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Thesium linophyllon parasitizes expansive Calamagrostis epigejos and restricts its dominance in a long-term vegetation surve

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Running title

Thesium linophyllon's effect on Calamagrostis epigejos

Keywords

Biological control; clonal grass; conservation management; haustorium; landfill restoration; species rich grassland

Abstract

- Root-hemiparasitic interaction between the dominant grass *Calamagrostis epigejos* and hemiparasitic *Thesium linophyllon* was studied to assess the potential of the parasite to regulate dominance of the grass expanding into species rich steppe grasslands.
- First, we aimed at identification of physiological links between the two species, as a principal indicator of the parasitic relationship. Second, we analysed the dynamics of the two species in a steppe grassland at the foot of the Bükk Mountains, Hungary,

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where their joint presence is recorded in a long-term permanent-plot monitoring

dataset to detect a pattern associated with the parasitic ecological interaction.

Numerous well-developed functional haustoria of *Th. linophyllon* were identified on

the root systems of C. epigejos. The joint dynamics of C. epigejos and Th. linophyllon

displayed clear signs of the parasitic interaction: 1. The dynamics of *Th. linophyllon*

frequency was positively associated with the initial cover of C. epigejos. 2. Maximal

recorded cover values of the two species were strongly positively correlated and 3.

The extent of *C. epigejos* decrease in the vegetation was significantly positively

associated with maximum Th. linophyllon cover recorded throughout the monitoring

period.

We demonstrate that C. epigejos can be parasitized by Th. linophyllon which restricts

its abundance. Th. linophyllon thus has a potential to act as a native biological control

of *C. epigejos* in steppe grasslands.

Nomenclature: Tutin et al. (1964-1993).

Introduction

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2 Calamagrostis epigejos is a range-expanding grass spreading into natural and seminatural 3 grasslands in Central Europe (Rebele & Lehmann 2001). This expansion is one of the most 4 prominent factors threatening biodiversity of these highly diverse communities. C. epigeios is 5 a clonal species displaying guerila clonal strategy to colonize previously unoccupied plots 6 (Klimešová & de Bello 2009). Consequently and thank to its tall habit it often attains 7 dominance. C. epigejos produces a thick layer of slowly decomposing litter which has a 8 strong suppressive effect on other species (Rebele & Lehmann 2002). As a result, 9 competitively weaker species are excluded from the community, which decreases its diversity 10 (Somodi et al. 2008; Rebele 2014). 11 Standard conservation management of *Calamagrostis epigejos*-infested grasslands includes 12 intense mowing (twice a year or more; Lehmann & Rebele 2002). While such management 13 can indeed suppress the grass, it may also have negative effect on the rest of the community. 14 In steppe grasslands, species with late phenology (e.g. Aster amellus, Aster linosyris, Odontites luteus) and characteristic Stipa grasses may react negatively to mowing. Moreover, 15 16 intense mowing is laborious and costly in particular considering the difficult terrain on which 17 these communities are often located. Recently, introduction of hemiparasitic Rhinanthus 18 species (Orobanchaceae) was suggested as an alternative or complement to mowing (Těšitel 19 et al. 2017). These parasitic plants parasitize C. epigejos inflicting a massive damage. As a 20 result, C. epigejos may be exterminated from the community in short term while the characteristic species composition is restored (Těšitel et al. 2017). Rhinanthus spp. as 21 22 biocontrol agents are however of limited use in steppe grasslands. They are rather rare in 23 these communities (Těšitel et al. 2015a) due to the sensitivity of their seedlings to drought 24 (Ameloot et al. 2006) and specialized physiology characterized by wasting water (Jiang et al. 25 2003).

There are nevertheless other root-hemiparasitic plants which inhabit steppe grasslands. Thesium linophyllon (Thesiaceae, Santalales) is a clonal perennial herb (Klimešová & de Bello 2009) typical of dry calcareous grasslands and steppes (Těšitel et al. 2015a). Thesium linophyllon is an unselective generalist hemiparasite forming haustoria on all species in its surrounding (Dostálek & Münzbergová 2010). Root hemiparasites are generally known to alter competitive hierarchies in plant communities potentially increasing biodiversity if suppressing dominant competitors (Westbury et al. 2006; Pywell et al. 2007; Mudrák & Lepš 2010). No such effect has however been reported for *Thesium linophyllon yet*. Here, we use the inspection of anatomic structures and long term vegetation monitoring data to explore the interaction between Calamagrostis epigejos and Thesium linophyllon. We hypothesize that Calamagrostis can serve as a host for Thesium and that Calamagrostis abundance may be reduced by presence of *Thesium*. **Materials and Methods**

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Our analyses are based on long-term monitoring data collected in a fragment of a formerly grazed, but still species-rich, grassland at the foot of the Bükk Mountains, Hungary (47°54' N. 20°35′ E). The area is relatively dreir compared to other occurrences of C. epigejos with 600 mm annual rainfall and 9 °C mean annual temperature. For further details of site conditions and species composition see also Virágh (1982) and Virágh & Fekete (1984). A long-term monitoring followed initial experiments with herbicides in 1979 (Virágh 1987, 1989). These previous experiments changed the species composition of the treated plots considerably in the early years, but the assemblages had completely regenerated by 1988

(Virágh 1989). The experimental area have been revisited since than in selected years with
various intervals.

C. epigejos appeared first in a control plot in 1983 and had acquired dominance on half of the experimental area by 2002. There was strong directionality in its expansion: it started to spread from a patch present before abandonment at the bottom of a small valley, close to the study area. The pattern of experimental treatments, however, did not influence the spread pattern of C. epigejos. C. epigejos greatly transformed species composition (Somodi et al. 2008), however unaffected plots largely retained their species composition as a species rich grassland dominated Festuca rupicola with a slight shift towards Danthonia alpina codominance. Thesium lynophyllon was present at the site from the start, elevated frequency was found in 2002 already, but started to gain dominance after 2005.

Sampling description

The sampling design installed for the previous experiments was used later in the monitoring. The experiments were carried out in 1 m \times 1 m non-contagious plots arranged systematically in a grid with 50 cm spaces (Virágh 1987). Each 1 m \times 1 m plot was subdivided into 25 20 cm \times 20 cm subplots. From the originally 45 experimental plots we choose plots for the current analysis, for which data was available for each of the studied years: 2002-2005, 2013-2015 and at least one of the two species in focus was present with higher than 1% cover in any year. This yielded 10 plots for considerations in our analysis. To avoid errors potentially induced by treating adjacent 20 cm \times 20 cm subplots as replicates and to increase plot size, we merged the 4 subplots in each corner of each plot and used these in the analysis. Thus we obtained four 40 x 40 cm quadrats within each plot which we consider as independent observations hierarchically nested within plots.

Presence of haustorial connections between *Thesium linophyllon* and *Calamagrostis* epigejos was examined by visual inspection of *Calamagrostis* root system after excavation. Observed haustorial connections were cut out, washed in distilled water and fixed in 2.5% glutaraldehyde phosphate buffer. Their transverse sections were subsequently prepared by hand cutting for inspection under a light microscope. Presence of a xylem bridge in the haustoria and host-parasite xylem contact were examined as indicators of haustorium functionality (Cameron & Seel 2007).

Vegetation data analysis

The parasitic interaction between organisms is characterized by resource flow from the host to the parasite. As a result the parasite should benefit from host presence or abundance. By contrast, the host should be suppressed if parasite is present and with the level of suppression being positively affected by parasite abundance. This relationship may be more complicated in case of a hemiparasitic association but in environments where abiotic resources are scarce (such as steppe grasslands) it should be largely retained (Těšitel *et al.* 2015b). To detect the signature of parasitic interaction between *Calamagrostis* and *Thesium*, we formulated three null hypotheses corresponding to neutral interaction between the species, which were subsequently tested by the data originating from the long-term vegetation survey: H0₁: *Thesium* frequency (presence/absence) and its dynamics in monitoring quadrats does not depend on *Calamagrostis* cover. H0₂: Maximal recorded cover of both species in individual quadrats throughout the monitoring period are not correlated. H0₃: *Calamagrostis* cover in monitoring quadrat does not depend on the interaction between year and maximal *Thesium* cover recorded throughout the monitoring period.

To test H0₁, frequency (presence/absence) of *Thesium linophyllon* in quadrats across individual sampling years was summarized in a contingency table. The table was analysed by generalized estimating equations (GEE) with *Thesium* presence/absence as a binomial response and year, initial *Calamagrostis* cover (in 2002) and their interaction as predictors. The GEE assumed first order-autoregressive correlation among residuals within each monitoring quadrat. This correlation structure is suitable for time series but it assumes a continuous time series, which does not hold for our data. However, a trial fit of GEE with unstructured correlation structure did not identify any major change of correlation structure which would correspond to the gap in the time series. H₀₂ was tested by Pearson correlation coefficient between the maximum cover of *Thesium* and *Calamagrostis* recorded in individual quadrats throughout the monitoring period. H0₃ was tested by a linear mixed effect model containing Calamagrostis cover as response, year, maximal Thesium recorded cover in given quadrat and their interaction as fixed effect categorical predictors and quadrat identity nested within block as a random effect predictor. To graphically illustrate the association between Thesium abundance and the trend in Calamagrostis cover, we constructed a series of scatterplots displaying dependence of difference of Calamagrostis cover in actual year compared to 2002 on *Thesium* cover recorded in actual year. All cover data were square-root transformed prior to analysis to improve normality and homoscedasticity of the residuals. Square root transformation was used due to presence of zeros in the data. A priori defined Helmert contrasts (contrasting actual factor level to the mean of previous levels) were used to assess differences between years. All analyses were conducted in R, version 3.3.2 (R Core Team 2016) and R packages *nlme* (Pinheiro *et al.*

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Results

2014) and geepack (Højsgaard et al. 2006).

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126 Examination of Calamagrostis epigejos root systems revealed numerous Thesium haustoria 127 attached to both roots and rhizomes (Fig. 1a). Xylem bridge and xylem contact between the 128 host and parasite were identified in their anatomical structure (Fig 1b,c), which indicates 129 functionality of the haustorial connections. 130 The generalized estimating equations rejected H₀₁ by demonstrating significant effects of 131 year, initial Calamagrostis cover and their interaction on actual Thesium frequency. The 132 frequency of *Thesium linophyllon* significantly increased throughout the monitoring period (Table 1; GEE: $\chi^2_6 = 18.9$, P = 0.004). The most pronounced differences occurred between 133 134 2005 and 2013, when *Thesium* frequency increased from one third to almost two thirds of the 135 quadrats. Thesium presence was significantly positively associated with initial Calamagrostis cover (GEE: $\chi^2_1 = 6.1$, P = 0.010). The interaction term (GEE: $\chi^2_6 = 21.9$, P = 0.001) indicates 136 137 that the dynamics of *Thesium* frequency in quadrats was affected by *Calamagrostis* cover at 138 the beginning of sampling period. The interaction coefficient was significantly negative in 139 2003 (Helmert contrast; est = -0.0212, Wald z = 3.90, P = 0.048), significantly positive in 140 2014 (Helmert contrast; est = 0.0077, Wald z = 5.35, P = 0.021), and marginally non-141 significantly positive in 2015 (Helmert contrast; est = 0.0076, Wald z = 3.70, P = 0.054). 142 The correlation coefficient between the maximum cover of *Thesium* recorded throughout 143 the monitoring period and that of *Calamagrostis* was significantly positive (Pearson r =144 0.443, P = 0.007), which rejected H0₂. 145 H₀₃ was rejected by a significant effect of the interaction between year and maximal cover 146 of *Thesium* recorded throughout the monitoring period on actual *Calamagrostis* cover (Table 147 2). The interaction coefficients were significantly negative in 2004 ($t_{2004} = -1.98$, P = 0.049), 148 2013 ($t_{2004} = -3.46$, P = 0.001), 2014 ($t_{2004} = -3.507$, P = 0.001) and 2015 ($t_{2004} = -2.92$, P = 0.001)

0.023). In correspondence to the mixed -effect model, significant negative correlations

between actual *Thesium* cover and *Calamagrostis* cover difference compared to 2002 were also observed in these years (Fig. 2).

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Discussion

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Functional haustorial connections between *Thesium linophyllon* and *Calamagrostis* epigejos represent a strong indication of parasitic interaction between these two species. The analyses of long-term vegetation data managed to reject all three null hypotheses which assumed independent vegetation dynamics of the two species. Thesium frequency was found to increase over the ten years period and the probability of emergence in previously unoccupied quadrats was positively associated with Calamagrostis cover. There was also a positive association between maximum recorded cover of *Thesium* and *Calamagrostis*. At the same time, a significant decrease of *Calamagrostis* was positively associated with *Thesium* cover. These results indicate, that *Thesium* benefitted from high *Calamagrostis* abundance, while Calamagrostis was reduced by Thesium as expected in a host-parasite interaction. Still, we admit that the evidence on parasitic interaction between the two species is only based on observation which makes it weaker than evidence based on manipulative experiments. Unfortunately, such experiments (e.g. experimental sowing) are extremely difficult to conduct with Thesium linophyllon due to its very low germination rate (Dostálek & Münzbergová 2010). The observed effect of *Thesium* on *C. epigejos* is rather moderate. It seems that the two species can coexist in a long term. However, *Thesium* seems to be able to establish in *C*. epigejos stands and decrease its dominance in the community. That is important for maintaining and restoring steppe grassland biodiversity since the loss of biodiversity following C. epigejos establishment is a slow process and most species perish only after C.

epigejos attains dominance (Somodi et al. 2008). Moreover, Thesium linophyllon has recently been demonstrated to belong within top 5% species associated with high species richness in the vegetation of the Czech Republic (Fibich et al. 2017). Therefore, promoting Thesium abundance may have also other positive effects on diversity in addition to preventing C. epigejos dominance. The use of (hemi)parasitic plants to suppress populations of competitively strong dominants, either native or alien invasives, is an emerging topic in applied plant ecology. Recent research has demonstrated drastic effects the parasitic plants on their competitive hosts; e.g. *Pedicularis palustris* on *Carex acuta* (Decleer *et al.* 2013), Cuscuta campestris on Mikania micrantha (Yu et al. 2008) or Rhinanthus alectorolophus on Calamagrostis epigejos (Těšitel et al. 2017). Our study indicating the less pronounced, yet significant effect of *Thesium linophyllon* demonstrates that even moderate effects of parasitic plants only detectable in a long term can have a value for biodiversity conservation and restoration. In contrast to the above mentioned examples, it seems that *Thesium* does not require a special management measure to establish in C. epigejos stand. Furthermore, T. *lynophyllon* remains part of the community and thus can control even a future increase in C. epigejos due to an unplanned fire for example, which is known to boost C. epigejos spread (Rebele & Lehmann 2001, Deák et al. 2014). The moderate effect of *Thesium linophyllon* on *Calamagrostis epigejos* is probably caused by the structure of the santalean haustoria. These haustoria do not feature an open vascular connection with the host xylem and the uptake of nutrients proceeds via a contact parenchyma (Tennakoon et al. 1997; Hibberd & Jeschke 2001). That limits the amount of nutrients and in particular water acquired from the host while the loss of water is probably the major mechanism inflicting harm to the hosts of hemiparasites in dry habitats (Těšitel et al. 2015b). In addition, *Thesium* is a clonal and perennial species (Klimešová & de Bello 2009). Therefore, its strategy may be based on a conservative host use to secure host resources for

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future vegetation seasons. This contrasts with the ecological behaviour of many annual hemiparasites which need to maximize the resource acquisition from the host and create gaps in the vegetation to facilitate their seedling establishment (Demey *et al.* 2015; Lepš & Těšitel 2015).

Applications and Perspectives

Our study indicates the potential of *Thesium linophyllon* to regulate local abundance of competitive *Calamagrostis epigejos* in dry grassland. This effect may possibly be used in nature conservation practice to reverse the biodiversity decline associated with *C. epigejos* expansion. However, further research of *Thesium linophyllon* reproductive biology and ecological requirements is needed to identify measures promoting its abundance and to develop methods of introduction to unoccupied target sites. Subsequently, *Thesium linophyllon* may be tested as a promising hemiparasitic species to colonize and increase diversity of extreme habitats such as post-mining sites, which have a successional potential to develop into steppe grasslands (Prach *et al.* 2013) but such development may be hindered by *Calamagrostis epigejos* and other synanthropic grass dominance (Prach & Pyšek 2001). *Thesium linophyllon* is native in Central and Eastern Europe (Meusel *et al.* 1965), where possible target post-mining or post-industrial sites are available in abundance.

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249 Hibberd J.M., & Jeschke W.D. (2001) Solute flux into parasitic plants. *Journal of* 250 *Experimental Botany*, **52**, 2043–2049. 251 Højsgaard S., Halekoh U., Yan J. (2006) The R Package geepack for Generalized Estimating 252 Equations. *Journal of Statistical Software*, **15**, 1-11. 253 Jiang F., Jeschke W.D., Hartung W. (2003) Water flows in the parasitic association 254 Rhinanthus minor/Hordeum vulgare. Journal of Experimental Botany, **54**, 1985–93. 255 Klimešová J., de Bello F. (2009) CLO-PLA: The database of clonal and bud bank traits of 256 Central European flora. *Journal of Vegetation Science*, **20**, 511–516. 257 Lehmann C., Rebele F. (2002) Successful management of Calamagrostis epigejos (L.) ROTH 258 on a sandy landfill site. *Journal of Applied Botany*, **76**, 77–81. 259 Lepš J., Těšitel J. (2015) Root hemiparasites in productive communities should attack 260 competitive host, and harm them to make regeneration gaps. Journal of Vegetation 261 Science, **26**, 407–408. 262 Meusel H., Jäger E., Weinert E. (1965) Vergleichende Chorologie der zentraleuropäischen 263 Flora. Volume 1. Gustav Fischer Verlag, Jena. 264 Mudrák O., Lepš, J. (2010) Interactions of the hemiparasitic species Rhinanthus minor with 265 its host plant community at two nutrient levels. Folia Geobotanica, 45, 407–424. Pinheiro J., Bates D., DebRoy S., Sarkar D. & R Core Team (2014) Nlme: Linear and 266 267 Nonlinear Mixed Effects Models. http://CRAN.R-project.org/package=nlme 268 Prach K., Pyšek P. (2001) Using spontaneous succession for restoration of human-disturbed 269 habitats: Experience from Central Europe. *Ecological Engineering*, 17, 55–62. Prach K., Lencová K., Řehounková K., Dvořáková H., Jírová A., Konvalinková P., Mudrák 270 271 O., Novák J., Trnková R. (2013) Spontaneous vegetation succession at different central 272 European mining sites: A comparison across seres. Environmental Science and Pollution

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Research, 20, 7680–7685.

274 Pywell R.F., Bullock J., Tallowin J., Walker K., Warman E., Masters G. (2007) Enhancing 275 diversity of species-poor grasslands: an experimental assessment of multiple constraints. 276 Journal of Applied Ecology, 44, 81–94. R Core Team (2014) R: A language and environment for statistical computing. R Foundation 277 278 for Statistical Computing, Vienna, Austria. URL http://www.R-project.org/ 279 Rebele F. (2014) Species composition and diversity of stands dominated by Calamagrostis 280 epigejos on wastelands and abandoned sewage farmland in Berlin. Tuexenia, 34, 247— 281 270. 282 Rebele F., Lehmann C. (2001) Biological flora of central Europe: Calamagnostis epigejos (L.) 283 Roth. Flora, 196, 325–344. 284 Rebele F., Lehmann, C. (2002) Restoration of a Landfill Site in Berlin, Germany by 285 Spontaneous and Directed Succession. *Restoration Ecology*, **10**, 340–347. 286 Somodi I., Virágh K., Podani J. (2008) The effect of the expansion of the clonal grass 287 Calamagrostis epigejos on the species turnover of a semi-arid grassland. Applied 288 *Vegetation Science*, **11**, 187–192. 289 Tennakoon, K.U., Pate, J.S., & Arthur, D. (1997) Ecophysiological aspects of the woody root 290 hemiparasite Santalum acuminatum (R. Br.) A. DC and its common hosts in south 291 western Australia. Annals of Botany, 80, 245–256. 292 Těšitel J., Fibich P., de Bello F., Chytrý M., Lepš J. (2015a) Habitats and ecological niches of 293 root-hemiparasitic plants: an assessment based on a large database of vegetation plots. 294 *Preslia*, **87**, 87–108. 295 Těšitel J., Těšitelová T., Fisher J.P., Lepš J., Cameron D.D. (2015b) Integrating ecology and 296 physiology of root-hemiparasitic interaction: interactive effects of abiotic resources 297 shape the interplay between parasitism and autotrophy. New Phytologist, 205, 350–360.

298	Těšitel J., Mládek J., Horník J., Těšitelová T., Adamec V., Tichý L. (2017) Suppressing
299	competitive dominants and community restoration with native parasitic plants using the
300	hemiparasitic Rhinanthus alectorolophus and the dominant grass Calamagrostis
301	epigejos. Journal of Applied Ecology, 54, 1487-1495
302	Tutin T.G., Heywood V.H., Burges N.A., Moore D.M., Valentine D.H., Walters S.M. &
303	Webb D.A. (Eds) (1964-1993). Flora Europaea, Vols. 1-5. Cambridge University Press,
304	Cambridge, UK: 2392 pp.
305	Virágh K., Fekete G. (1984) Degradation stages in a xeroseries: composition, similarity,
306	grouping, coordination. Acta Botanica Hungarica, 30, 427-459.
307	Virágh K. (1982) Vegetation dynamics induced by some herbicides in a perennial grassland
308	community, I. Acta Botanica Hungarica, 28, 427-447.
309	Virágh K. (1987) The effect of herbicides on vegetation dynamics: A five year study of
310	temporal variation of species composition in permanent grassland plots. Folia Geobo-
311	tanica and Phytotaxonomia, 22, 385-403.
312	Virágh K. (1989) The effect of selective herbicides on temporal population patterns in an old
313	perennial grassland community. Acta Botanica Hungarica, 35, 127-143.
314	Westbury D.B., Davies A., Woodcock BA, Dunnett N. (2006) Seeds of change: The value of
315	using Rhinanthus minor in grassland restoration. Journal of Vegetation Science, 17, 435-
316	446.
317	Yu H., Yu F.H., Miao S.L., Dong M. (2008). Holoparasitic Cuscuta campestris suppresses
318	invasive Mikania micrantha and contributes to native community recovery. Biological
319	Conservation, 141, 2653–2661.

321 **Table titles** 322 323 **Table 1.** Frequency of *Thesium linophyllon* in individual monitoring quadrats in the course of 324 the study period. 325 **Table 2.** Summary of linear mixed-effect model testing dependence of *Calamagrostis* cover 326 on maximum recorded *Thesium* cover in monitoring quadrats. 327 328 Figure captions 329 330 **Figure 1.** Morphology and anatomy of *Thesium linophyllon* haustoria attached to 331 Calamagrostis epigejos roots. (a) Outer morphology of the haustoria. (b) Cross-section of the 332 haustorium attached to the host root (c) Details of the xylem contact between the host and the 333 parasite. Ha: Haustorium, HR: Host root, PR: Parasite root, VC: Vascular core of the 334 haustorium, XB: Xylem bridge, HB: Hyalline body, PXy: Parasite xylem, HXy: Host xylem. 335 Figure 2. Trends in Calamagrostis epigejos abundance displayed by cover difference 336 compared to 2002 at individual monitoring quadrats in 2003-2005 and 2013-2015. Regression line is displayed for significant relationships. $^*P < 0.05$, $^{**}P < 0.01$, $^{***}P < 0.001$ 337