

This manuscript is contextually identical with the following published paper:

Heltai, B., Sály, P., Kovács, D., Kiss, I. (2015): Niche segregation of sand lizard (*Lacerta agilis*) and green lizard (*Lacerta viridis*) in an urban semi-natural habitat. *Amphibia-Reptilia* 36(4): 389-399.

DOI:10.1163/15685381-00003018

The original published pdf available in this website:

<http://booksandjournals.brillonline.com/content/journals/10.1163/15685381-00003018>

**Niche segregation of sand lizard (*Lacerta agilis*) and green lizard (*Lacerta viridis*)
in an urban semi-natural habitat**

Botond Heltai¹, Péter Sály², Dániel Kovács¹, István Kiss¹

¹Szent István University, Department of Zoology and Ecology, H-2103 Gödöllő, Páter K. u. 1., Hungary

²Balaton Limnological Research Institute of the Hungarian Academy of Sciences, H-8237 Tihany, Klebelsberg Kuno 3., Hungary

e-mail: heltaib2@gmail.com

Abstract. Different types of semi natural habitats has important role in long-term survival and maintenance of reptile species in urban environments. Heterogeneous urban green islands can provide conditions that enable competing species to live together in relatively small areas. However, the key mechanism of coexistence could vary from types of habitats and taxa. We investigated the population structure, the fine scale habitat segregation and diel activity pattern of two lizard species (*Lacerta viridis* and *L. agilis*) coexisting populations in a town cemetery. We hypothesized, that fine scale habitat segregation is a more important factor in coexistence than differences in diel activity pattern, because of the environmental dependent thermoregulation constrain. During the study, 178 *L. agilis* and 79 *L. viridis* occurrences were recorded. The daily activity patterns of both species were very similar, with peaks found in the forenoon and afternoon. Lizards were found to be the most active at 31–32°C. The probability of occurrence of the species was influenced more by the differences in the random selection of parcels rather than the random differences in the survey days. Around the proximate observation point the bush covered areas was significantly higher than average in the parcels. Our results showed that niche segregation based on fine scale habitat patches had a fundamental role in the coexistence of the two lizard species. *L. agilis* preferred the more open spaces, whereas *L. viridis* preferred areas with more bushes, but there was no difference in the daily activity pattern of the two species.

Keywords: habitat selection, activity pattern, urban ecology, public cemetery

Introduction

Long-term maintenance of lizard populations depends fundamentally on the quality and accessibility of suitable habitats. Expansions of the build-up areas, increasing gardening (e.g. mowing) often lead to fragmentation, degradation and loss of habitats in suburban and urban areas (Martens and Jelden, 1992). Due to habitat fragmentation the population size of lizards decreases, which could lead to local extinctions. Hence, green areas, enclosed in urban or inhabited areas or adjacent agricultural green areas, can provide suitable habitats for reptile species (Ilosvay, 1977; Bender, 1997; Schmidt-Loske, 1997; Rugiero, 2004; Mollov, 2005; Strugariu et al., 2007). Vignoli et al. (2009) found that the richness and the diversity of the herpetofauna are strongly influenced by the occurrence of large remnant woods, water and extended ecotonal habitats presence. Heterogeneous habitat patches within the towns can highly promote the accommodation success (Faeth et al., 2005). Because of their sizes and their undisturbed conditions cemeteries contribute significantly to habitats of lizard species in urban areas (Barbault and Mou, 1988; Schwartz, 2010). However, species can respond in different ways to urban effects (McKinney, 2008). Beside the less tolerant ones, some species are able to accommodate to the altered conditions.

Urban habitats can also have negative effects. The genetic isolation of urban lizard populations could be prevented by establishing green corridors (Strijbosch and Van Gelder, 1997; Altherr, 2007). The continuous anthropogenic effects of varying intensity, the drastic changes and disappearance of habitat patches and the increased number of potential predators are all intensified stresses for urban animals (Corbett, 1988, 2001; Koenig, Shine and Shea, 2002; Woods, McDonald and Harris, 2003; Ihász et al., 2006, Shochat, Lerman and Fernández-Juricic, 2010). The selection pressure could alter the

behaviour, morphology, genetic traits, of plant and animal species in urban conditions (Shochat et al., 2006). A small lizard population of an island could be similar to an urban lizard population structure. Cats seem to be the biggest threat to the lizards in a small island (van Bree, Plantaz and Zuiderwijk, 2006). Elbing (1997) found that the social behaviour and the distribution patterns of lizards can be different on islands, than habitats with optimal condition. Arntzen and Sá-Sousa (2007) found that insular lizard populations have some morphological differences, and a lower nuclear genetic heterozygosity than on the mainland, but no correspondence was observed between morphological and molecular patterns of intra-specific differentiation. Light pollution in cities or towns also affects reptiles (Perry et al., 2008), Carretero et al. (2012) recorded nocturnal activity in *Podarcis muralis*. The invasion of alien plant species has an adverse effect on the species richness of lizard species (Jellinek, Driscoll and Kirkpatrick, 2004). Capula, Luiselli and Rugiero (1993) investigated the competition and niche segregation between lizard species living together in urban conditions, while the habitat choice of reptile species was studied by Mollov (2011); Rugiero and Luiselli (2007). Structural features of habitats can sometimes override the effect of habitat fragmentation (Santos et al., 2008). Jellinek, Driscoll and Kirkpatrick (2004) found that the structure of lizard communities is primarily determined by the vegetation composition. In contrast, Garden et al. (2007) stated that the habitat structure is more important for the reptiles and small mammals than the vegetation composition. Generally, the increasing edge zones do not affect the species richness of a region or the abundance of lizards. Some species prefer habitats with heterogeneous structures, which can only be found in larger patches (Jellinek, Driscoll and Kirkpatrick, 2004). Some lizard species show a high degree of site fidelity, so they conquer other habitat types

only with difficulty, while others adapt well to changes and to the higher productivity of urban green areas (Koenig, Shine and Shea, 2001).

Our study focuses on two lizard species: *Lacerta agilis* and *L. viridis*. The sand lizard (*L. agilis*) is one of the most common lizard species in Hungary. It can be found in fields, forest edges, grassy or bushy edges of trenches, pastures, meadows, along railway embankments, marshes, peat-bogs, reedy habitats and urban gardens (Dely, 1983). The European surveys show a very diverse habitat use of the species (House and Spellerberg, 1983; Glandt, 1986; Strijbosch, 1986; Dent and Spellerberg, 1987; Amat, Llorente and Carretero, 2003; Edgar and Bird, 2006; Ceirâns, 2007; Èeirâns, 2007). Edge zones with heterogeneous vegetation and higher rates of shrub cover are also preferred by *L. agilis*. Shrubs are used both as hideouts and basking place (Stamps, 1977; Nemes, 2002; Nemes et al., 2006) while the over shaded woody habitats are avoided by the species (House and Spellerberg, 1983).

The green lizard (*Lacerta viridis*) is also a common species in Hungary, but its distribution is influenced especially by the presence of bushy edge zones. Hence, vineyards, orchards and cemeteries with tree groups and bushy patches could provide suitable habitats for the green lizard (Dely, 1983). *L. viridis* can occur in heavily disturbed areas (Iftime and Iftime, 2006). This species often climbs up trees (Mikátová, 2001; Iftime, 2005). According to some authors (Vasváry, 1926; Arnold, Burton and Oviden, 1978; Arnold, 1987; Mollov, 2005; Covaciu-Marcov et al., 2008; Strugariu and Gherghel, 2008), the green lizard prefers the woody, bushy habitats.

The seasonal activity patterns of the two species are slightly different and there are differences between males and females in this respect, too (Glandt, 1995; Mikátová, 2001; Amat, Llorente and Carretero, 2003). The two lizard species usually have two

daily activity peaks in the summer (House and Spellerberg, 1983; Korsós, 1984, 1986; Korsós and Gyovai, 1988; Sound and Veith, 2000; Amat, Llorente and Carretero, 2003; Kuranova et al., 2005). However, if the temperature at noon does not rise too high, there can be only one activity peak for sand lizard (House, Taylor and Spellerberg, 1979; Korsós, 1984, 1986).

In urban areas there could be higher competition between species, than in natural open areas so the aim of our study was to examine how the niche segregation of the two species living together in an urban green area in a cemetery takes place. Our specific questions were: 1) Does the spatial distribution of the lizard population relate to the habitat diversity of the parcels? 2) Can habitat preferences be established? 3) Are there any differences in the daily activity patterns of the two species and what factors can influence them?

Materials and methods

Study area

The total area of Public Cemetery of Dunaújváros (70 km from Budapest) is 21 ha. It is surrounded by houses and roads, but it is attached to the forest belt and a streambed, which provide corridor for lizards (Dent and Spellerberg, 1988; Scali and Zuffi, 1994; Iftime, 2005; Sos, 2007; Covaciu-Marcov et al., 2009; Molloy, 2011). In the cemetery the original indigenous vegetation with Tartar maple on loess could not be found, most of the plants having been planted (Simon, 1967; Zólyomi, 1967).

Eight parcels with the same area (2500 m²) were selected for the investigation. In the parcels we assessed the stone covered area and the perimeter of all gravestones, the surface of grass and shrub covered areas, and the area shaded by shrubs and trees. The opening year of the parcels was given by the year of the earliest grave.

Field work

The study was carried out during the period of 1–7 July 2012. Walking of the parcels took place over the same period of time and at an even pace to spot the highest number of individuals. On the first day the parcels were walked through four times, henceforth three times a day while two walks were completed on

the last day. To discover the daily activity pattern of lizards we walked each parcel during different parts of the day (early morning, at noon, in the afternoon) in random distribution and several repeats.

Always the same observer made the surveys. We walked through every other row without disturbing the lizards and always following the same route to make the surveys comparable. Usually the lizards were not frightened, they were staying at the same small plot. We recorded the species, gender, age of each individual lizard. We estimated the percentage of grass covered, grave covered and shrub covered areas around the observed individuals in a range of 1 m (4 m²). While the eight parcels were walked, the air temperature was measured four times in the grass (twice in the sunlight and twice in the shade) and four times on the gravestone (twice in the sunlight and twice in the shade). The mean of these eight measurements was used in the analysis. The thermometer was set 5 cm above the ground. At the beginning of surveys we recorded the number of people present in the given parcel to specify the disturbing effects.

Data analyses

We used generalized linear mixed models (Zuur et al., 2009) with binomial error distribution and logit link function to reveal the relationship between the probability of occurrence of the lizards and the studied explanatory factors. We used nlme, lme4, MASS R packages for analysis. The time of observation, the mean temperature during observation and the number of humans staying within the parcel during the observation period were the fix explanatory variables in the models. In the reduced model, which describes the probability of occurrence, time, temperature and the interaction of the two had effect, whereas the number of humans staying within the parcel during the observation period was no significant effect. The parcel and the date of observation nested within the parcel were used as random factors in the models. Our models assumed that the effects of the explanatory variables on the probability of occurrence would be the same in all parcels and on each day of the observation period and differences between the parcels and days would only contribute to the occurrence probability with random fluctuations. We fitted three models, one on the general occurrence of the lizards (i.e. irrespectively of the species identity), and two others on the sand lizard and green lizard data, respectively. In order to control the autocorrelative nature of the subsequent observations within a certain parcel on a certain day, a first-order autocorrelation structure was built in the models (Zuur et al., 2009).

Principal component analysis was performed to determine which are the most important environmental factors classifying the parcels.

Pearson's Chi-squared test with Yates' continuity correction was used to compare the apparent sex-ratio and the age structure between species and within species against the hypothetical 0.5:0.5.

Two-sample Kolmogorov-Smirnov test was used to find differences between the two species' daily activity patterns and the same test was used to find differences between the two species' relative occurrence depending on the temperature.

All the tests were conducted in R statistical environment (version number 3.2) (R Core Team, 2012).

Polynomial regression model was used in two cases: 1st to indicate the daily activity patterns of the two species and 2nd to show the relationship between the observed number of individuals and the vegetation covering the parcels. The tests were conducted in Microsoft Office Excel (Winston, 2007).

Analysis of Variance (ANOVA) was used to find differences between the two species' habitat patch preference followed by the Fisher's Least Significant Difference post hoc test (LSD test). Namely we test whether the surface vegetation cover percentages measured near the observation points of the two species were different between species; or rather they were different from the mean cover of the parcels. All percentage variables were arcsine square-root transformed in order to normalize data distribution before the statistical analysis. The mean of percentage grass cover, shrub cover and gravestone cover were compared with the mean surface cover data of the eight parcels. To carry out the statistical analysis the IBM SPSS software was used (IBM Corp. Released, 2011).

Results

Differences between parcels

The traits of chosen parcels proved to be different, mainly with regard to the vegetation cover and the percentage of the shade. The results of a principal components analysis based on the covariance matrix showed that there are two main principal components. The first principal component (PC1) explained 58.8 % of total variance (eigenvalue: 2.9414) and the second principal component (PC2) explained 21.6 % of the variance (eigenvalue: 1.078). The last three PC's have little apparent significance. The percentage

of grass covered areas and the absence or the very short cumulated perimeters of gravestones and absence of shrubs were the most dominant traits in the first PC. The percentage of grass covered area was higher in the E2 and XXXIV parcels. The absence of the area shaded by shrubs and trees and the absence of shrub covered area were dominant traits in the second PC (Fig. 1.).

Observation data of lizard species

178 observations were made for *Lacerta agilis* in 7 days. In parcel E2 no lizards were found. There were fewer observations made for *L. viridis*, only 79 during the study period.

In both lizard species the proportion of females among the observed individuals was higher, but the differences were not significant (Chi-squared test; $\chi^2_{L.a.} = 0, P = 1$, $\chi^2_{L.v.} = 0.3369, P = 0.5616$), and the differences in the apparent sex-ratio between the two species were also not significant ($\chi^2 = 0.4087, P = 0.5226$). Ratio of number of adults and subadults was 3.91 in *L. agilis*, and 7.77 in *L. viridis*. No juveniles were observed during the study period. The ratios of adults were significantly higher than the hypothetical 0.5 in both species ($\chi^2_{L.a.} = 32.8599, P < 0.001$, $\chi^2_{L.v.} = 25.8908, P < 0.001$), but the apparent age groups between the two species did not differ significantly ($\chi^2 = 2.3622, P = 0.1243$).

Occurrence and daily activity pattern of lizard species

The probability of occurrence described in the reduced model was mainly dependent on the time, temperature and interaction between the two (Table 1.). The variance of the probability of occurrence calculated on the basis of the differences among the parcels was 34.79 (logit scale), which was 4.28-fold higher than the variance calculated based on the days, which was 8.13 (logit scale). During a short, approx. one-week-long

summer observation period the traits of the parcels were found to influence the daily activity patterns of the lizards more than the differences between the days. With regard to the variance of the probability of occurrence of *L. agilis* based on the differences of parcels was 1.63 (logit scale), while the variance based on the differences of the days was 0.10 (logit scale) (Table 1.). The probability of occurrence of the *L. viridis* showed no correspondence with any of the studied traits (Table 1.).

The mean temperature, slightly influenced the relative occurrence of the two lizard species (Fig. 2.), and the difference between the two species was not significant (Kolmogorov-Smirnov test; $D = 0.2273$ $P = 0.6208$). Both species were found to be the most active around 31–32°C.

The sample size allowed to model only the relationship between the temperature and the probability of occurrence (Table 1.) of *L. agilis*. The sand lizard probability of occurrence was the highest around 33°C, except for two parcels (Fig. 3.). In parcel XXXIV a low number of individuals were recorded in the forenoon, while in parcels E2 no individuals were observed.

The daily activity patterns of the two species were found to be very similar (Fig. 4.) and did not differ significantly (Kolmogorov-Smirnov test; $D = 0.3077$ $P = 0.5696$). Both species had their first activity peaks in morning from 9 to 10 o'clock whereas in early afternoon the observed number of lizards decreased. In late afternoon they showed a second activity peak, which was much higher than the first. For both species the fourth power polynomial trend line, which describes the changes indicated a strong correlation with the mean temperature ($R^2 = 0.8489$ for *L. agilis* and $R^2 = 0.7445$; *L. viridis*, $y_{L.a} = -0.045x^4 + 2.5312x^3 - 51.695x^2 + 450.85x - 1397.8$, $y_{L.v.} = -0.0324x^4 + 1.827x^3 - 37.369x^2 + 326.42x - 1010.5$).

Lizard species habitat patch choice

We found strong relationship between the vegetation covering the parcels and the number of pooled data of observed individuals. The shrub covered area and the number of individuals showed strong correlation (based on the quadratic polynomial trend line $y = -1.5833x^2 + 23.083x - 28.75$, $R^2 = 0.789$).

The highest number of individuals was observed in the parcels with 5–8% shrub cover rate. In the new parcels, where there was no undergrowth or the shrub cover percentage was 3%, similarly to the very shaded parcel where the shrub cover percentage was 1,24% *L. viridis* was not and only a few *L. agilis* individuals were observed. The number of observed individuals was higher in the more heterogeneous group of parcels containing more shrubs.

The additive effect of trees and shrubs shade on the observed number of individuals showed a strong correlation (based on the quadratic polynomial trend line $y = -4.1071x^2 + 38.321x - 32.964$, $R^2 = 0.6211$). The maximum number of individuals was found in parcel XXV with 30% shade. The parcel with little vegetation cover and the ones with too much shaded area were found to provide unsuitable habitats for the lizards.

ANOVAs scores on the land cover percentage revealed significant effect on the lizard species for grass- and shrub coverage, (ANOVA; Grass %: $F_{2, 261}=12.418$, $P < 0.001$; Shrub%: $F_{2, 261}=10.601$, $P < 0.001$) but no significant effect on tombstone coverage (Tombstone %: $F_{2, 261}=0.052$, $P = 0.949$).

When measuring habitat preferences the grass cover around the observation point of individuals differed significantly between the two species (LSDtest; $P < 0.001$). While in case of *L. viridis* we found significant difference between the grass cover

around the observation point of individuals and the average in the parcels ($P < 0.01$), *L. agilis* did not show the same pattern ($P = 0.162$) (Fig. 5/A.).

The shrub cover around the direct environment of the detected individuals of *L. viridis* was found to be significantly higher, than in the direct environment of *L. agilis* (LSDtest; $P < 0.001$). The shrub cover of the close environment of *L. viridis* proved to be significantly higher ($P < 0.05$), than the shrub cover of the parcels (Fig. 5/B.).

With regard to the gravestone cover the differences between the cover rates around the two species' individuals were not significant (LSDtest; $P = 0.88$). The gravestone cover around the observation points of *L. agilis* ($P = 0.79$) and *L. viridis* ($P = 0.75$) did not differ significantly from the parcels' means, but the gravestone covered areas were lower around the two species (Fig. 5/C.).

Discussion

This study investigated the coexistence of two lacertid lizard species (*Lacerta agilis* and *L. viridis*) in a town cemetery serving as semi-natural green island habitat. Although our survey provides only a snapshot on the lizard populations, it seems that cemeteries with various semi-natural microhabitat patches, such as shrubs, grassy areas, or tombs could offer suitable habitats to coexist for *L. agilis* and *L. viridis*. This corresponds to former studies (Ilosvay, 1977; Bender, 1997; Schmidt-Loske, 1997; Rugiero, 2004; Mollov, 2005) that also reported urban green areas as suitable lizard habitats.

In the chosen parcels of the cemetery lizards were found except in parcel E2, which was the consequence of the high degree of shade and the scarce undergrowth vegetation. From day 3 to 4 during the survey time the observed number of individuals showed a decreasing tendency, which could be explained by the slight drop in the daily

mean temperature. The number of females was not significantly higher than males in both lizard populations.

We identified that the mean temperature consisting of the temperatures measured in the sunlight and in the shade near the ground influenced the occurrence of *L. agilis* and the *L. viridis* but the two species did not differ significantly, so the air temperature had no role in their niche segregation. The optimal air temperature was between 31–32°C, when the lizards were the most active. This corresponded to the findings of Castilla, Van Damme and Bauwens (1999).

The daily activity patterns of both species were similar, with the activity peaks in forenoon and afternoon and their retreats in the early afternoon. As a result, their daily activity patterns did not seem to play any important role in their niche segregation. The activity patterns of the two lizard species detected in summer time were identical with the results of other studies showing two activity peaks (House and Spellerberg, 1983; Korsós, 1984, 1986; Korsós and Gyovai, 1988; Sound and Veith, 2000; Amat, Llorente and Carretero, 2003; Kuranova et al., 2005). The reason for the differences in the lizards' daily activities can be attributed to their different body size (House, Taylor and Spellerberg, 1979; Strijbosch, 1986) or differences in their thermoregulation (Rismiller and Heldmaier, 1982; Castilla, Van Damme and Bauwens, 1999; Grbac and Bauwens, 2001).

Their probability of occurrence was influenced more profoundly by the differences in the characteristics of the parcels rather than the differences between the days. During a short observation period the differences arising from the random selection of the parcels had a more substantial influence on their activity patterns than the random differences between the days.

The middle aged heterogeneous parcels were suitable for the lizards unlike the shaded or lawned ones which were seldom frequented by them. The most suitable parcels were heterogeneous with bushy parts providing a suitable percentage of shaded surface. Korsós (1984, 1986) did not find statistically significant proof to support the statement that the rate of shrubbery areas was higher in the environment of the *L. viridis*, than in the environment of the *L. agilis*. Cooke (1991) obtained similar results according to which the habitat is suitable for the *L. agilis* up to 9% shrub cover. Bushes are mainly used as hiding place by the *L. agilis* (Nemes, 2002; Nemes et al., 2006), but House and Spellerberg (1983) observed them climbing up and basking on shrubs. The *L. viridis* also uses these sites as basking place (Mikátová, 2001) and hiding place, too. We also observed this behaviour.

The fine spatial scale habitat choice of the two lizard species was slightly different, which makes their coexistence in one particular habitat possible. The *L. viridis* tends to choose areas with more shrubs, than *L. agilis*, while the sand lizard usually occurs in more open areas with higher percentage of grass cover. In the majority of the parcels we observed high rates of grass cover, which could be suitable for the sand lizard preferring more open spaces, while the *L. viridis* tends to avoid the open, grass-covered areas (House and Spellerberg, 1983; Glandt, 1986; Strijbosch, 1986; Arnold, 1987; Cooke, 1991; Amat, Llorente and Carretero, 2003; Nemes et al., 2006; Ceirâns, 2007). In their close environment the rates of open areas are low. In the shaded parcels lizards cannot find large enough sunny sites, which are therefore not suitable for them (House and Spellerberg, 1983; Dent and Spellerberg, 1987; Cooke, 1991). Most of the green lizard individuals were centred around the bushy habitats offering them hiding place. The gravestone covered areas did not contribute to the separation of their

habitats. In our study the preferential use of stone (gravestone) as sunny site or shelters by the lizards was not observed, as it was detected for other green lizard species before (Díaz, Monasterio and Salvador, 2006).

Our findings show that there are no differences between lizards' habitat use, and daily activity pattern in urban area and natural area. Lizards need a structurally complex habitat with a combination of different patches. This corresponds to former study (Becker 2015) that also reported urban areas as suitable lizard refuges; urbanisation do not affected sand lizard population densities. The *L. agilis* tends to choose more open areas, than *L. viridis*. Angelici, Luiselli and Rugiero (1997) found, that microhabitat differences could explain the dietary differences between young and adult lizards. Therefore predation may play a role in their niche segregation (Korsós 1984).

To conclude, our findings highlight the role of cemeteries as less disturbed, semi-natural urban areas in conservation of lizard populations. In highly populated towns, cemeteries along with city parks should be considered as potential refuge for less mobile territorial animals such as lizards. This point of view poses a question of considering the possibility of attributing limited protection to these urban areas.

Acknowledgements

We thank Dunaújvárosi Vagyonkezelő ZRT (Property-holder of Dunaújváros PLC) for making the fieldwork possible and for supplying us with data. We would like to thank J. Vörös for comments on earlier draft of the manuscript. We are grateful to S. Farkas and B. Szabó for language corrections. The study was supported by the Hungarian Ministry of Human Resources (Research Centre of Excellence - 8526-5/2014/TUDPOL).

References

- Altherr, G. (2007): From genes to habitats: effects of urbanisation and urban areas on biodiversity. Dissertation, University of Basel.
- Amat, F., Llorente, G.A., Carretero, M.A. (2003): A preliminary study on thermal ecology, activity times and microhabitat use of *Lacerta agilis* (Squamata: Lacertidae) in the Pyrenees. *Folia Zool.* **52**: 413-422.
- Angelici, M.A., Luiselli, L., Rugiero, L. (1997): Food habits of the green lizard, *Lacerta bilineata*, in central Italy and a reliability test of faecal pellet analysis. *Ital. J. Zool.* **64**: 267-272.
- Arnold, E.N., (1987): Resource partition among lacertid lizards in southern Europe. *J. Zool.* **1**: 739–782.
- Arnold, E.N., Burton, J.A., Oviden, D. (1978): A field guide to the reptiles and amphibians of Britain and Europe. London, Collins.
- Arntzen, J.W., Sá-Sousa, P. (2007): Morphological and genetical differentiation of lizards (*Podarcis bocagei* and *P. hispanica*) in the Ria de Arosa Archipelago (Galicia, Spain) resulting from vicariance and occasional dispersal. In: *Biogeography, Time, and Place: Distributions, Barriers, and Islands*. p. 365-401. Renema, W., Ed., Dordrecht, Springer.
- Barbault, R., Mou, Y.P. (1988): Population dynamics of the common wall lizard, *Podarcis muralis*, in southwestern France. *Herpetologica* **44**: 38-47.

Becker, M., Buchholz, S. (2015): The sand lizard moves downtown – habitat analogues for an endangered species in a metropolitan area. *Urban Ecosyst.* 08/2015: DOI: 10.1007/s11252-015-0497-x.

Bender, C. (1997): Demography of a small population of the endangered common wall lizard (*Podarcis muralis*, Lacertidae) in Western Germany. In: *Herpetologia Bonnensis*, p. 27-34. Böhme, W., Bischoff, W., Ziegler, T., Eds, Bonn, Societas Europaea Herpetologica.

Capula, M., Luiselli, L., Rugiero, L. (1993): Comparative ecology in sympatric *Podarcis muralis* and *P. sicula* (Reptilia: Lacertidae) from the historical centre of Rome: What about competition and niche segregation in an urban habitat? *Ital. J. Zool.* **60**: 287-291.

Carretero, M.A., Sillero, N., Lazić, M.M., Crnobrnja-Isailović, J. (2012): Nocturnal activity in a Serbian population of *Podarcis muralis*. *Herpetozoa* **25**(1/2): 87-89.

Castilla, A.M., Van Damme, R., Bauwens, D. (1999): Field body temperatures, mechanisms of thermoregulation and evolution of thermal characteristics in lacertid lizards. *Nat. Croat.* **8**: 253-274.

Ceirâns, A. (2007): Distribution and habitats of the Sand Lizard (*Lacerta agilis*) in Latvia. *Acta Univ. Latv.* **723**: 53-59.

Cooke, A.S., (1991): The habitat of sand lizards *Lacerta agilis* at Merseyside, English Nature, Peterborough, Nature Conservancy Council.

Corbett, K. (1988): Distribution and status of the Sand lizard, *Lacerta agilis agilis*, in Britain. *Mertensiella* **1**: 92-100.

Corbett, K. (2001): Status, Threats and Conservation Requirements of *Lacerta bilineata* on Jersey. *Mertensiella* **13**: 98-104.

Covaciu-Marcov, S.D., Bogdan, H.V., Cristiana, P., Toader, S., Condure, N. (2008): The herpetofauna of the north-western region of Bihor County, Romania. *Bihorean Biologist* **2**: 5-13.

Covaciu-Marcov, S.D., Cicort-Lucaciu, A.Ş., Gaceu, O., Sas, I., Ferenţi, S., Bogdan, H.V. (2009): The herpetofauna of the south-western part of Mehedinţi County, Romania. *North-Western J. Zool.* **5**: 142-164.

Dely, O.Gy. (1983): Hüllők-Reptilia. Magyarország Állatvilága (Fauna Hungariae) **20** (4), Budapest, Academic Press.

Dent, S., Spellerberg, I.F. (1987): Habitat of the lizards *Lacerta agilis* and *Lacerta vivipara* on forest ride vergers in Britain. *Biol. Conserv.* **42**: 273-286.

Dent, S., Spellenberg, I.F. (1988): Use of forest ride verges in Southern England for the conservation of the sand lizard *Lacerta agilis* L. *Biol. Conserv.* **45**: 267-277.

Díaz, J.A., Monasterio, C., Salvador, A. (2006): Abundance, microhabitat selection and conservation of eyed lizards (*Lacerta lepida*): a radiotelemetric study. *J. Zool.* **268**: 295-301.

Edgar, P., Bird, D.R. (2006): Action Plan for the Conservation of the Sand Lizard (*Lacerta agilis*) in Northwest Europe. In: Convention on the conservation of European

wildlife and natural habitats-Standing Committee, Strasbourg, 2006. Bournemouth, The Herpetological Conservation Trust.

Èeirâns, A. (2007): Microhabitat characteristics for reptiles *Lacerta agilis*, *Zootoca vivipara*, *Anguis fragilis*, *Natrix natrix*, and *Vipera berus* in Latvia. Russ. J. Herpetol. **14**: 172-176.

Elbing, K. (1997): How to arrange rendezvous-distribution patterns and reproduction events in a small population of sand lizard (*Lacerta agilis*). In: Herpetologia Bonnensis, p. 99-104. Böhme, W., Bischoff, W., Ziegler, T., Eds, Bonn, Societas Europaea Herpetologica.

Faeth, S.H., Warren, P.S., Shochat, E., Marussich, W.A. (2005): Trophic dynamics in urban communities. BioScience **55**: 399-407.

Garden, J.G., Clive, A., Mcalpine, C.A., Possingham, H.P., Jones, D.N. (2007): Habitat structure is more important than vegetation composition for local-level management of native terrestrial reptile and small mammal species living in urban remnants: A case study from Brisbane, Australia. Austral Ecol. **32**: 669–685.

Glandt, D. (1986): Substrate choice of the sand lizard *Lacerta agilis* and the common lizard *Lacerta vivipara*. In: Studies in Herpetology, p. 143-146. Van Gelder, J.J., Strijbosch, H., Bergers, P.J.M., Eds, Prague, Charles University for the Societas Europaea Herpetologica.

Glandt, D. (1995): Seasonal activity of the sand lizard (*Lacerta agilis*) and the common lizard (*Lacerta vivipara*) in an experimental outdoor enclosure. In: Scientia

Herpetologica, p. 229-231. Llorente, G.A., Montori, A., Santos, X., Carretero, M.A., Eds, Barcelona, Asociacion Herpetologica Espanola.

Grbac, I., Bauwens, D. (2001): Constraints on temperature regulations in two sympatric podarcis lizards during autumn. *Copeia* **2001**:178-186.

House, S.M., Spellerberg, I.F. (1983): Ecology and conservation of the sand lizard (*Lacerta agilis* L.) habitat in Southern England. *J. Appl. Ecol.* **20**: 417-437.

House, S.M., Taylor, P.J., Spellerberg, I.F. (1979): Patterns of daily behaviour in two lizard species *Lacerta agilis* L. and *Lacerta vivipara* Jacquin. *Oecologia* **44**: 396-402.

IBM Corp. Released (2011): IBM SPSS Statistics for Windows, Version 20.0. Armonk, New York, IBM Corp.

Iftime, A. (2005): Herpetological observations in the Danube floodplain sector in the Giurgiu County (Romania). *Trav Mus Natl Hist Nat Grigore Antipa Grigore Antipa* **4**: 339-348.

Iftime, A., Iftime, O. (2006): Preliminary data on the herpetofauna of the Cozia massif (Romania). 1. Reptiles. *Trav Mus Natl Hist Nat Grigore Antipa* **49**: 331-340.

Ihász, N., Bayer, K., Kopena, R., Molnár, O., Herczeg, G., Török, J. (2006): Szemben a ragadozóval – a zöld gyík (*Lacerta viridis*) búvóhelyközpontú menekülési stratégiája. *Állattani Közlemények* **90**: 127-138.

Ilosvay, Gy. (1977): Effect of urbanization on the herpetofauna of a settlement at the Tisza (Szeged). *Tiscia* **12**: 123-130.

Jellinek, S., Driscoll, D.A., Kirkpatrick, J.B. (2004): Environmental and vegetation variables have a greater influence than habitat fragmentation in structuring lizard communities in remnant urban bushland. *Austral Ecol.* **29**: 294–304.

Koenig, J., Shine, R., Shea, G. (2001): The ecology of an Australian reptile icon: how do blue-tongued lizards (*Tiliqua scincoides*) survive in suburbia? *Wildlife Res.* **28**: 214–227.

Koenig, J., Shine, R., Shea, G. (2002): The dangers of life in the city: patterns of activity, injury and mortality in suburban lizards (*Tiliqua scincoides*). *J. Herpetol.* **36**: 62-68.

Korsós, Z. (1984): Comparative niche analysis of two sympatric lizard species (*Lacerta viridis*) and (*Lacerta agilis*). *Vert. Hung.* **21**: 5-14.

Korsós, Z. (1986): Ecological Comparison of *Lacerta viridis* and *L. agilis*. In: *Studies in Herpetology*, p. 455-458. Rocec, Z., Ed., Prague, Charles University for the Societas Europaea Herpetologica.

Korsós, Z., Gyovai, F. (1988): Habitat dimension and activity pattern differences in allopatric populations of *Lacerta agilis*. *Mertensiella* **1**: 235-244.

Kuranova, V.N., Patrakov, S.V., Bulakova, N.A., Krechetova, O.A. (2005): The study of the ecological niche segregation for sympatric species of lizards *Lacerta agilis* and *Zootoca vivipara*. In: *Herpetologia Petropolitana. Proc. of the 12th Ord. Gen. Meeting Soc. Eur. Herpetol.*, August 12 – 16, 2003, St. Petersburg. Ananjeva, N., Tsinenko, O., Eds, *Russ. J. Herpetol.* **12**(Suppl.): 171-175.

Legendre, P., Legendre, L. (1998): Numerical ecology. 2nd English Edition, Amsterdam, Elsevier Science BV.

Loos, J., Dayan, T., Drescher, N., Levanony, T., Maza, E., Shacham, B., Talbi, R., Assmann, T. (2011): Habitat preferences of the Levant Green Lizard, *Lacerta media israelica* (Peters, 1964). Zool. Middle East **52**: 17-28.

Martens, H., Jelden, D. (1992): Pet trade in reptiles and amphibians- perspectives of their conservation status within the European Community (EC). In: Proc. Sixth Ord. Gen. Meet. S. E. H., p. 309-313. Korsós, Z., Kiss, I., Eds, Budapest, Hungarian Natural History Museum.

McKinney, M.L. (2008): Effect of urbanization on species richness: a review of plants and animals. Urban Ecosyst. **11**: 161-176.

Mikátová, B. (2001): The Green Lizard, *Lacerta viridis* (Laurenti, 1786), in the Czech Republic: Distribution, Ecology and Conservation aspects. Mertensiella **13**: 138-149.

Mollov, I.A. (2005): A study on the amphibians (*Amphibia*) and reptiles (*Reptilia*) from three urban protected areas in the town of Plovdiv (South Bulgaria). Scientific Studies of the University of Plovdiv-Biology, Animalia **41**: 79-94.

Mollov, I.A. (2011): Habitat distribution of the amphibians and reptiles in the city of Plovdiv, Bulgaria. Biharean Biologist **5**: 25-31.

Nemes, Sz. (2002): Foraging mode of the sand lizard, *Lacerta agilis*, at the beginning of its yearly activity period. Russ. J. Herpetol. **9**: 57-62.

Nemes, Sz., Vogrin, M., Hartel, T., Öllerer, K. (2006): Habitat selection at the sand lizard (*Lacerta agilis*): ontogenetic shifts. *North-Western J. Zool.* **2**: 17-26.

Olsson, M. (1986): Spatial distribution and home range size in the Swedish sand lizard (*Lacerta agilis*) during the mating season. In: *Studies in Herpetology*, p. 597-600. Roček, Z., Ed., Prague, Charles University for the Societas Europaea Herpetologica.

Perry, G., Buchanan, B.W., Fisher, R.N., Salmon, M., Sharon, E., Wise, S.E. (2008): Effects of artificial night lighting on amphibians and reptiles in urban environments. In: *Urban Herpetology, Herpetological Conservation*, p. 239-256. Mitchell, J.C., Jung Brown R.E., Bartholomew B., Eds, Salt Lake City, Society for the Study of Amphibians and Reptiles.

R Core Team (2012): *R: A language and environment for statistical computing*. Vienna, R Foundation for Statistical Computing.

Rismiller, P.D., Heldmaier, G. (1982): The effect of photoperiod on temperature selection in the European green lizard, *Lacerta viridis*. *Oecologia* **53**: 222–226.

Rose, B. (1982): Lizard home ranges: methodology and functions. *J. Herpetol.* **16**: 253-269.

Rugiero, L. (2004): Composition of the reptile communities in five urban protected areas of different isolation degrees. *Herpetozoa* **16**: 151-155.

Rugiero, L., Luiselli, L. (2007): Null model analysis of lizard communities in five urban parks of Rome. *Amphibia-Reptilia* **28**: 547-553.

Santos, T., Díaz, J.A., Pérez-Tris, J., Carbonell, R., Tellería, J.L. (2008): Habitat quality predicts the distribution of a lizard in fragmented woodlands better than habitat fragmentation. *Anim. Conserv.* **11**: 46-56.

Scali, S., Zuffi, M. (1994): Preliminary report on a reptile community ecology in a suburban habitat of northern Italy. *Boll. Zool.* **61**: 73-76.

Schmidt-Loske, K. (1997): Some remarks on habitat use by *Podarcis muralis* Laurenti, 1768 in vine-growing parts of the Ahr valley near Bad Neuenahr-Ahrweiler (Rhineland-Palatinate). In: *Herpetologia Bonnensis*, p. 331-340. Böhme, W., Bischoff, W., Ziegler, T., Eds, Bonn, Societas Europaea Herpetologica.

Schwartz, M. (2010): Beobachtungen an einer Population der Zauneidechse (*Lacerta agilis*) auf einem Friedhof im Münsterland (NRW). *Zeitschrift für Feldherpetologie* **17**: 77-88.

Shochat, E., Lerman, S., Fernández-Juricic, E. (2010): Birds in urban ecosystems: population dynamics, community structure, biodiversity and conservation. In: *Urban Ecosystem Ecology*, p. 75-86. Aitkenhead-Peterson J., Volder A., Eds, Madison, American Society of Agronomy.

Shochat, E., Warren, P.S., Faeth, S.H., McIntyre, N.E., Hope, D. (2006): From patterns to emerging processes in mechanistic urban ecology. *Trends Ecol. Evol.* **21**: 186–191.

Simon, T. (1967): Dunamenti síkság természetes növényzete. In: *A dunai Alföld*, p. 204-207. Pécsi M., Ed., Budapest, Academic Press.

Sos, T. (2007): Notes on distribution and current status of herpetofauna in the northern area of Braşov County (Romania). *North-Western J. Zool.* **3**: 34-52.

Sound, P., Veith M. (2000): Weather effects on intrahabitat movements of the western green lizard, *Lacerta bilineata* (Daudin, 1802), at its northern distribution range border: a radio-tracking study. *Can. J. Zool.* **78**: 1831-1839.

Stamps, J.A. (1977): Social behaviour and spacing patterns in lizards. *Biol. Reptilia* **7**: 265-334.

Strijbosch, H. (1986): Niche Segregation in Sympatric *Lacerta agilis* and *L. vivipara*. In: *Studies in Herpetology*, p. 449-454. Rocec, Z., Ed., Prague, Charles University for the Societas Europaea Herpetologica.

Strijbosch, H., Van Gelder, J.J. (1997): Population structure of lizards in fragmented landscapes and causes of their decline. In: *Herpetologia Bonnensis*, p. 347-357. Böhme, W., Bischoff, W., Ziegler, T., Eds, Bonn, Societas Europaea Herpetologica.

Strugariu, A., Gherghel, I. (2008): A preliminary report on the composition and distribution of the herpetofauna in the Lower Prut River Basin (Romania). *North-Western J. Zool.* **4**: 49-69.

Strugariu, A., Gherghel, I., Huţuleac-Volosciuc, M.V., Puşcaşu, C.M. (2007): Preliminary aspects concerning the herpetofauna from urban and peri-urban environments from North-Eastern Romania: a case study in the city of Suceava. *Herpetologica Romanica* **1**: 53-61.

van Bree, H., Plantaz, R., Zuiderwijk, A. (2006): Dynamics in the sand lizard (*Lacerta agilis*) population at Forteiland, IJmuiden, The Netherlands. In: Herpetologia Bonnensis II, p. 187-190. Vences, M., Köhler, J., Ziegler, T., Böhme, W., Eds, Bonn, Societas Europaea Herpetologica.

Vasváry, M. (1926): Adatok a zöldség- és formakör ismeretéhez. Állattani Közlemények **23**: 34-66.

Vignoli, L., Mocaer, I., Luiselli, L., Bologna, M.A. (2009): Can a large metropolis sustain complex herpetofauna communities? An analysis of the suitability of green space fragments in Rome. Anim. Conserv. **12**:456-466.

Winston, W. (2007): Microsoft Office Excel 2007: data analysis and business modeling. Redmond, WA, USA, Microsoft Press.

Woods, M., McDonald, R.A., Harris, S. (2003): Predation of wildlife by domestic cats *Felis catus* in Great Britain. Mammal Rev. **33**: 174-188.

Zólyomi B. (1967): Mezőföld természetes növényzete. In: A dunai Alföld, p. 285–288. Pécsi M., Ed., Budapest, Academic Press.

Zuur, A., Ieno, E.N., Walker, N., Saveliev, A.A., Smith, G.M. (2009): Mixed effects models and extensions in ecology with R. New York, Springer Science & Business Media.

Table 1. Results of the generalized linear mixed model (GLMM) describing the general activity of lizards. Parameter estimates and their standard errors (SE) of the fixed and interaction effects. *L.a.*: *Lacerta agilis*, *L.v.*: *Lacerta viridis*, D.F: Degrees of freedom, t: t-value, p: p-value

Model	Species	Parameter estimates (logit scale)	SE	D.F.	t	p
time	<i>L.a. & L.v.</i>	6.726	1.131	277	5.947	< 0.001
	<i>L. a.</i>	1.586	0.570	256	2.780	0.006
	<i>L. v.</i>	-0.374	0.669	248	-0.559	0.577
mean temperature	<i>L.a. & L.v.</i>	1.657	0.331	277	5.005	< 0.001
	<i>L. a.</i>	0.439	0.163	256	2.702	0.007
	<i>L. v.</i>	-0.102	0.197	248	-0.519	0.604
time:mean temperature	<i>L.a. & L.v.</i>	-0.195	0.0319	277	-6.111	< 0.001
	<i>L. a.</i>	-0.046	0.0158	256	-2.888	0.004
	<i>L. v.</i>	0.01	0.019	248	0.528	0.598

Figure 1. Principal component analysis of the eight parcels based on the five traits investigated.

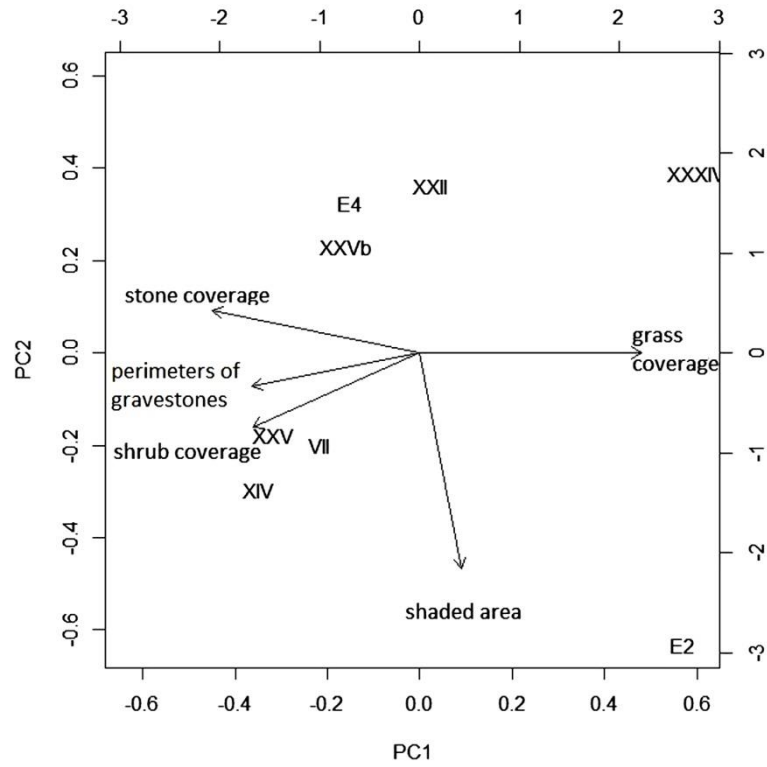


Figure 2. Relationship between mean temperature and the activity of the two species (filled circles: *Lacerta agilis*, open circles: *L. viridis*).

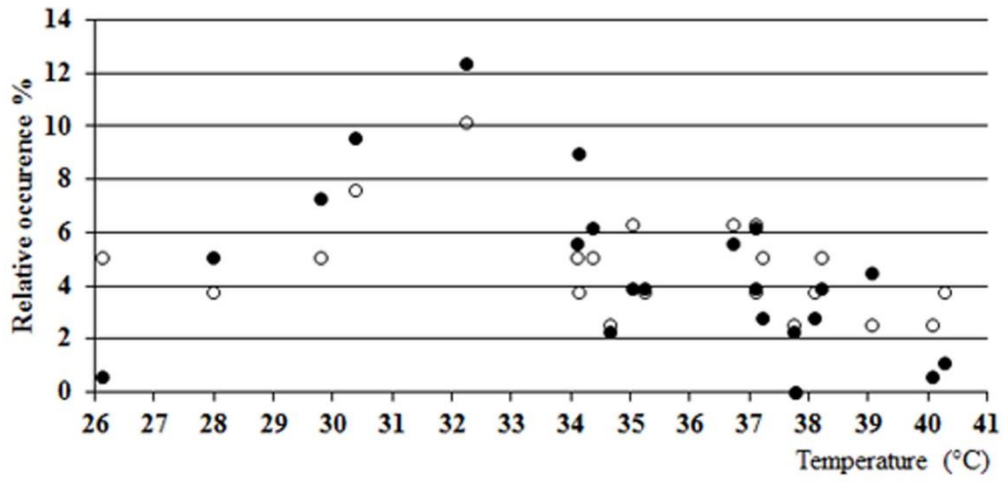


Figure 3. Probability of occurrence of *Lacerta agilis* depending on temperature.

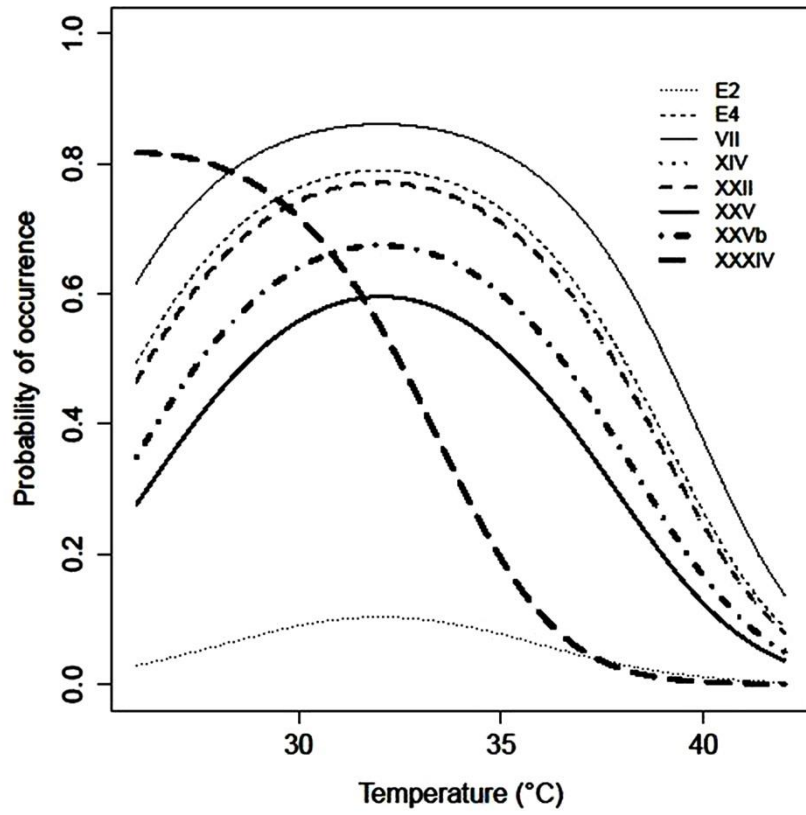


Figure 4. Relative occurrence of *Lacerta agilis* and *L. viridis* during the day (filled circles: *L. agilis*, open circles: *L. viridis*, solid line: trend line for *L. agilis*, dotted line: trend line for *L. viridis*).

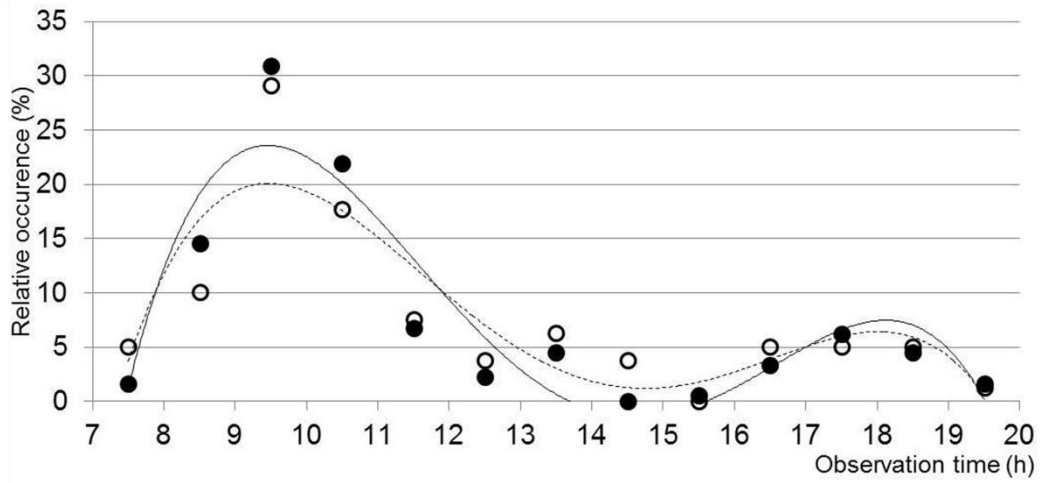


Figure 5. Percentage of areas covered with grass (A), shrub (B) or gravestone (C) on the occurrence points of the two lizard species and on the full area of parcels.

