al., 2010; Kolenda et al., 2017), Germany (Ohst et al., 2013),
77 Austria (Sztatecsny and Glaser, 2011), Slovakia (Baláž et al., 2014b) and Italy (Federici et
78 al., 2008; Tessa et al., 2013). New data indicates that the fungus is present also in the
79 Balkans, e.g. in Serbia (Mali et al., 2017), Albania, Montenegro and Macedonia (Vojar et al.,
80 2017). Though, interestingly, no negative effects or *Bd*-linked population declines have been
81 detected from Central-Eastern-Europe so far (Vörös et al., 2014).

82 Some aspects of chytridiomycosis epizootics show environmental correlates (Olson et
83 al., 2013). *Bd* presents a reasonably wide environmental tolerance under a variety of
84 temperature and precipitation regimes (Ron, 2005), but previous studies postulated that
85 climate (Berger et al., 2004; Bosch et al., 2007; Murray et al., 2009; Blaustein et al., 2010;
86 Rohr et al., 2010; Rödder et al., 2010) and elevation (Lips et al., 2008; Walker et al., 2010;
87 Becker and Zamudio, 2011) can significantly influence *Bd* outbreaks. Furthermore, large
88 intra- and interspecific variations exist, especially in the prevalence (Gründler et al., 2012;
89 Böll et al., 2014; Spitzen-Van Der Sluijs et al., 2014), but also in the intensity of infection
90 (Van Sluys and Hero, 2009; Baláž et al., 2014a; Spitzen-Van Der Sluijs et al., 2014). In
91 addition, behavioural differences influence the susceptibility to *Bd* which is further affected

92 by the intraspecific variability related to sex and life stage (Blaustein et al., 2005, Garcia et
93 al., 2006, Williams and Groves, 2014).

94 Hungary is situated in the Carpathian Basin, a region with high amphibian diversity
95 due to different climatic and zoogeographical influences (Vörös et al., 2014). Previous
96 findings about the occurrence of *Bd* in Hungary are restricted to a few areas and species
97 where the presence was initially detected (Gál et al., 2012; Baláz et al., 2014b, Vörös et al.,
98 2014, Drexler et al., 2017). Therefore, no large-scale distribution data on *Bd* presence is
99 available to date from the country.

100 Our study displays multiple goals. First, we present a general overview on the
101 occurrence of *Bd* in Hungary summarising data collected between the years 2009-2015. The
102 data set includes the general occurrence of *Bd* on sixteen amphibian taxa with a special focus
103 on the yellow-bellied toad *Bombina variegata* and water frogs belonging to the *Pelophylax*
104 *esculentus* complex. We selected these two target taxa because these species may present
105 high levels of infection intensity in Europe and so they may also act as sentinel taxa (Baláz et
106 al., 2014b); in addition, they can play a role in the spread and the persistence of the disease
107 (Baláz et al., 2014a).

108 Second, by studying *B. variegata* populations in Hungary we assessed whether
109 distinct phylogenetic lineages – Alpine (West of the Danube) and Carpathian, occurring in
110 the North Hungarian Range East of the Danube (Vörös et al., 2006) – express differences in
111 prevalence and infection intensity. Moreover, to explore the historical distribution of *Bd* in
112 Hungary field surveys were complemented with available archived samples of *Bombina* spp.
113 from museum collections which comprise a dataset covering a 70 years' time frame (1936-
114 2005) prior to our field sampling.

115 Third, in order to further monitor *Bd* infection levels of amphibians in Hungary, we
116 selected one population of two of the most susceptible taxa in Central-Eastern Europe, *B.*
117 *variegata* and the *P. esculentus* complex (Baláz et al., 2014b), and extensively sampled these
118 populations for three consecutive years in two core areas. Finally, we aimed to use climatic
119 data (monthly mean precipitation and monthly mean air temperature) in these core areas to
120 test if there is any correlation between the previously mentioned climatic variables and the
121 occurrence of *Bd*.

123 MATERIALS AND METHODS

124 *Data collection*

125 Altogether 1233 specimens belonging to sixteen amphibian taxa were studied in the
126 field between 2009-2015. Sampling was conducted in fourteen different regions in 45 distinct
127 sampling points throughout Hungary, covering a great variety of wetland habitats (i.e.
128 irrigation canals, streams, marshlands, ponds, fishponds, water reservoirs and temporary
129 wetland habitats) and elevations ranging between 84 and 734 m a.s.l. (Fig. 1, Table 1).
130 *Bombina variegata* was surveyed in five regions from Transdanubia (Region 1, 2, 3, 5 and 8
131 in Table 1 and Fig. 1) representing the Alpine (Western) genetic lineage, and in three regions
132 from the North Hungarian Mountains (Region 10, 12 and 13 in Table 1 and Fig. 1)
133 representing the Carpathian (Eastern) genetic lineage, covering the distribution of the species
134 in Hungary (Vörös et al., 2006). Identification of the two *Bombina* species and their hybrids
135 was performed considering morphological characters plus genetic information provided by
136 previous researches in Hungary (Vörös et al., 2006, 2007). Members of the *Pelophylax*
137 *esculentus* complex were sampled in eight regions (Region 1, 3, 4, 7, 8, 9, 10 and 14 in Table

138 1 and Fig. 1). Age classes were characterized as tadpoles, juveniles and adults based on the
139 external features of each species examined in the field. In those cases when we couldn't
140 distinguish between age and sex of an individual we discarded the sample for further
141 analysis. Additionally, 127 ethanol-fixed specimens of *Bombina* spp., deposited in the
142 Hungarian Natural History Museum (Budapest, Hungary) and Savaria Museum
143 (Szombathely, Hungary), collected between 1936 and 2005 from regions matching the
144 current distribution of the species were swabbed (Appendix 1).

145

146 *Systematic sampling of sentinel taxa in two core areas*

147 Core areas were selected based on the prevalence found previously or in the first year
148 of sampling (Gál et al., 2012; Baláz et al., 2014b). In Bakony Mts, *B. variegata* was
149 systematically sampled in 2010-2012. Data of 2010 were published previously (Gál et al.,
150 2012), thus our analyses includes a comparison of data from 2010 and new data from 2011
151 and 2012. Surveys were completed between March and September in 2010, April and
152 September in 2011, May and July in 2012. The assigned locality, Iharkút (see asterisk on Fig.
153 1), is an old open bauxite mine, where human activities are common due to being a famous
154 paleontological research site (Ősi et al., 2012). In Iharkút we were able to locate only two
155 water bodies: a small lake and a nearby stream. Because of the close proximity (ca. 50
156 meters) and the presumed connection of the two habitats, all the toads belonged to the same
157 population.

158 Members of the *P. esculentus* complex were screened for *Bd* in the Hortobágy
159 National Park (HNP; see asterisk on Fig. 1). HNP is the largest continuous alkaline steppe in
160 Europe covering 80.000 hectares. This natural reserve is abundant in wetland habitats like

161 alkaline marshes, fishponds, wet grasslands and wet meadows (Ecsedi, 2004). *Pelophylax*
162 species were sampled in three sites at HNP – Nádudvar-Kösély canal near the city Nádudvar,
163 a fish pond system located eastwards to Hortobágy and a marshland system at Egyek-
164 Pusztakócs – between April and October during three consecutive years (2012-2014).

165

166 *Taxonomic identification of Pelophylax esculentus complex*

167 Water frog taxon identification was determined using the technique described by
168 Hauswaldt et al. (2012), and is based on allele-size polymorphism in intron-1 of the serum
169 albumin gene (SAI-1; Plötner et al., 2009), with a slight modification in PCR protocol
170 (Herczeg et al., 2017). To verify SAI-1 fragments we sequenced representative alleles on a
171 Hitachi 3130 Genetic Analyzer (Applied Biosystems, UK). Consensus sequences were
172 compiled using BioEdit version 7.0.9.0 (Hall, 1999) and aligned manually. If genetic samples
173 were not available we referred to the individuals as *Pelophylax* sp.

174

175 *Sampling protocol*

176 We collected *Bd* samples following Hyatt et al. (2007) by either swabbing the skin of
177 the individuals or clipping one of the toes. According to Hyatt et al. (2007) skin swabbing
178 and toe clipping show similar performances in detectability of *Bd*. Skin swabbing was
179 performed using two types of sterile swabs (SWA90006; Biolab, Budapest, Hungary, 5 mm
180 diameter; and MW100-100; Medical Wire and Equipment, Wiltshire, England, 3 mm
181 diameter). We collected each sample in a standardized way with three strokes on each side of
182 the abdominal midline, the inner thighs, hands and feet. Toe clipping was performed using
183 sterilized scissors and toe clips were stored in 70% EtOH in a freezer at -80 °C. Skin swabs

184 were stored dry in individually labelled vials and transferred to a freezer for longer storage
185 throughout the field season. For both sampling procedures we used a new pair of disposable
186 gloves per individual, and after each sampling event we sterilized all the used sampling
187 equipment in order to avoid cross-contamination. Mouthpart (oral disc) of larvae were
188 swabbed following Hyatt et al. (2007). Ethanol-fixed specimens of *Bombina* spp. were
189 screened by skin swabbing following methodology presented above.

190

191 *Genetic analysis of Bd samples*

192 DNA was extracted using PrepMan Ultra Sample Preparation Reagent (Thermo
193 Fisher Scientific, Waltham, Massachusetts, USA) following the recommendations of Boyle
194 et al. (2004). Because of size differences between swabs (i.e. 3 mm vs. 5 mm; see above),
195 only the top 3 mm of the larger swabs was used in all cases. Extracted DNA was analysed
196 using real-time quantitative polymerase chain reaction (qPCR) following the amplification
197 methodology of Boyle et al. (2004) and Hyatt et al. (2007) targeting the partial ITS-1 – 5.8S
198 rRNA regions. Samples were run in triplicate and an internal positive control was included
199 (TaqMan exogenous internal positive control reagents; 4308323; Thermo Fisher Scientific,
200 Waltham, Massachusetts, USA) to detect potential inhibitors present in the DNA
201 extractions. We considered evidence of infection if genomic equivalents (GE) were ≥ 0.1 and
202 we considered a sample positive if all three wells returned a positive reaction. When a sample
203 returned an equivocal result, it was re-run. If it again returned an equivocal result, it was
204 considered negative (N = 17, 1.3% of total samples). The templates were run on a Rotor-
205 Gene 6000 real-time rotary analyser (Corbett Life Science, Sydney, Australia). GE were
206 estimated from standard curves based on positive controls of 100, 10, 1, 0.1 developed from

207 the *Bd* isolate IA 2011, from Acherito Lake, Spain. Finally, GE values of the three positive
208 replicates were averaged.

209 In order to identify lineages of *Bd* found on amphibians in Hungary, 2 µl of DNA
210 extract from three individuals (one juvenile *P. ridibundus* plus one juvenile *B. variegata* from
211 Bakony Mts, and one adult *B. variegata* from Órség) were selected as template for
212 amplification of a partial fragment of ITS1 rRNA. Nested PCR approach described by
213 Gaertner et al. (2009) was performed. The amplified fragments were sequenced on an
214 Applied Biosystems/Hitachi 3130 Genetic Analyser (Thermo Fisher Scientific, Waltham,
215 Massachussets, USA). Sequences were aligned manually using BioEdit version 7.0.9.0.
216 (Hall, 1999) and were blasted against available sequences from GenBank for identification.

217

218 *Climatic data*

219 Climatic data were provided by the Hungarian Meteorological Service (OMSZ). For
220 the core areas of *B. variegata* and *P. esculentus* complex climatic data were obtained from
221 the closest meteorological station of each sampling site: Pápa city (47.29, 17.37), 135.5 m
222 a.s.l, 21.5 km distance from Iharkút (Bakony Mts), and Kunmadaras village (47.46, 20.89),
223 88.8 m a.s.l. 12.5 km distance from Egyek-Pusztakócs (HNP), which is the closest sampling
224 point to the station. We used monthly mean precipitation and monthly mean air temperature
225 data for the period 2010-2014 to test if any relationship between climate and prevalence or
226 infection intensity exists.

227

228 *Statistical analyses*

229 Statistical analyses were performed in R (version 3.4.4; R Core Team, 2018).
230 Prevalence was expressed as a discrete binomial variable (uninfected vs. infected). Infection
231 intensity was expressed through GE value. First, we calculated infection prevalence (%) of
232 different amphibian species together with their 95% Clopper-Pearson confidence intervals
233 (95% CI) as follows. Prevalence values were obtained by dividing the cumulative number of
234 positive samples with the total number of samples per species and multiplied with 100 to
235 obtain percentile values, while 95% CI values were calculated using the R package 'PropCIs'
236 (function 'exactci'; Scherer, 2018). In *Bd* infected species we calculated the mean, median,
237 SD and range of GE values as well. The same statistics were run to compare the two
238 phylogenetic lineages of *B. variegata*, and in the two sentinel taxa (i.e. *B. variegata* and *P.*
239 *esculentus* complex) we also tested for differences between study years, sexes and age
240 classes. Prevalence values were compared with Chi-square tests, while infection intensities
241 were compared using Mood's median test, as implemented in the R package
242 'RVAideMemoire' (function 'mood medtest'; Hervé, 2018).

243 Finally, in the two sentinel taxa we tested the relationship between climatic variables
244 and prevalence and infection intensity. We note here that the data set of the *P. esculentus*
245 complex was restrained only on *P. ridibundus*, as the *Bd* infection of *P. esculentus* was very
246 low (i.e. two infected individuals in total) and the sample size of *P. lessonae* was also not
247 representative ($N = 1$). The relationship between the climatic factors and infection prevalence
248 was tested using generalized linear mixed models (GLMMs) with binomial error distribution
249 term and the relationship between the climatic factors and infection intensity was analysed
250 using linear mixed models with Gaussian distribution (LMMs). Prevalence and infection
251 intensity, respectively, were entered as dependent variables in the models, while the focal

252 climatic variable (i.e. air temperature or precipitation) was set as continuous predictor. In all
253 models sampling year was entered as a random effect to control for the interannual variations
254 in infection prevalence or intensity. Additionally, in the case of *P. ridibundus*, collection site
255 ID within the HNP was entered also as a random factor to account for the variations in
256 prevalence and intensity between collection sites. To assure the adequate distribution of
257 model residuals, for the LMMs GE values were $\log(x+1)$ -transformed. Prior entering into the
258 models, $\log(x+1)$ -transformed GE values and the continuous predictor were scaled to mean =
259 0 and SD = 1 to improve model convergence (see also Schielzeth 2010). Model fits were
260 checked visually by plot diagnosis. In all cases for the statistical comparison of infection
261 intensities only infected species/individuals were used. Mixed models were constructed using
262 the 'lme4' package for R (Bates et al., 2015), and P-values for the linear mixed models were
263 obtained using the function 'Anova' (type III) from the R package 'car' (Fox and Weisberg,
264 2011). We used a significance level of $P \leq 0.05$ throughout.

265

266 RESULTS

267 *Bd* occurrence in Hungary

268 In Hungary, nine regions were infected with *Bd* and the overall prevalence was 7.46%
269 (95% CI: 6.05–9.07), indicating a low presence of the fungus in the country (Table 1).
270 Among the sixteen sampled amphibian taxa seven were found infected with *Bd*, including
271 one unidentified *Pelophylax* individual (Table 2). Details on prevalence and summary
272 statistics of GE values are presented in Table 2; while the geographic distribution of the
273 sampling sites with the site-specific prevalence is shown in Fig. 1.

274

275 *Bd* occurrence in *Bombina variegata*

276 In *B. variegata* the overall prevalence was 12.69% (95% CI: 9.91–15.92). Details on
277 prevalence and summary statistics of GE values for the different regions are presented in
278 Table 3. We found no significant difference between the two lineages of *B. variegata* in
279 infection prevalence ($N_{\text{Alpine}} = 422$, $N_{\text{Carpathian}} = 82$; $\chi^2 = 0.155$, $df = 1$, $P = 0.693$) and intensity
280 ($N_{\text{Alpine}} = 52$, $N_{\text{Carpathian}} = 12$, $P = 0.750$). *Bd* was not detected among the ethanol-fixed *B.*
281 *variegata* specimens.

282 In Bakony Mts between 2010 and 2012 we sampled 310 individuals of *B. variegata*,
283 among which 32 individuals were found to be infected with *Bd*. Here the overall prevalence
284 was 10.32 % (95% CI: 7.16–14.25), and the mean, median, SD and range of GE values were
285 15.92, 5.09, 38.60 and 0.159–210.3, respectively. There was no significant difference in
286 infection prevalence ($N_{2010} = 80$, $N_{2011} = 144$, $N_{2012} = 86$; $\chi^2 = 4.980$, $df = 2$, $P = 0.082$) nor in
287 intensity between the three study years ($N_{2010} = 13$, $N_{2011} = 14$, $N_{2012} = 5$, $P = 0.201$), and we
288 found no significant difference in prevalence ($N_{\text{males}} = 113$, $N_{\text{females}} = 90$; $\chi^2 = 0.241$, $df = 1$, P
289 $= 0.623$) and infection intensity between sexes ($N_{\text{males}} = 8$, $N_{\text{females}} = 2$, $P = 0.545$). However,
290 there was a significant difference in prevalence between the two age classes ($N_{\text{juveniles}} = 105$,
291 $N_{\text{adults}} = 204$; $\chi^2 = 11.563$, $df = 1$, $P < 0.001$), with juveniles being more infected than adults
292 (proportion of individuals infected: 19.04% versus 5.88%). Differences in infection intensity
293 between the two age classes were not significant ($N_{\text{juveniles}} = 20$, $N_{\text{adults}} = 12$, $P = 0.273$). There
294 was significant negative relationship between infection prevalence and monthly mean air
295 temperature ($\chi^2 = 4.482$ $df = 1$, $P = 0.034$), and a marginally significant positive relationship
296 between prevalence and monthly mean precipitation ($\chi^2 = 3.611$, $df = 1$, $P = 0.057$). There
297 was no significant relationship between infection intensity and monthly mean air temperature

298 ($\chi^2 = 0.180$, $df = 1$, $P = 0.671$). However, there was a significant positive relationship between
299 infection intensity and monthly mean precipitation ($\chi^2 = 4.227$, $df = 1$, $P = 0.039$); though,
300 this significant relationship disappeared after removing one outlier GE value from the data
301 set ($\chi^2 = 1.510$, $df = 1$, $P = 0.219$).

302 All the three sequences (i.e. sequences obtained from juvenile *P. ridibundus* and *B.*
303 *variegata* from Bakony Mts, and one adult *B. variegata* from Örség) were identified as ITS1
304 rRNA of *Bd*, belonging to the globally dispersed *Bd*-GPL lineage (GenBank accession
305 numbers: MH745069-71). One sequence showed 100% identity with *Bd* from Cape Cod
306 (GenBank accession number: FQ176489.1, FQ176492.1), South Africa (JQ582903-4, 15,
307 37), and Italy (FJ010547). The second sequence was 100% identical with a sequence of *Bd*
308 from Equador (FJ232009.1), and the third sequence represented a unique haplotype. Genetic
309 distance (p-distance) among sequences ranged between 0.005–0.035.

310

311 *Bd* occurrence in *Pelophylax ridibundus*

312 In Hortobágy between 2012 and 2014 we sampled 100 individuals of *P. ridibundus*,
313 among which thirteen were found to be infected with *Bd*. Here the overall prevalence was
314 13.00% (7.10–21.20), and the mean, median, SD and range of GE values were 11.52, 1.59,
315 19.63 and 0.635–57.905, respectively. We found a significant difference in infection
316 prevalence between years ($N_{2012} = 35$, $N_{2013} = 48$, $N_{2014} = 17$; $\chi^2 = 27.750$, $df = 2$, $P < 0.001$);
317 all the infected individuals being captured in 2012 (prevalence: 37.14%), while no infected
318 individuals being found in 2013-2014. We found no significant difference in prevalence
319 ($N_{\text{males}} = 42$, $N_{\text{females}} = 30$; $\chi^2 = 0.002$, $df = 1$, $P = 0.958$) and infection intensity between sexes
320 ($N_{\text{males}} = 7$, $N_{\text{females}} = 6$, $P = 1.000$). Age classes did not differ in infection prevalence ($N_{\text{juveniles}}$

321 = 9, $N_{\text{adults}} = 72$; $\chi^2 = 0.827$, $df = 1$, $P = 0.363$). Infection intensities of the different age
322 classes cannot be compared because no infected juveniles were captured. We found no
323 significant relationship between infection prevalence and monthly mean air temperature ($\chi^2 =$
324 2.375 , $df = 1$, $P = 0.123$), and between prevalence and monthly mean precipitation ($\chi^2 =$
325 0.010 , $df = 1$, $P = 0.920$). Since infection prevalence was relatively low in the *P. esculentus*
326 complex and infected individuals were captured in the same month and year, the relationship
327 between climatic variables and infection intensity could not be tested in this taxa.

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329

DISCUSSION

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343

Low *Batrachochytrium dendrobatidis* prevalence was experienced throughout the country (Table 1, Table 2), with similar or slightly lower values than in neighbouring countries e.g. Czech Republic (Baláz et al., 2014a; 19% average at country level), Austria (Sztatecsny and Glaser, 2011; 5.9-45% at country level) or Poland (Kolenda et al., 2017; 18% average at country level). Overall, seven taxa carried the infection: *Bombina bombina*, *Bombina variegata*, *Bufo viridis*, *Pelophylax ridibundus*, *Pelophylax esculentus*, *Pelophylax* sp. and *Ichthyosaura alpestris*. In accordance with previous studies in Central Europe (Ohst et al., 2013; Baláz et al., 2014a,b; Kolenda et al., 2017), *B. variegata* and the members of the *P. esculentus* complex showed the highest prevalence and *Bd* infection intensity in Hungary. On the other hand, there was no difference in prevalence and infection intensity was detected between the two ancient phylogenetic lineages of *B. variegata*. *Bd* was present in eight of the fourteen studied regions. The highest prevalence was experienced in the Alpine foothills at Órség (Region 1), Soproni Mts (Region 2), and in the Zemplén Mts (Region 13). These three regions represent the margins of the Alps and Carpathians (respectively) hosting populations

344 with continuous distribution towards the higher regions. On the other hand, the remnant
345 mountain regions, where prevalence was much lower (Regions 3, 10 and 11), are
346 geographically isolated from other higher elevations. In contrast, amphibians from five
347 regions (Regions 5, 6, 8, 9 and 12) seemed to not carry *Bd*. This either indicates that *Bd* has
348 not reached these parts of the country yet, or more comprehensive sampling would be needed
349 to locate its presence.

350 The Carpathian Basin combines the characteristics of the neighbouring regions.
351 Despite the relatively small extent of Hungary, the climatic elements have distinct temporal
352 and spatial characters (Mezősi, 2017). Although the majority of the country has an elevation
353 of less than 300 m a.s.l., Hungary has several moderately high ranges of mountains and the
354 highest peak located in the Mátra Mts at 1014 m a.s.l. (Table 1, Region 10). Overall, our
355 results rather supporting the relationship between the measured climatic variables and
356 prevalence or infection intensity. We found significant relationship regarding *B. variegata*
357 individuals in the Bakony Mts core area, where prevalence was negatively affected by
358 monthly mean temperature. Furthermore, the monthly mean precipitation positively affected
359 the *Bd* infection intensity. Nonetheless, the robustness of the latter result is questionable,
360 since the relationship disappeared when we excluded an outlier value from the analysis. This
361 substantial effect of one outlier value could have on the outcomes of this analysis suggests
362 the need for an extensive sampling in order to test whether this result is a statistical artefact
363 or a real biological phenomenon.

364 To determine the time and location of the emergence or introduction of *Bd* in different
365 regions worldwide, it is important to study archived specimens deposited to museum
366 collections. To examine the historical presence of the fungus in Hungary we screened

367 archived specimens of *Bombina* spp. collected in the regions 1, 2, 3, 8, 10, 12, 13 and the
368 Kőszeg Mts (archived data only) between 1936 and 2005. In total 127 specimens were
369 analysed and all of the samples were *Bd* negative. Both for field and for museum samples we
370 used the same detection methodology, following Hyatt et al. (2007). The detection
371 probability with qPCR is more sensitive and accurate compared to conventional PCR or
372 histology (Annis et al., 2004; Boyle et al., 2004; Kriger et al., 2006). There is no difference in
373 regard of *Bd* detectability between sample collection techniques (i.e. skin swabbing, brushing
374 or scraping). Nonetheless, preservation methodology and storage history may have influence
375 on the results (Soto-Azat et al., 2009). The Amphibian Collection of the Hungarian Natural
376 History Museum is stored in ethanol, but no record is available about the mode of initial
377 preparation. As formaldehyde is known to inhibit PCR reaction, there is therefore a slight
378 chance that qPCR reactions failed to detect *Bd* in our archived samples; however, this may be
379 an unlikely possibility.

380 Although with testing archived specimens we did not find evidence on when *Bd* might
381 have been introduced into the country, our genetic analyses showed that the fungus found on
382 amphibians in Hungary is a member of the *Bd*-GPL lineage. This was confirmed by a recent
383 study tracking the origin of *Bd* using a full genome approach, which detected *Bd*-GPL
384 lineage in Hungary (from Iharkút, Bakony Mts; O’Hanlon et al., 2018) and is in line with
385 previous findings reporting that this lineage has a widespread distribution in Europe (Farrer
386 et al., 2007).

387 During the surveys in the core area of Bakony Mts (Region 3, Table 1) juvenile *B.*
388 *variegata* individuals showed a significantly higher prevalence compared to adults. The same
389 pattern was observed for two *B. variegata* populations in a seven-year period study in the

390 Netherlands, which the authors explained by the less developed immune responses, or
391 immunosuppression, following the stress of metamorphosis (Spitzen-van der Sluijs et al., 2017).
392 Quite surprisingly, during our study, two juveniles changed infection state once (recovered
393 from *Bd positive*). It is a relatively common phenomenon in the field, when infected adult
394 frogs lose and regain the infection which may be caused by overwintering tadpoles or larvae
395 acting as reservoirs (Briggs et al., 2010, Spitzen-van der Sluijs et al., 2017). In contrast, it is
396 less frequent with juvenile individuals as it was experienced in our study. Similar pattern was
397 observed for *Epidalea calamita* in Spain, where juveniles changed infection state towards the
398 end of metamorphosis, possibly mediated by the increasing water temperature in permanent
399 ponds (Bosch et al., pers. comm.).

400 In Iharkút (Bakony Mts), during our study period the environmental conditions
401 changed unexpectedly. The lake which hosted most of the amphibian species – including *B.*
402 *variegata* – dried out after the first season of sample collection. In the second year only four
403 individuals of *B. variegata* were captured around this locality, however the rest of the
404 specimens (N = 181) found shelter in a nearby stream unsuitable for breeding. During the
405 third year the lake kept dry and only seven out of 87 individuals were found in or around the
406 lake. Even though there was no difference in prevalence between the three years, they
407 showed a downward trend towards significance. Already low prevalence (23%) dropped
408 down to 11% in the second and to 5% in the third year. This trend could be associated with
409 the differences in habitat type, as it was observed for *Salamandra salamandra* in the
410 Guadarrama National Park, Spain (Medina et al, 2015). Here, *Bd* infection was greater in
411 salamander larvae from permanent ponds, while it was absent or weak in temporary water
412 bodies and permanent streams. Also, infection intensity in larval cohorts was reduced when

413 water was flowing rather than standing. Same authors suggested that increased water flow
414 rate reduce the likelihood of successful pathogen transmission.

415 Chytridiomycosis is limited to the keratinized tissues of the host individual, therefore
416 tadpoles and post-metamorphic amphibians are mostly affected by the disease (Rachowicz
417 and Vredenburg, 2004). Our dataset covered all life stages of amphibians and the presence of
418 the infection was not detected in tadpoles of *B. bufo* and *R. dalmatina* (N = 39). On the other
419 hand, post-metamorphic and juvenile individuals were found infected in the regions 1, 3, 10
420 and 13 of *B. variegata* and the members of the *P. esculentus* complex, even though all
421 sampled individuals apparently didn't display any clinical sign of chytridiomycosis.

422 In Central Europe the *P. esculentus* complex is formed by two sexual species, the *P.*
423 *ridibundus* and the *P. lessonae* and their interspecific mating produces the hybridogenetic *P.*
424 *esculentus*. Overall, our results in the core area of Hortobágy National Park showed higher
425 *Bd* prevalence in *P. ridibundus* compared to the hybrid *P. esculentus* (Table 2) which is
426 related to the fact that the hybrids have more effective peptide defence system against *Bd* and
427 have a richer peptide repertoire than both parental species (Daum et al., 2012). Further,
428 contrary to what was observed in *B. variegata* in the Bakony Mts core area, we did not find
429 differences in *Bd* infection between life stages and sexes in *P. ridibundus* individuals.

430 Our results fit into the general pattern showing significant variability in the effects of
431 chytridiomycosis across Europe. The marked difference in species susceptibility between
432 amphibian species/communities of Western and Central-Eastern Europe might be determined
433 by multiple linked factors, e.g. virulence of different *Bd* strains (Farrer et al., 2007), genotype
434 (Savage and Zamudio, 2011), behaviour (Williams and Groves, 2014), microbial skin
435 community compound of host species (Bletz et al., 2013), or structure of amphibian

436 communities (Becker et al., 2014). In the Iberian Peninsula – that received the most attention
437 due to mass amphibian mortalities caused by chytridiomycosis – infection was clustered
438 within high-altitude areas, where environmental conditions are the most optimal for growth
439 of *Bd* (Piotrowski et al., 2004). In contrast, Hungary harbours only low-elevation Mountains,
440 where environmental conditions might be less favourable for *Bd*-linked epidemics.
441 Differences in elevation might explain the relatively lower impact and infection values of
442 amphibians in Hungary, than it was reported for surrounding countries in Central and Eastern
443 Europe (e.g. Austria, Sztatecsny and Glaser, 2011; Czech Republic, Baláž et al., 2014a or
444 Poland, Kolenda et al., 2017).

445 Since *Bd*-related disease outbreak have been proven to be climate-driven (Bosch et
446 al., 2007), amphibians of Central-Eastern Europe might be heavily impacted in the future due
447 to global climate change. Changes in the climate might alter *Bd* diffusion and make it's
448 spreading less predictable, thus areas not yet affected by epidemics require particular
449 attention and constant monitoring.

450

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SUPPLEMENTARY MATERIAL

473 Supplementary material associated with this article can be found at <
474 <http://www.unipv.it/webshi/appendix> > Manuscript number 22611: Appendix 1.

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703 **Table 1.** Summary of regions, sampling locations, coordinates and sampled species in our inventory. mtDNA lineages were indicated as
 704 Alpine (Alp) or Carpathian (Carp) in the case of *B. variegata*. Lat = Latitude; Long = Longitude N = Number of individuals sampled; Prev
 705 = Prevalence; GE = Genomic equivalents; NA = not applicable

| Nr. of region | Alt | Lat | Long | Species | mtDNA lineage <i>B. variegata</i> | N | Positive/Sampled | Prev (%) | Prev 95% CI (%) | GE mean | GE median | GE SD | GE range |
|---------------|-------|-------|-------|-------------------------------|--------------------------------------|----|------------------|----------|-----------------|---------|-----------|-------|---------------|
| 1-Őrség | 315.0 | 46.87 | 16.13 | <i>Bombina variegata</i> | Alp | 2 | 16 / 68 | 23.53 | 14.09 – 35.38 | 34.45 | 5.01 | 58.32 | 0.20 – 182.78 |
| | 264.0 | 46.87 | 16.45 | <i>Bombina variegata</i> | Alp | 7 | | | | | | | |
| | 253.0 | 46.89 | 16.43 | <i>Hyla arborea</i> | | 1 | | | | | | | |
| | 253.0 | 46.89 | 16.43 | <i>Lissotriton vulgaris</i> | | 1 | | | | | | | |
| | 253.0 | 46.89 | 16.43 | <i>Rana arvalis</i> | | 1 | | | | | | | |
| | 253.0 | 46.89 | 16.43 | <i>Rana dalmatina</i> | | 4 | | | | | | | |
| | 315.0 | 46.90 | 16.24 | <i>Bombina variegata</i> | Alp | 48 | | | | | | | |
| | 315.0 | 46.90 | 16.24 | <i>Ichthyosaura alpestris</i> | | 1 | | | | | | | |
| | 267.0 | 46.91 | 16.23 | <i>Pelophylax esculentus</i> | | 1 | | | | | | | |
| | 315.0 | 46.90 | 16.24 | <i>Rana temporaria</i> | | 2 | | | | | | | |
| 2-Soproni | 493.0 | 47.65 | 16.48 | <i>Bombina</i> | Alp | 14 | 4 / 14 | 28.57 | 8.38 – | 2.05 | 2.40 | 1.13 | 0.48 – |

| | | | | | | | | | | | | | |
|----------|-------|-------|-------|---------------------|-----|-----|----------|-------|--------|-------|------|-------|--------|
| Mts | | | | <i>variegata</i> | | | | 58.10 | | | | 2.90 | |
| 3-Bakony | 455.0 | 47.06 | 17.67 | <i>Bombina</i> | | 2 | 37 / 606 | 6.11 | 4.33 – | 21.15 | 5.19 | 45.58 | 0.16 – |
| Mts | | | | <i>bombina</i> | | | | | 8.32 | | | | 210.30 |
| | 316.0 | 47.23 | 17.74 | <i>Bombina</i> | Alp | 3 | | | | | | | |
| | | | | <i>variegata</i> | | | | | | | | | |
| | 327.0 | 47.27 | 17.69 | <i>Bombina</i> | Alp | 15 | | | | | | | |
| | | | | <i>variegata</i> | | | | | | | | | |
| | 327.0 | 47.27 | 17.69 | <i>Bufo bufo</i> | | 2 | | | | | | | |
| | 327.0 | 47.27 | 17.69 | <i>Ichthyosaura</i> | | 12 | | | | | | | |
| | | | | <i>alpestris</i> | | | | | | | | | |
| | 327.0 | 47.27 | 17.69 | <i>Lissotriton</i> | | 19 | | | | | | | |
| | | | | <i>vulgaris</i> | | | | | | | | | |
| | 327.0 | 47.27 | 17.69 | <i>Rana</i> | | 25 | | | | | | | |
| | | | | <i>dalmatina</i> | | | | | | | | | |
| | 348.0 | 47.23 | 17.64 | <i>Bombina</i> | | 2 | | | | | | | |
| | | | | <i>bombina</i> | | | | | | | | | |
| | 348.0 | 47.23 | 17.64 | <i>Bombina</i> | Alp | 310 | | | | | | | |
| | | | | <i>variegata</i> | | | | | | | | | |
| | 356.0 | 47.23 | 17.65 | <i>Bufo bufo</i> | | 61 | | | | | | | |
| | 356.0 | 47.23 | 17.65 | <i>Bufo viridis</i> | | 39 | | | | | | | |
| | 348.0 | 47.23 | 17.64 | <i>Lissotriton</i> | | 5 | | | | | | | |
| | | | | <i>vulgaris</i> | | | | | | | | | |
| | 356.0 | 47.23 | 17.65 | <i>Pelophylax</i> | | 24 | | | | | | | |
| | | | | <i>ridibundus</i> | | | | | | | | | |
| | 348.0 | 47.23 | 17.64 | <i>Pelophylax</i> | | 4 | | | | | | | |
| | | | | sp. | | | | | | | | | |
| | 348.0 | 47.23 | 17.64 | <i>Rana</i> | | 83 | | | | | | | |
| | | | | <i>dalmatina</i> | | | | | | | | | |
| 4-Hanság | 113.0 | 47.66 | 16.74 | <i>Bombina</i> | | 4 | 3 / 33 | 9.09 | 1.92 – | 0.56 | 0.16 | 0.70 | 0.15 – |

| | | | | | | | | | | | | | |
|------------------------------|-------|-------|-------|----------------------------------|-----|----|--------|-------|-----------------|-------|-------|-------|-----------------|
| | | | | <i>bombina</i> | | | | 24.33 | | | | | 1.37 |
| | 116.0 | 47.63 | 17.08 | <i>Pelophylax ridibundus</i> | | 29 | | | | | | | |
| 5-Mecsek Mts | 381.0 | 46.22 | 18.33 | <i>Bombina variegata</i> | Alp | 12 | 0 / 23 | 0.00 | 0.00 – 14.82 | NA | NA | NA | NA |
| | 232.0 | 46.16 | 18.24 | <i>Bombina variegata</i> | Alp | 8 | | | | | | | |
| | 415.0 | 46.20 | 18.33 | <i>Bombina variegata</i> | Alp | 3 | | | | | | | |
| 6- Kiskunság | 89.0 | 46.61 | 19.12 | <i>Triturus dobrogicus</i> | | 13 | 0 / 13 | 0.00 | 0.00 – 24.71 | NA | NA | NA | NA |
| 7- Budapest | 100.0 | 47.18 | 18.53 | <i>Bombina bombina</i> | | 4 | 2 / 18 | 11.11 | 1.38 – 34.71 | 36.77 | 36.77 | 50.11 | 1.34 – 72.20 |
| | 111.0 | 47.42 | 19.14 | <i>Bufo viridis</i> | | 4 | | | | | | | |
| | 156.0 | 47.53 | 19.22 | <i>Pelophylax ridibundus</i> | | 10 | | | | | | | |
| 8-Pilis- Visegrádi Mts | 168.0 | 47.78 | 19.04 | <i>Bombina bombina</i> | | 1 | 0 / 78 | 0.00 | 0.00 – 4.62 | NA | NA | NA | NA |
| | 418.0 | 47.78 | 19.00 | <i>Rana dalmatina</i> | | 5 | | | | | | | |
| | 261.0 | 47.57 | 18.94 | <i>Bufo bufo</i> | | 1 | | | | | | | |
| | 261.0 | 47.57 | 18.94 | <i>Salamandra salamandra</i> | | 35 | | | | | | | |
| | 216.0 | 47.64 | 18.78 | <i>Bombina bombina</i> | | 2 | | | | | | | |
| | 329.0 | 47.76 | 18.85 | <i>Rana temporaria</i> | | 2 | | | | | | | |

| | | | | | | | | | | | | | |
|-----------------|-------|-------|-------|---|----|-----|--------|------|-------------|----|----|----|----|
| | 183.0 | 47.76 | 18.91 | <i>Salamandra salamandra</i> | 7 | | | | | | | | |
| | 234.0 | 47.61 | 18.88 | <i>Hyla arborea</i> | 1 | | | | | | | | |
| | 234.0 | 47.61 | 18.88 | <i>Pelophylax</i> sp. | 3 | | | | | | | | |
| | 208.0 | 47.85 | 19.12 | <i>Rana temporaria</i> | 1 | | | | | | | | |
| | 209.0 | 47.85 | 19.11 | <i>Salamandra salamandra</i> | 1 | | | | | | | | |
| | 107.0 | 47.77 | 19.09 | <i>Hyla arborea</i> | 2 | | | | | | | | |
| | 107.0 | 47.77 | 19.09 | <i>Pelophylax</i> sp. | 4 | | | | | | | | |
| | 358.0 | 47.72 | 19.06 | <i>Bombina bombina</i> x <i>variegata</i> | 1 | | | | | | | | |
| | 358.0 | 47.72 | 19.06 | <i>Bombina variegata</i> | 2 | Alp | | | | | | | |
| | 301.0 | 47.78 | 18.99 | <i>Pelophylax ridibundus</i> | 8 | | | | | | | | |
| | 301.0 | 47.78 | 18.99 | <i>Rana temporaria</i> | 2 | | | | | | | | |
| 9-Gödöllő Hills | 224.0 | 47.63 | 19.38 | <i>Lissotriton vulgaris</i> | 20 | | 0 / 56 | 0.00 | 0.00 – 6.38 | NA | NA | NA | NA |
| | 156.0 | 47.53 | 19.22 | <i>Pelophylax ridibundus</i> | 1 | | | | | | | | |
| | 111.0 | 47.76 | 17.34 | <i>Rana arvalis</i> | 1 | | | | | | | | |
| | 96.0 | 47.26 | 19.23 | <i>Rana arvalis</i> | 17 | | | | | | | | |

| | | | | | | | | | | | | | |
|---------------|-------|-------|-------|------------------------------------|------|----|---------|------|--------------|------|------|------|--------------|
| | 96.0 | 47.26 | 19.23 | <i>Rana dalmatina</i> | | 3 | | | | | | | |
| | 96.0 | 47.26 | 19.23 | <i>Triturus dobrogicus</i> | | 14 | | | | | | | |
| 10-Máttra Mts | 492.0 | 47.90 | 19.98 | <i>Bombina variegata</i> | Carp | 2 | 7 / 103 | 6.80 | 2.78 – 13.50 | 6.93 | 2.13 | 9.19 | 0.61 – 23.55 |
| | 648.0 | 47.93 | 19.89 | <i>Bombina variegata</i> | Carp | 2 | | | | | | | |
| | 648.0 | 47.93 | 19.89 | <i>Salamandra salamandra</i> | | 6 | | | | | | | |
| | 598.0 | 47.90 | 19.97 | <i>Bombina bombina</i> | | 2 | | | | | | | |
| | 587.0 | 47.85 | 19.96 | <i>Bombina variegata</i> | Carp | 3 | | | | | | | |
| | 316.0 | 47.97 | 19.52 | <i>Salamandra salamandra</i> | | 1 | | | | | | | |
| | 720.0 | 47.90 | 19.93 | <i>Bombina variegata</i> | Carp | 4 | | | | | | | |
| | 403.0 | 47.92 | 19.97 | <i>Bombina bombina</i> | | 2 | | | | | | | |
| | 304.0 | 47.93 | 19.98 | <i>Bombina bombina x variegata</i> | | 1 | | | | | | | |
| | 636.0 | 47.87 | 19.97 | <i>Bombina variegata</i> | Carp | 32 | | | | | | | |
| | 727.0 | 47.88 | 20.01 | <i>Bufo bufo</i> | | 1 | | | | | | | |
| | 727.0 | 47.88 | 20.01 | <i>Ichthyosaura alpestris</i> | | 11 | | | | | | | |

| | | | | | | | | | | | | | |
|----------------|-------|-------|-------|----------------------------------|------|----|-------|-------|-----------------|------|------|----|----|
| | 411.0 | 47.93 | 19.96 | <i>Pelophylax esculentus</i> | | 1 | | | | | | | |
| | 727.0 | 47.88 | 20.01 | <i>Rana temporaria</i> | | 1 | | | | | | | |
| | 727.0 | 47.88 | 20.01 | <i>Salamandra salamandra</i> | | 3 | | | | | | | |
| | 364.0 | 47.90 | 19.74 | <i>Bombina bombina</i> | | 3 | | | | | | | |
| | 362.0 | 47.93 | 19.76 | <i>Bombina variegata</i> | Carp | 1 | | | | | | | |
| | 522.0 | 47.89 | 20.10 | <i>Bombina bombina</i> | | 6 | | | | | | | |
| | 274.0 | 47.91 | 20.14 | <i>bombina x variegata</i> | | 1 | | | | | | | |
| | 633.0 | 47.89 | 20.11 | <i>Bombina variegata</i> | Carp | 12 | | | | | | | |
| | 636.0 | 47.93 | 19.93 | <i>bombina x variegata</i> | | 5 | | | | | | | |
| | 411.0 | 47.93 | 19.96 | <i>Bombina variegata</i> | Carp | 2 | | | | | | | |
| | 411.0 | 47.93 | 19.96 | <i>Pelophylax esculentus</i> | | 1 | | | | | | | |
| 11-Bükk Mts | 249.0 | 48.12 | 20.24 | <i>Bufo bufo</i> | | 1 | 1 / 9 | 11.11 | 0.28 – 48.25 | 8.10 | 8.10 | NA | NA |
| | 320.0 | 48.15 | 20.10 | <i>Rana temporaria</i> | | 1 | | | | | | | |

| | | | | | | | | | | | | | |
|--------------------------|-------|-------|-------|-------------------------------|------|----|----------|-------|------------------|--------|--------|--------|----------------------|
| | 443.0 | 48.04 | 20.56 | <i>Ichthyosaura alpestris</i> | | 6 | | | | | | | |
| | 330.0 | 48.15 | 20.08 | <i>Rana temporaria</i> | | 1 | | | | | | | |
| 12- Aggtelek Karst | 286.0 | 48.54 | 20.66 | <i>Bombina variegata</i> | Carp | 6 | 0 / 12 | 0.00 | 0.00 – 26.46 | NA | NA | NA | NA |
| | 238.0 | 48.53 | 20.64 | <i>Salamandra salamandra</i> | | 6 | | | | | | | |
| 13- Zemplén Mts | 468.0 | 48.27 | 21.29 | <i>Bombina variegata</i> | Carp | 10 | 6 / 22 | 27.27 | 10.73 – 50.22 | 244.00 | 101.15 | 328.43 | 13.03 – 882.54 |
| | 281.0 | 48.48 | 21.33 | <i>Bombina variegata</i> | Carp | 6 | | | | | | | |
| | 341.0 | 48.48 | 21.32 | <i>Rana temporaria</i> | | 1 | | | | | | | |
| | 341.0 | 48.48 | 21.32 | <i>Salamandra salamandra</i> | | 4 | | | | | | | |
| | 449.0 | 48.40 | 21.45 | <i>Bombina variegata</i> | Carp | 1 | | | | | | | |
| 14- Hortobágy | 86.0 | 47.57 | 20.94 | <i>Pelophylax esculentus</i> | | 18 | 16 / 178 | 8.99 | 5.23 – 14.19 | 10.48 | 1.48 | 17.98 | 0.64 – 57.91 |
| | 84.0 | 47.60 | 20.88 | <i>Pelophylax lessonae</i> | | 1 | | | | | | | |
| | 86.0 | 47.57 | 20.94 | <i>Pelophylax ridibundus</i> | | 2 | | | | | | | |
| | 85.0 | 47.62 | 21.08 | <i>Pelophylax esculentus</i> | | 25 | | | | | | | |

| | | | | |
|------|-------|-------|----------------------------------|----|
| 86.0 | 47.61 | 21.07 | <i>Pelophylax ridibundus</i> | 56 |
| 86.0 | 47.63 | 21.08 | <i>Pelophylax sp.</i> | 12 |
| 85.0 | 47.44 | 21.14 | <i>Pelophylax esculentus</i> | 20 |
| 85.0 | 47.44 | 21.14 | <i>Pelophylax ridibundus</i> | 42 |
| 84.0 | 47.45 | 21.17 | <i>Pelophylax sp.</i> | 2 |

Total 1233

706

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707 **Table 2.** *Batrachochytrium dendrobatidis* (*Bd*) infection in amphibian species sampled in Hungary between the years 2009 and 2015. Prev =
 708 prevalence; GE = genomic equivalents of zoospores; NA = not applicable

| Species | Positive/Sampled | Prev (%) | Prev 95% CI (%) | GE mean | GE median | GE SD | GE range |
|---|------------------|----------|-----------------|---------|-----------|--------|---------------|
| Order Anura | | | | | | | |
| Family Bombinatoridae | | | | | | | |
| <i>Bombina bombina</i> | 1 / 29 | 3.45 | 0.09 – 17.76 | 16.41 | 16.41 | NA | NA |
| <i>Bombina variegata</i> | 64 / 504 | 12.70 | 9.92 – 15.92 | 40.08 | 4.96 | 120.76 | 0.16 – 882.54 |
| <i>Bombina bombina</i> x <i>variegata</i> | 0 / 8 | NA | 0.00 – 36.94 | NA | NA | NA | NA |
| Family Bufonidae | | | | | | | |
| <i>Bufo bufo</i> | 0 / 66 | NA | 0.00 – 5.44 | NA | NA | NA | NA |
| <i>Bufo viridis</i> | 2 / 43 | 4.65 | 0.57 – 15.81 | 36.77 | 36.77 | 50.11 | 1.34 – 72.20 |
| Family Hylidae | | | | | | | |
| <i>Hyla arborea</i> | 0 / 4 | NA | 0.00 – 60.24 | NA | NA | NA | NA |
| Family Ranidae | | | | | | | |
| <i>Pelophylax esculentus</i> | 2 / 66 | 3.03 | 0.37 – 10.52 | 1.07 | 1.07 | 0.41 | 0.78 – 1.36 |
| <i>Pelophylax lessonae</i> | 0 / 1 | NA | 0.00 – 97.5 | NA | NA | NA | NA |
| <i>Pelophylax ridibundus</i> | 21 / 164 | 12.80 | 8.10 – 18.91 | 20.21 | 1.59 | 41.72 | 0.15 – 164.30 |
| <i>Pelophylax</i> sp. | 1 / 33 | 3.03 | 0.08 – 15.76 | 15.75 | 15.75 | NA | NA |
| <i>Rana dalmatina</i> | 0 / 120 | NA | 0.00 – 3.03 | NA | NA | NA | NA |
| <i>Rana arvalis</i> | 0 / 19 | NA | 0.00 – 17.65 | NA | NA | NA | NA |
| <i>Rana temporaria</i> | 0 / 11 | NA | 0.00 – 28.49 | NA | NA | NA | NA |
| Order Caudata | | | | | | | |
| Family Salamandridae | | | | | | | |
| <i>Salamandra salamandra</i> | 0 / 63 | NA | 0.00 – 5.69 | NA | NA | NA | NA |
| <i>Triturus dobrogicus</i> | 0 / 27 | NA | 0.00 – 12.77 | NA | NA | NA | NA |

| | | | | | | | |
|-------------------------------|-----------|------|--------------|------|------|----|----|
| <i>Lissotriton vulgaris</i> | 0 / 45 | NA | 0.00 – 7.87 | NA | NA | NA | NA |
| <i>Ichthyosaura alpestris</i> | 1 / 30 | 3.33 | 0.08 – 17.22 | 8.10 | 8.10 | NA | NA |
| Total | 92 / 1233 | 7.46 | 6.05 – 9.07 | | | | |

accepted manuscript

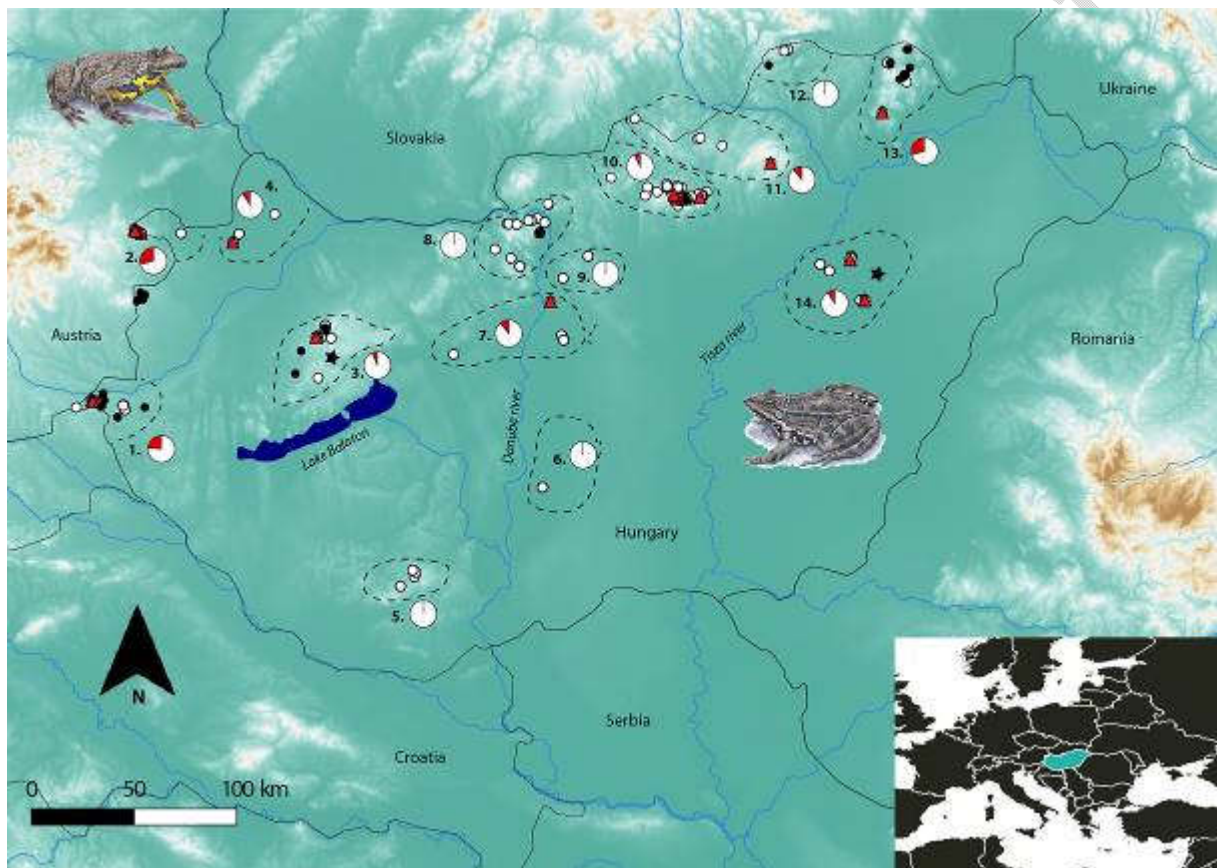
709 **Table 3.** *Batrachochytrium dendrobatidis* (*Bd*) detection in regions representing the surveyed local populations of *B. variegata* in Hungary. GE =
 710 genomic equivalents of zoospores; NA = not applicable

| Genetic lineage | Region | Positive/Sampled | Prev (%) | Prev 95% CI (%) | GE mean | GE median | GE SD | GE range |
|------------------------|---------------------|-------------------------|-----------------|------------------------|----------------|------------------|--------------|-----------------|
| Alpine | Őrség | 16 / 57 | 28.07 | 16.97 – 41.54 | 34.45 | 5.01 | 58.32 | 0.20 – 182.78 |
| | Soproni Mts | 4 / 14 | 28.57 | 8.39 – 58.10 | 2.05 | 2.40 | 1.13 | 0.48 – 2.90 |
| | Bakony Mts | 32 / 328 | 9.76 | 6.77 – 13.49 | 15.93 | 5.09 | 38.61 | 0.16 – 210.30 |
| | Mecsek Mts | 0 / 23 | 0.00 | 0.00 – 14.82 | 0.00 | 0.00 | 0.00 | NA |
| | Pilis-Visegrádi Mts | 0 / 2 | 0.00 | 0.00 – 84.19 | 0.00 | 0.00 | 0.00 | NA |
| Carpathian | Mátra Mts | 6 / 58 | 10.34 | 3.89 – 21.17 | 5.36 | 1.86 | 8.97 | 0.61 – 23.55 |
| | Aggtelek Karst | 0 / 6 | 0.00 | 0.00 – 45.93 | 0.00 | 0.00 | 0.00 | NA |
| | Zemplén Mts | 6 / 16 | 37.50 | 15.20 – 64.57 | 244.00 | 101.15 | 328.43 | 13.03 – 882.54 |
| Total | | 64 / 508 | 12.59 | 9.83 – 15.80 | | | | |

711

712 **Fig.1.** Map of Hungary showing sampling locations of *Bd* negative (black filled circles), *Bd*
713 positive (red/grey triangles) and archived (white circles) samples. Pie charts indicate *Bd*
714 prevalence of the 14 studied geographic regions. Numbers of regions correspond to Table 1.
715 The two core areas are marked with asterisk (Region 3 and 14). Drawing of *Bombina*
716 *variegata* and *Pelophylax ridibundus* courtesy of Márton Zsoldos.

717



718

719

Appendix 1 List of archived and analysed samples of *Bombina* spp. in this study

| Region | Locality | Code | Species | Date of collection | Catalogue number | WGSX | WGSY |
|----------------|-------------------------------|------|---|--------------------|-----------------------|-------|-------|
| Aggtelek karts | Aggtelek, Vörös lake | MA1 | <i>B. bombina</i> x <i>B. variegata</i> | 1990.04.19 | HNHMHER 2002.680.1 | 48.47 | 20.54 |
| Aggtelek karts | Aggtelek, Vörös lake | MA2 | <i>B. bombina</i> x <i>B. variegata</i> | 1990.04.19 | HNHMHER 2002.680.2 | 48.47 | 20.54 |
| Aggtelek karts | Aggtelek, Vörös lake | MA3 | <i>B. bombina</i> x <i>B. variegata</i> | 1990.04.19 | HNHMHER 2002.680.3 | 48.47 | 20.54 |
| Aggtelek karts | Aggtelek, Vörös lake | MA4 | <i>B. bombina</i> x <i>B. variegata</i> | 1990.04.19 | HNHMHER 90.18.1 | 48.47 | 20.54 |
| Bakony Mts | Bakonybél, Vörös János stream | MB1 | <i>B. bombina</i> | 1959.05.20-21. | HNHMHER 2002.419.1 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB2 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.609.1 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB3 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.609.2 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB4 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.609.3 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB5 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.609.4 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB6 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.609.5 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB7 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.610.1 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB8 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 2002.610.2 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB13 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 60.27.1 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB14 | <i>B. bombina</i> | 1959.05.20-21. | HNHMHER 60.28.1 | 47.27 | 17.70 |
| Bakony Mts | Bakonybél, Vörös János stream | MB17 | <i>B. variegata</i> | 1959.05.20-21. | HNHMHER 76.166.1 | 47.27 | 17.70 |
| Bakony Mts | Németbánya | MB9 | <i>B. variegata</i> | 1964.06.12-13. | HNHMHER 2002.615.1 | 47.17 | 17.55 |
| Bakony Mts | Németbánya | MB10 | <i>B. bombina</i> x <i>B. variegata</i> | 1964.06.12-13. | HNHMHER 2002.678.1 | 47.17 | 17.55 |
| Bakony Mts | Németbánya | MB11 | <i>B. bombina</i> x <i>B. variegata</i> | 1964.06.12-13. | HNHMHER 2002.678.2 | 47.17 | 17.55 |
| Bakony Mts | Németbánya | MB12 | <i>B. bombina</i> x <i>B. variegata</i> | 1964.06.12-13. | HNHMHER 2002.678.3 | 47.17 | 17.55 |

| | | | | | | | |
|-------------|------------------------------|------|----------------------------------|----------------|------------------------------|-------|-------|
| Bakony Mts | Németbánya | MB15 | <i>B. bombina x B. variegata</i> | 1964.06.12-13. | HNHMHER 64.47.1 | 47.17 | 17.55 |
| Bakony Mts | Németbánya | MB16 | <i>B. variegata</i> | 1964.06.12-13. | HNHMHER 64.49.1 | 47.17 | 17.55 |
| Kőszegi Mts | Cák, rock mine | SM7 | <i>B. variegata</i> | 1976.07.05 | SAMU 87.150.1.4. | 47.36 | 16.52 |
| Kőszegi Mts | Cák, rock mine | SM8 | <i>B. variegata</i> | 1976.07.05 | SAMU 87.150.1.4. | 47.36 | 16.52 |
| Kőszegi Mts | Kőszeg | SM6 | <i>B. variegata</i> | 1936.07.15 | SAMU 2002.36.1 | 47.38 | 16.53 |
| Kőszegi Mts | Kőszeg | SM11 | <i>B. variegata</i> | 1981.07.25 | SAMU 87.75.1.2. | 47.36 | 16.49 |
| Kőszegi Mts | Kőszeg | SM12 | <i>B. variegata</i> | 1981.07.25 | SAMU 87.75.1.2. | 47.36 | 16.49 |
| Mátra Mts | Padrag | MM11 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.1 HNHMHER | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM12 | <i>B. bombina</i> | 1957.06.20 | 2006.49.10 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM13 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.2 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM14 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.3 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM15 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.4 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM16 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.5 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM17 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.6 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM18 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.7 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM19 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.8 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM20 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 2006.49.9 | 47.07 | 17.52 |
| Mátra Mts | Padrag | MM21 | <i>B. bombina</i> | 1957.06.20 | HNHMHER 57.241.1 HNHMHER | 47.07 | 17.52 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM1 | <i>B. variegata</i> | 1969.07.07 | 2002.618.1 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM2 | <i>B. variegata</i> | 1969.07.07 | 2002.618.2 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM3 | <i>B. variegata</i> | 1969.07.07 | 2002.618.3 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM4 | <i>B. bombina x B. variegata</i> | 1969.07.07 | 2002.677.1 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM5 | <i>B. bombina x B. variegata</i> | 1969.07.07 | 2002.677.2 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM6 | <i>B. bombina x B. variegata</i> | 1969.07.07 | 2002.677.3 HNHMHER | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM7 | <i>B. bombina x B. variegata</i> | 1967.05.12 | 2002.682.1 | 47.88 | 20.01 |

| | | | | | | | |
|-----------|-------------------------------|------|----------------------------------|----------------|-----------------------|-------|-------|
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM8 | <i>B. bombina x B. variegata</i> | 1967.05.12 | HNHMHER 2002.682.2 | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM9 | <i>B. bombina x B. variegata</i> | 1967.05.12 | HNHMHER 2002.682.3 | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM10 | <i>B. bombina x B. variegata</i> | 1967.05.12 | HNHMHER 2002.682.4 | 47.88 | 20.01 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM22 | <i>B. bombina x B. variegata</i> | 1967.05.12 | HNHMHER 67.18.1 | 48.54 | 21.45 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM23 | <i>B. variegata</i> | 1969.07.07 | HNHMHER 69.8.1 | 48.54 | 21.45 |
| Mátra Mts | Parádfürdő, Pisztrángos-lake | MM24 | <i>B. bombina x B. variegata</i> | 1969.07.07 | HNHMHER 69.9.1 | 48.43 | 21.43 |
| Órség | Cák | SM13 | <i>B. variegata</i> | 1976.04.05 | SAMU 87.143.1.4. | 47.36 | 16.51 |
| Órség | Cák | SM14 | <i>B. variegata</i> | 1976.04.05 | SAMU 87.143.1.4. | 47.36 | 16.51 |
| Órség | Cák | SM15 | <i>B. variegata</i> | 1976.04.05 | SAMU 87.143.1.4. | 47.36 | 16.51 |
| Órség | Cák | SM16 | <i>B. variegata</i> | 1976.04.05 | SAMU 87.143.1.4. | 47.36 | 16.51 |
| Órség | Farkasfa | SM4 | <i>B. variegata</i> | 1983.05.27 | SAMU 87.158.1.1. | 46.91 | 16.31 |
| Órség | Orfalu | SM3 | <i>B. variegata</i> | 1977.07.13 | SAMU 87.156.1.2. | 46.88 | 16.29 |
| Órség | Óriszentpéter, Disznós stream | MO1 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 2002.614.1 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO2 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 2002.614.2 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO3 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 2002.614.3 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO4 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 2002.614.4 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO5 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 2002.614.5 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO6 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 2002.644.1 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO7 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 2002.644.2 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO8 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 2002.644.3 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO9 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 2002.644.4 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO10 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 2002.644.5 | 46.84 | 16.40 |
| Órség | Óriszentpéter, Disznós stream | MO11 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER | 46.84 | 16.40 |

| | | | | | | | |
|---------------|-------------------------------|------|----------------------------------|----------------|------------------|-------|-------|
| | | | | | 2002.644.6 | | |
| | | | | | HNHMHER | | |
| Őrség | Őriszentpéter, Disznós stream | MO12 | <i>B. bombina</i> | 1970.08.05-08. | 2002.644.7 | 46.84 | 16.40 |
| Őrség | Őriszentpéter, Disznós stream | MO17 | <i>B. bombina</i> | 1970.08.05-08. | HNHMHER 70.88.1 | 46.84 | 16.40 |
| Őrség | Őriszentpéter, Disznós stream | MO18 | <i>B. variegata</i> | 1970.08.05-08. | HNHMHER 70.89.1 | 46.84 | 16.40 |
| Őrség | Sopron | SM5 | <i>B. variegata</i> | 1971.06.10 | SAMU 87.147.1.1. | 47.39 | 16.50 |
| Őrség | Szentgotthárd | SM9 | <i>B. variegata</i> | 1977.07.13 | SAMU 87.205.1.2. | 46.94 | 16.30 |
| Őrség | Szentgotthárd | SM10 | <i>B. variegata</i> | 1977.07.13 | SAMU 87.205.1.2. | 46.94 | 16.30 |
| | | | | | HNHMHER | | |
| Őrség | Szőce | MO13 | <i>B. bombina x B. variegata</i> | 1990.08.13 | 2002.673.1 | 46.89 | 16.57 |
| | | | | | HNHMHER | | |
| Őrség | Szőce | MO14 | <i>B. bombina x B. variegata</i> | 1990.08.13 | 2002.673.2 | 46.89 | 16.57 |
| | | | | | HNHMHER | | |
| Őrség | Szőce | MO15 | <i>B. bombina x B. variegata</i> | 1990.08.13 | 2002.673.3 | 46.89 | 16.57 |
| | | | | | HNHMHER | | |
| Őrség | Szőce | MO16 | <i>B. bombina x B. variegata</i> | 1990.08.13 | 2002.673.4 | 46.89 | 16.57 |
| Őrség | Szőce | MO19 | <i>B. bombina x B. variegata</i> | 1990.08.13 | HNHMHER 90.53.1 | 46.89 | 16.57 |
| Őrség | Velem | SM17 | <i>B. variegata</i> | 1982.05.13 | SAMU 87.124.1.3 | 47.35 | 16.48 |
| Őrség | Velem | SM18 | <i>B. variegata</i> | 1982.05.13 | SAMU 87.124.1.3 | 47.35 | 16.48 |
| Őrség | Velem | SM19 | <i>B. variegata</i> | 1982.05.13 | SAMU 87.124.1.3 | 47.35 | 16.48 |
| Visegrádi Mts | Leányfalu, Csíkos lake | MP8 | <i>B. bombina</i> | 1952.05.20 | HNHMHER 57.16.1 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Rekettyés lake | MP1 | <i>B. bombina</i> | 1960.04.04 | 2002.554.1 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Rekettyés lake | MP2 | <i>B. bombina</i> | 1960.04.04 | 2002.554.2 | 47.72 | 19.06 |
| Visegrádi Mts | Leányfalu, Rekettyés lake | MP9 | <i>B. bombina</i> | 1960.04.04 | HNHMHER 60.77.1 | 47.72 | 19.06 |
| Visegrádi Mts | Leányfalu, Rekettyés lake | MP10 | <i>B. bombina</i> | 1960.04.04 | HNHMHER 60.91.1 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Sztradovavally | MP4 | <i>B. variegata</i> | 2005.06.09 | 2006.136.1 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Sztradovavally | MP5 | <i>B. variegata</i> | 2005.06.09 | 2006.136.2 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Sztradovavally | MP6 | <i>B. variegata</i> | 2005.06.01 | 2007.109.1 | 47.72 | 19.06 |
| | | | | | HNHMHER | | |
| Visegrádi Mts | Leányfalu, Sztradovavally | MP7 | <i>B. variegata</i> | 2005.06.01 | 2007.109.2 | 47.72 | 19.06 |

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|-------------|--------------------------------------|------|---------------------|----------------|-----------------------------|-------|-------|
| Zemplén Mts | Füzér | MZ36 | <i>B. variegata</i> | 1977.04.03-08. | HNHMHER 76.138.1 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér | MZ5 | <i>B. variegata</i> | 1969.05.10 | 2002.612.1 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér, belowthecastle | MZ1 | <i>B. variegata</i> | 1959.06.12 | 2002.611.1 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér, belowthecastle | MZ2 | <i>B. variegata</i> | 1959.06.12 | 2002.611.2 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér, belowthecastle | MZ3 | <i>B. variegata</i> | 1959.06.12 | 2002.611.3 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér, belowthecastle | MZ4 | <i>B. variegata</i> | 1959.06.12 | 2002.611.5 | 48.54 | 21.46 |
| Zemplén Mts | Füzér, belowthecastle | MZ32 | <i>B. variegata</i> | 1960.05.16-21. | HNHMHER 59.228.1 HNHMHER | 48.54 | 21.46 |
| Zemplén Mts | Füzér, Great Milic | MZ12 | <i>B. variegata</i> | 1960.07.12-14. | 2002.631.10 HNHMHER | 48.43 | 21.44 |
| Zemplén Mts | Füzér, Great Milic | MZ13 | <i>B. variegata</i> | 1960.07.12-14. | 2002.631.12 HNHMHER | 48.43 | 21.43 |
| Zemplén Mts | Füzér, Great Milic | MZ14 | <i>B. variegata</i> | 1960.07.12-14. | 2002.631.4 HNHMHER | 48.43 | 21.43 |
| Zemplén Mts | Füzér, Great Milic | MZ15 | <i>B. variegata</i> | 1960.07.12-14. | 2002.631.7 | 48.41 | 21.40 |
| Zemplén Mts | Füzér, Great Milic | MZ34 | <i>B. variegata</i> | 1960.07.13-15. | HNHMHER 60.172.1 HNHMHER | 48.88 | 21.46 |
| Zemplén Mts | Istvánkút, Pálháza | MZ7 | <i>B. variegata</i> | 1960.05.16-21. | 2002.626.8 HNHMHER | 48.46 | 21.47 |
| Zemplén Mts | Istvánkút, Pálháza | MZ8 | <i>B. variegata</i> | 1957.05.30 | 2002.627.1 HNHMHER | 48.46 | 21.47 |
| Zemplén Mts | Kőkapu | MZ9 | <i>B. variegata</i> | 1957.05.30 | 2002.629.4 HNHMHER | 48.43 | 21.44 |
| Zemplén Mts | Kőkapu | MZ10 | <i>B. variegata</i> | 1957.05.30 | 2002.629.9 HNHMHER | 48.43 | 21.44 |
| Zemplén Mts | Kőkapu | MZ11 | <i>B. variegata</i> | 1957.05.31 | 2002.630.1 | 48.43 | 21.44 |
| Zemplén Mts | Kőkapu | MZ29 | <i>B. variegata</i> | 1958.06.12 | HNHMHER 57.186.1 | 48.43 | 21.44 |
| Zemplén Mts | Pálháza, Istvánkút, Istvánkútispring | MZ28 | <i>B. variegata</i> | 1957.05.31 | HNHMHER 57.182.1 | 48.46 | 21.47 |
| Zemplén Mts | Pálháza, Istvánkút, Istvánkútispring | MZ33 | <i>B. variegata</i> | 1960.07.12-14. | HNHMHER 60.112.1 HNHMHER | 48.46 | 21.47 |
| Zemplén Mts | Rostalló | MZ24 | <i>B. variegata</i> | 1977.04.03-08. | 2002.638.1 | 48.42 | 21.43 |
| Zemplén Mts | Rostalló | MZ25 | <i>B. variegata</i> | 1977.04.03-08. | HNHMHER | 48.42 | 21.43 |

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|-------------|------------------------|------|---------------------|----------------|-----------------------|-------|-------|
| | | | | | 2002.638.5 | | |
| Zemplén Mts | Rostalló | MZ26 | <i>B. variegata</i> | 1977.04.03-08. | HNHMHER 2002.638.7 | 48.42 | 21.43 |
| Zemplén Mts | Rostalló | MZ27 | <i>B. variegata</i> | 1957.05.30 | HNHMHER 2002.638.8 | 48.42 | 21.43 |
| Zemplén Mts | Rostalló, Pálháza | MZ37 | <i>B. variegata</i> | 1977.04.03-08. | HNHMHER 77.43.1 | 48.42 | 21.43 |
| Zemplén Mts | Suslya Hill | MZ16 | <i>B. variegata</i> | 1960.07.13-15. | HNHMHER 2002.634.1 | 48.41 | 21.40 |
| Zemplén Mts | Suslya Hill | MZ17 | <i>B. variegata</i> | 1960.07.13-15. | HNHMHER 2002.634.3 | 48.43 | 21.43 |
| Zemplén Mts | Suslya Hill | MZ18 | <i>B. variegata</i> | 1960.07.13-15. | HNHMHER 2002.634.5 | 48.43 | 21.43 |
| Zemplén Mts | Suslya Hill | MZ19 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 2002.634.6 | 48.43 | 21.43 |
| Zemplén Mts | Suslya Hill | MZ35 | <i>B. variegata</i> | 1969.05.10 | HNHMHER 60.181.1 | 48.43 | 21.43 |
| Zemplén Mts | Telkibánya, Ósvavalley | MZ20 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 2002.637.1 | 48.48 | 21.34 |
| Zemplén Mts | Telkibánya, Ósvavalley | MZ21 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 2002.637.2 | 48.48 | 21.34 |
| Zemplén Mts | Telkibánya, Ósvavalley | MZ22 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 2002.637.4 | 48.48 | 21.34 |
| Zemplén Mts | Telkibánya, Ósvavalley | MZ23 | <i>B. variegata</i> | 1977.04.03-08. | HNHMHER 2002.637.5 | 48.48 | 21.34 |
| Zemplén Mts | Telkibánya, Ósvavalley | MZ31 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 59.227.1 | 48.48 | 21.34 |
| Zemplén Mts | Vadásztető | MZ6 | <i>B. variegata</i> | 1958.06.12 | HNHMHER 2002.625.5 | 48.46 | 21.47 |
| Zemplén Mts | Vadásztető | MZ30 | <i>B. variegata</i> | 1959.06.12 | HNHMHER 58.686.1 | 48.46 | 21.47 |